# Yakima/Klickitat Fisheries Project 

## Monitoring and Evaluation



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# MONITORING AND EVALUATION 

PROJECT NUMBER 95-063-25

THE CONFEDERATED TRIBES AND BANDS OF THE YAKAMA NATION

# YAKIMA/ KLICKITAT FISHERIES PROJECT 

FINAL REPORT 2002

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## MONITORING AND EVALUATION PROJECT REPORT

## Preface

The monitoring and evaluation objectives and tasks have been developed through a joint process between the co-managers, Y akama Nation (Y N, Lead Agency) and Washington Department of Fish and Wildlife (WDFW). The Science/ Technical Advisory Committee (STAC), which consists of core members from the co-managers, employs the services of a work committee of scientists, the Monitoring Implementation Planning Team (MIPT) to develop the Monitoring and Evaluation (M\&E) Plan.

The process employed by STAC to verify these designated activities and the timing of their implementation involved the utilization of the following principles:

1. YKFP monitoring should evaluate the success (or lack of it) of project supplementation efforts and its impacts, including juvenile post release survival, natural production and reproductive success, ecological interactions, and genetics;
2. YKFP monitoring should be comprehensive and,
3. YKFP monitoring should be done in such a way that results are of use to salmon production efforts throughout and Columbia basin and the region.

Utilizing these principles, STAC and MIPT developed this M\&E action plan in three phases. The first phase was primarily conceptual. STAC and MIPT defined critical issues and problems and identified associated response variables. The second phase was quantitative, which determined the scale and size of an effective monitoring effort. A critical element of the quantitative phase was an assessment of the precision with which response variables can be measured, the probability of detecting real impacts and the sample sizes required for a given level of statistical precision and power. The third phase is logistical. The feasibility of monitoring measures was evaluated as to practicality and cost. The Policy Group has determined that the M\&E activities covered by this agreement are necessary, effective and cost-efficient.

## Introduction

The FY 2002 monitoring and evaluation program for the YK FP was organized into four categories- Natural Production (tasks 1.a - 1.y), Harvest (task 2.b), Genetics (tasks 3.a - 3.c) and Ecological Interactions (tasks 4.a - 4.f). This annual report specifically discusses tasks directly conducted by the Yakama Nation. Those tasks that are conducted directly by the Washington State Department of Fish and Wildlife cite the written report where a complete discussion of that task can be found. IntStats provides the biometrical support for the YKFP and IntStats' written reports for tasks 1.e, 1.f., 1.g and 4.c are included in full as appendices to this report.

Special acknowledgement and recognition is owed to all of the dedicated YKFP personnel who are working on various tasks. The referenced accomplishments and achievements are a direct result of their dedication and desire to seek positive results for the betterment of the resource. The readers of this report are requested to pay special attention to the Personnel Acknowledgements. Also, these achievements are attainable because of the efficient and essential administrative support received from all of the office and administrative support personnel for the YKFP.

## NATURAL PRODUCTION

Overall Objective: Develop methods of detecting indices of increasing natural production, as well as methods of detecting a realized increase in natural production, with specified statistical power.

## Task 1.a Modeling

Rationale: To design complementary supplementation/ habitat enhancement programs for targeted stocks with computer models incorporating empirical estimates of life-stage-specific survival and habitat quality \& quantity.

Methods: To diagnose the fundamental environmental factors limiting natural production, and to estimate the relative improvements in production that would result from a combination of habitat enhancement and supplementation using the "Ecosystem Diagnosis and Treatment" (EDT) model. A brief description of the EDT model can be found on the Mobrand Biometrics Incorporated (MBI) website at www.mbi.com.

## Progress:

Yakima
The baseline and diagnostic reports were completed for upper Yakima and Naches spring chinook; lower Yakima and Marion Drain fall chinook; upper Yakima, Naches, Toppenish and Satus summer steelhead; and upper Yakima and Naches coho. The baseline report consists of abundance, productivity and diversity index values for specific geographic areas for each species. The baseline report includes output for both current and historic conditions.

Baseline reports for each species/ subbasin are presented in Tables 1-9. The diagnostic reports for purposes of this report have been expressed in terms of Tornado charts (Figures 1-10), which depict the top restoration and preservation reaches in terms of abundance, productivity and diversity index for each species/ spawning aggregate previously mentioned for the baseline reports. Restoration is defined in the EDT model as the difference between historic and current conditions for abundance, productivity and diversity index. Clearly in many cases it is not feasible to restore reach conditions completely back to historic conditions. Nevertheless, the restoration potential value provides a place to start when assessing which reaches may provide the best restoration opportunities. Obviously the next step is to assess realistically the ability to restore a specific reach given existing physical limitations (e.g.
freeways, roads, dikes, floodplain development) and community values. Preservation is defined as the relative importance a specific reach is to the overall abundance, productivity and diversity index for a particular species/ spawning aggregate. A detailed reach level analysis depicting affected (or non affected) Level 3 attributes and specific salmonid life stages is presented in the EXCEL Report 2 Viewer files for each species/ spawning aggregate. Because of their large size (output of one page per reach) these files have been posted to the YKFP website under Technical Reports and Publications at ykfp@yakama.com for readers interested in obtaining reach specific information on habitat conditions, etc.

Further calibration of the summer steelhead and coho models is warranted based on the initial baseline model runs complete. It was recognized two years ago that the EDT model over predicted the upper Y akima steelhead population abundance based on the available habitat. It was hypothesized that this was a consequence of past operational procedures at Roza D am that resulted in the resident form of 0 mykiss to expand and eventually hold a competitive advantage over the anadromous form. A mathematical analog has been developed to correct for this biological interaction, but needs further refinement so that all four populations are calibrated in relative terms in proportions similar to those currently observed in the Y akima Basin. The coho EDT model may currently predict low numbers of adults. Because coho are currently being reintroduced into historic subbasins using a non-indigenous stock it's difficult to use current adult return rates as good means to calibrate the model since this population has not reach equilibrium in terms of geographic distribution and population size. We will be reviewing the habitat EDT Level 2 attributes for accuracy for the major floodplain reaches (most important for coho) before accepting current EDT model outputs parameters for coho.

With completion of the initial baseline and diagnostic EDT model runs for the key salmonid species attention in FY 03 will focus on application of the model to assess specific habitat restoration projects (or scenarios in the EDT vernacular). Work will also focus on reviewing and upgrading Level 2 attributes for the Wilson-Nanuem watershed, which is the focus of current habitat assessment surveys and removal of passage barriers.

In FY 2002 CWU was contracted to develop a biophysical classification scheme and measurement protocols designed to increase the precision of selected Level 2 attributes EDT model (see Appendix A for a complete discussion).

Table 1. Summary of upper Yakima basin spring chinook performance (by diagnostic area) in terms of equilibrium abundance, productivity and life history diversity for current and historical conditions.

| Baseline Performance Analysis, Upper Yakima Spring Chinook |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CURRENT CONDITIONS |  |  |  |  |  |  |  |  |
|  | Equilibrium Abundance |  |  | Smolt-to-adult Survival (at Prosser Dam) | Productivity |  |  | Life History Diversity |
| Diagnostic Area | Smolts at Prosser | Adults at Prosser | Adults on Spawning Grounds |  | Smolts per Spawner (at Prosser) | Adult Returns per Spawner (at Prosser) | Adult Returns per Spawner (on Spawning Grounds) |  |
| Upper Yakima Tributaries (Cle Elum R and Taneum, Swauk, Manastash, Big and Little Creeks) | 21,576 | 761 | 485 | 3.5\% | 166 | 6.0 | 5 | 13.0\% |
| Upper Yakima Mainstem, Wilson Cr to Keechelus Dam | 53,244 | 1,920 | 1,436 | 3.6\% | 120 | 4.4 | 3.5 | 96.0\% |
| Upper Yakima Mainstem, Ahtanum Cr to Wilson Cr | 32,557 | 1,095 | 680 | 3.4\% | 98 | 3.4 | 2.2 | 37.0\% |
| Teanaway Watershed | 2,985 | 100 | 46 | 3.4\% | 73 | 2.4 | 1.4 | 4.0\% |
| Wilson/Naneum Watershed | 1,964 | 66 | 32 | 3.4\% | 78 | 2.6 | 1.8 | 2.0\% |
| Ahtanum/Wide Hollow Watershed | 3,866 | 138 | 60 | 3.6\% | 76 | 2.8 | 1.4 | 1.0\% |
| Wenas Creek Watershed | 0 | 0 | 0 | N.A. |  | 0.0 | 0 | 0.0\% |
| HISTORICAL/NORMATIVE CONDITIONS |  |  |  |  |  |  |  |  |
|  | Equilibrium Abundance |  |  |  | Productivity |  |  |  |
| Diagnostic Area | Smolts at Prosser | Adults at Prosser | Adults on Spawning Grounds | Smolt-to-adult Survival (at Prosser Dam) | Smolts per Spawner (at Prosser) | Adult Returns per Spawner (at Prosser) | Adult Returns per Spawner (on Spawning Grounds) | Life History Diversity |
| Upper Yakima Tributaries (Cle Elum R and Taneum, Swauk, Manastash, Big and Little Creeks) | 239,072 | 26,096 | 20,064 | 10.9\% | 303 | 34.0 | 30 | 100\% |
| Upper Yakima Mianstem, Wilson Cr to Keechelus Dam | 285,528 | 31,425 | 27,801 | 11.0\% | 344 | 39.0 | 35 | 100\% |
| Upper Yakima Mainstem, Ahtanum $\mathbf{C r}$ to Wilson Cr | 298,657 | 30,564 | 25,565 | 10.2\% | 366 | 39.0 | 35 | 100\% |
| Teanaway Watershed | 138,696 | 15,196 | 12,520 | 11.0\% | 326 | 37.0 | 34 | 99\% |
| Wilson/Naneum Watershed | 64,858 | 7,051 | 4,856 | 10.9\% | 208 | 23.0 | 20 | 100\% |
| Ahtanum/Wide Hollow Watershed | 169,798 | 17,581 | 13,491 | 10.4\% | 248 | 26.0 | 24 | 100\% |
| Wenas Creek Watershed | 42,854 | 4,259 | 3,170 | 9.9\% | 225 | 24.0 | 22 | 100\% |

## Upper Yakima Spring Chinook

Relative Importance Of Geographic Areas For Protection and Restoration Measures

| Geographic Area | Protection benefit | Restoration benefit | Change in Abundance with |  | Change in Productivity with |  | Change in Diversity Index with |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Category/rank | Category/rank | Degradation | Restoration | Degradation | Restoration | Degradation | Restoration |
| Yakima Ahtanum to Naches | 3 | 4 |  |  |  |  |  |  |
| Yakima Roza Dam to Wilson Cr | 5 | 3 |  |  |  |  |  |  |
| Yakima Manastash to Taneum | 1 | 9 |  |  |  |  |  |  |
| Columbia above Estuary | 10 | 1 |  |  |  |  |  |  |
| Yakima Wilson to Manastash | 5 | 6 |  |  |  |  |  |  |
| Yakima Naches to Roza Dam | 8 | 4 |  |  |  |  |  |  |
| Yakima Cle Elum to Easton Dam | 2 | 12 |  |  |  |  |  |  |
| Yakima Easton Dam to Keechelus Dam | 5 | 15 |  |  |  |  |  |  |
| Columbia Estuary | 14 | 7 |  |  |  |  |  |  |
| Yakima above storage dams | 16 | 7 |  |  |  |  |  |  |
| Yakima Teanaway to Cle Elum | 8 | 17 |  |  |  |  |  |  |
| Yakima Taneum to Teanaway | 4 | 22 |  |  |  |  |  |  |
| Teanaway drainge above forks | 18 | 11 |  |  |  |  |  |  |
| Teanaway drainge below forks | 27 | 2 |  |  |  |  |  |  |
| Yakima SSide Dam to Ahtanum Cr | 17 | 16 |  |  |  |  |  |  |
| Cle Elum R below Cle Elum Dam | 11 | 24 |  |  |  |  |  |  |
| Taneum Drainage | 15 | 20 |  |  |  |  |  |  |
| Wilson Drainage | 21 | 14 |  |  |  |  |  |  |
| Yakima Toppenish to Sunnyside Dam | 12 | 24 |  |  |  |  |  |  |
| Yakima Prosser Dam to Satus | 19 | 18 |  |  |  |  |  |  |
| Manastash drainge | 26 | 12 |  |  |  |  |  |  |
| Ahtanum Amer Fruit Br to Upper WIP | 31 | 10 |  |  |  |  |  |  |
| Yakima Satus to Toppenish | 13 | 29 |  |  |  |  |  |  |
| Lower Ahtanum \& Wide Hollow | 24 | 19 |  |  |  |  |  |  |
| Ahtanum drainage above WIP | 23 | 26 |  |  |  |  |  |  |
| Yakima delta to Horn Dam | 20 | 30 |  |  |  |  |  |  |
| Wenas Cr Drainage | 32 | 20 |  |  |  |  |  |  |
| Yakima Horn Dam to Benton | 24 | 28 |  |  |  |  |  |  |
| Yakima Chandler Bypass Reach | 22 | 31 |  |  |  |  |  |  |
| Big Cr Drainage | 33 | 23 |  |  |  |  |  |  |
| Swauk Drainage | 30 | 27 |  |  |  |  |  |  |
| Yakima Benton to Powerplant | 28 | 32 |  |  |  |  |  |  |
| Yakima delta | 29 | 33 |  |  |  |  |  |  |
| Coastal Zone | 33 | 34 |  |  |  |  |  |  |
| Offshore Marine | 33 | 34 |  |  |  |  |  |  |
| $\begin{array}{ccc}-75 \% & 0 \% & +75 \% \\ \text { Percentage change }\end{array}$ |  |  |  |  | $\begin{array}{ccc}-75 \% & 0 \% \\ \text { Percentage change }\end{array}$ |  | $-75 \%$ $0 \%$ $+75 \%$ <br> Percentage change   |  |

Figure 1. Summary of the relative importance of habitat (geographic area) restoration and preservation to spring chinook in the upper Yakima basin.

Table 2. Summary of Naches basin spring chinook performance (by diagnostic area) in terms of equilibrium abundance, productivity and life history diversity for current and historical conditions.

| Baseline Performance Analysis, Naches Drainage Spring Chinook |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CURRENT CONDITIONS |  |  |  |  |  |  |  |  |
|  | Equilibrium Abundance |  |  | Smolt-to-adult <br> Survival (at Prosser Dam) | Productivity |  |  | Life History Diversity |
| Diagnostic Area | Smolts at Prosser | Adults at Prosser | Adults on Spawning Grounds |  | Smolts per <br> Spawner (at Prosser) | Adult Returns per Spawner (at Prosser) | Adult Returns per <br> Spawner (on <br> Spawning Grounds) |  |
| American River and tributaries | 12,091 | 335 | 303 | 2.8\% | 168 | 4.9 | 4.5 | 67.0\% |
| Bumping River, below Bumping Dam | 4,719 | 161 | 112 | 3.4\% | 115 | 4.0 | 2.9 | 81.0\% |
| Bumping River, above Bumping Dam | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0.0\% |
| Little Naches River and tributaries | 11,579 | 357 | 264 | 3.1\% | 160 | 5.1 | 4.4 | 74.0\% |
| Rattlesnake Creek and tributaries | 5,248 | 166 | 127 | 3.2\% | 118 | 3.9 | 3.1 | 83.0\% |
| Tieton River and tributaries below Rimrock Dam | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0.0\% |
| Tieton River and tributaries above Rimrock Dam | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0.0\% |
| Cowiche Creek and tributaries | 211 | 8 | 5 | 3.8\% | 46 | 1.8 | 1.2 | 1.0\% |
| Naches River mainstem below Tieton confluence | 11,527 | 312 | 144 | 2.7\% | 142 | 4.5 | 2.1 | 24.0\% |
| Naches River mainstem above Tieton confluence | 15,617 | 494 | 328 | 3.2\% | 106 | 3.5 | 2.4 | 86.0\% |
| HISTORICAL/NORMATIVE CONDITIONS |  |  |  |  |  |  |  |  |
|  | Equilibrium Abundance |  |  |  |  |  |  |  |
| Diagnostic Area | Smolts at Prosser | Adults at Prosser | Adults on Spawning Grounds | Smolt-to-adult <br> Survival (at Prosser Dam) | Smolts per Spawner (at Prosser) | Adult Returns per Spawner (at Prosser) | Adult Returns per <br> Spawner (on <br> Spawning Grounds) | Life History Diversity |
| American River and tributaries | 81,106 | 7,031 | 5,898 | 8.7\% | 448 | 41.2 | 38.9 | 95\% |
| Bumping River, below Bumping Dam | 51,976 | 5,260 | 4.491 | 10.1\% | 427 | 43.7 | 39.7 | 100\% |
| Bumping River, above Bumping Dam | 12,357 | 1,005 | 786 | 8.1\% | 399 | 37.4 | 34.4 | 100\% |
| Little Naches River and tributaries | 85,826 | 8,629 | 6,863 | 10.1\% | 469 | 47.7 | 43.5 | 100\% |
| Rattlesnake Creek and tributaries | 42,992 | 4,276 | 3,388 | 9.9\% | 396 | 39.8 | 35.8 | 100\% |
| Tieton River and tributaries below Rimrock Dam | 76,970 | 7,620 | 6,183 | 9.9\% | 404 | 40.6 | 35.2 | 100\% |
| Tieton River and tributaries above Rimrock Dam | 109,504 | 10,927 | 8,235 | 10.0\% | 342 | 34.8 | 32.2 | 100\% |
| Cowiche Creek and tributaries | 51,111 | 4,024 | 2,953 | 7.9\% | 469 | 44.5 | 39.6 | 100\% |
| Naches River mainstem below Tieton confluence | 146,932 | 12,129 | 10,507 | 8.3\% | 550 | 53.4 | 48.4 | 100\% |
| Naches River mainstem above Tieton confluence | 141,753 | 13,891 | 10,837 | 9.8\% | 479 | 48.4 | 44.7 | 100\% |

## Naches River Drainage Spring Chinook (including the American River stock)

Relative Importance Of Geographic Areas For Protection and Restoration Measures

| Geographic Area | Protection |  | Restoration |  | Change in Abundance with |  | Change in Productivity with |  |  | Change in Diversity Index with |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Category/rank |  | Category/rank |  | Degradation | Restoration | Degradation | Restoration |  | Degradation | Restoration |  |  |
| Naches Cowiche to Tieton | A | 3 | A | 2 |  |  |  |  |  |  |  |  |  |
| Columbia above Estuary | B | 7 | A | 1 |  |  |  |  |  |  |  |  |  |
| Naches Nile to L Naches/Bumping | A | 1 | B | 10 |  |  |  |  |  |  |  |  |  |
| Bumping below Bumping Dam | B | 7 | B | 7 |  |  |  |  |  |  |  |  |  |
| Yakima Toppenish to Sunnyside Dam | A | 4 | B | 12 |  |  |  |  |  |  |  |  |  |
| L. Naches mouth to Salmon Falls | B | 9 | B | 8 |  |  |  |  |  |  |  |  |  |
| Naches Tieton to Rattlesnake | B | 9 | B | 10 |  |  |  |  |  |  |  |  |  |
| Yakima SSide Dam to Ahtanum Cr | C | 14 | B | 5 |  |  |  |  |  |  |  |  |  |
| L. Naches above Salmon Falls | B | 6 | C | 15 |  |  |  |  |  |  |  |  |  |
| Columbia Estuary | D | 16 | B | 6 |  |  |  |  |  |  |  |  |  |
| Yakima Ahtanum to Naches | B | 5 | C | 17 |  |  |  |  |  |  |  |  |  |
| American R. and tribs | A | 1 | D | 22 |  |  |  |  |  |  |  |  |  |
| Tieton below Rimrock Dam | D | 22 | B | 3 |  |  |  |  |  |  |  |  |  |
| Naches mouth to Cowiche | C | 13 | C | 13 |  |  |  |  |  |  |  |  |  |
| Rattlesnake Drainage | B | 9 | C | 18 |  |  |  |  |  |  |  |  |  |
| Naches Rattlesnake to Nile | C | 12 | C | 16 |  |  |  |  |  |  |  |  |  |
| Tieton drainage above Rimrock Dam | E | 25 | B | 4 |  |  |  |  |  |  |  |  |  |
| Yakima Prosser Dam to Satus | D | 17 | C | 14 |  |  |  |  |  |  |  |  |  |
| Cowiche Drainage | E | 24 | B | 8 |  |  |  |  |  |  |  |  |  |
| Yakima Satus to Toppenish | C | 15 | D | 21 |  |  |  |  |  |  |  |  |  |
| Yakima Horn Dam to Benton | D | 20 | D | 19 |  |  |  |  |  |  |  |  |  |
| Yakima delta to Horn Dam | D | 18 | E | 23 |  |  |  |  |  |  |  |  |  |
| Yakima Chandler Bypass Reach | D | 18 | E | 24 |  |  |  |  |  |  |  |  |  |
| Bumping above Bumping Dam | E | 25 | D | 19 |  |  |  |  |  |  |  |  |  |
| Yakima Benton to Powerplant | D | 20 | E | 25 |  |  |  |  |  |  |  |  |  |
| Yakima delta | E | 23 | E | 26 |  |  |  |  |  |  |  |  |  |
| Coastal Zone | E | 25 | E | 27 |  |  |  |  |  |  |  |  |  |
| Offshore Marine | E | 25 | E | 27 |  |  |  |  |  |  |  |  |  |
| $-75 \% \quad 0 \% \quad+75 \%$Percentage change |  |  |  |  |  |  | $\begin{aligned} & -75 \% \quad 0 \% \quad+75 \% \\ & \text { Percentage change } \end{aligned}$ |  |  | $\begin{array}{ccc} -75 \% & 0 \% & +75 \% \\ \text { Percentage change } \\ \hline \end{array}$ |  |  |  |

Figure 2. Summary of the relative importance of habitat (geographic area) restoration and preservation to spring chinook in the Naches basin.

Table 3. Summary of upper Yakima basin coho performance (by diagnostic area) in terms of equilibrium abundance, productivity and life history diversity for current and historical conditions.

| Baseline Performance Analysis, Upper Yakima Coho |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CURRENT CONDITIONS |  |  |  |  |  |  |  |  |
|  | Equilibrium Abundance |  |  | Smolt-to-adult <br> Survival (at <br> Prosser Dam) | Productivity |  |  | Life History Diversity |
| Diagnostic Area | Smolts at Prosser | Adults at Prosser | Adults on Spawning Grounds |  | Smolts per Spawner (at Prosser) | Adult Returns per Spawner (at Prosser) | Adult Returns per Spawner (on Spawning Grounds) |  |
| Upper Yakima Tributaries (Cle Elum R and Taneum, Swauk, Manastash, Big and Little Creeks) | 4,888 | 90 | 85 | 1.8\% | 81 | 1.5 | 1.48 | 4.6\% |
| Upper Yakima Mainstem, Wilson Cr to Keechelus Dam | 15,951 | 287 | 281 | 1.8\% | 88 | 1.6 | 1.58 | 0.2\% |
| Upper Yakima Mainstem, Ahtanum Cr to Wilson Cr | 5.326 | 78 | 75 | 1.5\% | 98 | 1.5 | 1.42 | 9.1\% |
| Teanaway Watershed | 2,691 | 45 | 43 | 1.7\% | 78 | 1.3 | 1.24 | 0.3\% |
| Wilson/Naneum Watershed | 132 | 2 | 2 | 1.5\% | 72 | 1.2 | 1.05 | 0.3\% |
| Ahtanum/Wide Hollow Watershed | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0.0\% |
| Wenas Creek Watershed | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0.0\% |
| HISTORICAL/NORMATIVE CONDITIONS |  |  |  |  |  |  |  |  |
|  | Equilibrium Abundance |  |  |  | Productivity |  |  |  |
| Diagnostic Area | Smolts at Prosser | Adults at Prosser | Adults on Spawning Grounds | Smolt-to-adult <br> Survival (at Prosser Dam) | Smolts per Spawner (at Prosser) | Adult Returns per Spawner (at Prosser) | Adult Returns per Spawner (on Spawning Grounds) | Life History Diversity |
| Upper Yakima Tributaries (Cle Elum R and Taneum, Swauk, Manastash, Big and Little Creeks) | 210,512 | 14,395 | 13,434 | 6.8\% | 206 | 14.1 | 13.9 | 68\% |
| Upper Yakima Mainstem, Wilson Cr to Keechelus Dam | 268,566 | 17.608 | 17,135 | 6.6\% | 234 | 15.8 | 15.6 | 98\% |
| Upper Yakima Mainstem, Ahtanum Cr to Wilson Cr | 207,970 | 13,054 | 12,648 | 6.3\% | 324 | 20.4 | 20.3 | 100\% |
| Teanaway Watershed | 449,730 | 30,818 | 28,499 | 6.9\% | 280 | 19.4 | 19.3 | 77\% |
| Wilson/Naneum Watershed | 77,059 | 5,014 | 4.666 | 6.5\% | 201 | 13.3 | 13.2 | 100\% |
| Ahtanum/Wide Hollow Watershed | 164,196 | 10,455 | 9,850 | 6.4\% | 230 | 14.8 | 14.7 | 96\% |
| Wenas Creek Watershed | 45,252 | 2,868 | 2,713 | 6.3\% | 216 | 13.9 | 13.7 | 77\% |

## Upper Yakima Coho

Relative Importance Of Geographic Areas For Protection and Restoration Measures


Figure 3. Summary of the relative importance of habitat (geographic area) restoration and preservation to coho in the upper Yakima basin.

Table 4. Summary of Naches basin coho performance (by diagnostic area) in terms of equilibrium abundance, productivity and life history diversity for current and historical conditions.

| Baseline Performance Analysis, Naches Drainage Coho |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CURRENT CONDITIONS |  |  |  |  |  |  |  |  |
|  | Equilibrium Abundance |  |  | Smolt-to-adult Survival (at Prosser Dam) | Productivity |  |  | Life History Diversity |
| Diagnostic Area | Smolts at Prosser | Adults at Prosser | Adults on Spawning Grounds |  | Smolts per Spawner (at Prosser) | Adult Returns per Spawner (at Prosser) | Adult Returns per Spawner (on Spawning Grounds) |  |
| American River and tributaries | 340 | 4 | 4 | 1.2\% | 87 | 1.1 | 1.1 | 1.0\% |
| Bumping River, below Bumping Dam | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0.0\% |
| Bumping River, above Bumping Dam | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0.0\% |
| Little Naches River and tributaries | 2,015 | 34 | 33 | 1.7\% | 79 | 1.4 | 1.3 | 1.0\% |
| Rattlesnake Creek and tributaries | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0.0\% |
| Nile Cree | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0.0\% |
| Tieton River and tributaries below Rimrock Dam | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0.0\% |
| Tieton River and tributaries above Rimrock Dam | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0.0\% |
| Cowiche Creek and tributaries | 81 | 1 | 1 | 1.2\% | 68 | 1.2 | 1.1 | 1.0\% |
| Naches River mainstem below Tieton confluence | 2,439 | 37 | 33 | 1.5\% | 89 | 1.4 | 1.2 | 7.0\% |
| Naches River mainstem above Tieton confluence | 1,564 | 26 | 26 | 1.7\% | 73 | 1.3 | 1.2 | 1.0\% |
| HISTORICAL/NORMATIVE CONDITIONS |  |  |  |  |  |  |  |  |
|  | Equilibrium Abundance |  |  | Smolt-to-adult Survival (at Prosser Dam) | Productivity |  |  | Life History Diversity |
| Diagnostic Area | Smolts at Prosser | Adults at Prosser | Adults on Spawning Grounds |  | Smolts per <br> Spawner (at <br> Prosser) | Adult Returns per <br> Spawner (at <br> Prosser) | Adult Returns per Spawner (on Spawning Grounds) |  |
| American River and tributaries | 42,507 | 3,103 | 2,867 | 7.3\% | 197 | 14.3 | 14.2 | 23\% |
| Bumping River, below Bumping Dam | 54.891 | 3.949 | 3.826 | 7.2\% | 174 | 12.7 | 12.6 | 94\% |
| Bumping River, above Bumping Dam | 8,189 | 605 | 562 | 7.4\% | 150 | 11.1 | 11 | 43\% |
| Little Naches River and tributaries | 97,712 | 7,204 | 6,664 | 7.4\% | 186 | 13.8 | 13.7 | 83\% |
| Rattlesnake Creek and tributaries | 27,135 | 1,903 | 1,797 | 7.0\% | 225 | 15.9 | 15.9 | 78\% |
| Nile Cree | 8,260 | 584 | 509 | 7.1\% | 216 | 15.5 | 15.4 | 100\% |
| Tieton River and tributaries below Rimrock Dam | 62,171 | 4,405 | 4,211 | 7.1\% | 221 | 15.7 | 15.6 | 100\% |
| Tieton River and tributaries above Rimrock Dam | 64,741 | 4,715 | 4,466 | 7.3\% | 168 | 12.4 | 12.3 | 67\% |
| Cowiche Creek and tributaries | 26,488 | 1,861 | 1,707 | 7.0\% | 271 | 19.0 | 18.9 | 100\% |
| Naches River mainstem below Tieton confluence | 106,601 | 7,434 | 7,192 | 7.0\% | 343 | 24.0 | 23.9 | 100\% |
| Naches River mainstem above Tieton confluence | 204,846 | 15,073 | 14,056 | 7.4\% | 237 | 17.5 | 17.4 | 100\% |

## Naches Drainage Coho

Relative Importance Of Geographic Areas For Protection and Restoration Measures

| Geographic Area | Protection | Restoration | Change in Abundance with |  | Change in Productivity with |  | Change in Diversity Index |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Category/rank | Category/rank | Degradation | Restoration | Degradation | Restoration | Degradation | Restoration |
| Naches Cowiche to Tieton | 1 | 1 |  |  |  |  |  |  |
| Columbia above Estuary | 2 | 3 |  |  |  |  |  |  |
| Naches Nile to L Naches/Bumping | 5 | 2 |  |  |  |  |  |  |
| Columbia Estuary | 3 | 9 |  |  |  |  |  |  |
| L. Naches above Salmon Falls | 5 | 9 |  |  |  |  |  |  |
| Yakima Ahtanum to Naches | 3 | 16 |  |  |  |  |  |  |
| Bumping below Bumping Dam | 13 | 7 |  |  |  |  |  |  |
| L. Naches mouth to Salmon Falls | 11 | 11 |  |  |  |  |  |  |
| Naches mouth to Cowiche | 10 | 13 |  |  |  |  |  |  |
| Tieton below Rimrock Dam | 19 | 4 |  |  |  |  |  |  |
| Naches Rattlesnake to Nile | 9 | 15 |  |  |  |  |  |  |
| Naches Tieton to Rattlesnake | 8 | 17 |  |  |  |  |  |  |
| Yakima Toppenish to Sunnyside Dam | 7 | 18 |  |  |  |  |  |  |
| Cowiche Drainage | 22 | 5 |  |  |  |  |  |  |
| Yakima Prosser Dam to Satus | 13 | 14 |  |  |  |  |  |  |
| Rattlesnake Drainage | 20 | 8 |  |  |  |  |  |  |
| Tieton drainage above Rimrock Dam | 26 | 6 |  |  |  |  |  |  |
| Yakima SSide Dam to Ahtanum Cr | 20 | 12 |  |  |  |  |  |  |
| Yakima Satus to Toppenish | 15 | 20 |  |  |  |  |  |  |
| Yakima Horn Dam to Benton | 17 | 19 |  |  |  |  |  |  |
| Yakima delta to Horn Dam | 15 | 22 |  |  |  |  |  |  |
| American R. and tribs | 12 | 26 |  |  |  |  |  |  |
| Yakima Chandler Bypass Reach | 18 | 23 |  |  |  |  |  |  |
| Yakima Naches to Roza Dam | 25 | 21 |  |  |  |  |  |  |
| Yakima Benton to Powerplant | 23 | 25 |  |  |  |  |  |  |
| Bumping above Bumping Dam | 26 | 23 |  |  |  |  |  |  |
| Yakima delta | 24 | 27 |  |  |  |  |  |  |
| Coastal Zone | 26 | 28 |  |  |  |  |  |  |
| $-350 \% \quad 0 \% \quad+350 \%$ Percentage change |  |  |  |  | -350\% 0\% +350\% Percentage change |  | $-350 \% \quad 0 \% \quad+350 \%$ <br> Percentage change |  |

Figure 4. Summary of the relative importance of habitat (geographic area) restoration and preservation to coho in the Naches basin.

Table 5. Summary of upper Yakima basin steelhead performance (by diagnostic area) in terms of equilibrium abundance, productivity and life history diversity for current and historical conditions.

| Baseline Performance Analysis, Upper Yakima Summer Steelhead <br> ("Upper Yakima" defined as all mainstem reaches and tributaries upstream of the Yakima-Ahtanum Cr confluence, excluding the Naches River.) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CURRENT CONDITIONS |  |  |  |  |  |  |  |  |  |
|  | Equilibrium Abundance |  |  |  | Smolt-to-adult Survival (at Prosser Dam) | Productivity |  |  | Life History Diversity |
| Diagnostic Area | Smolts at Prosser | Adults at Prosser | Adults on Spawning Grounds | Adjusted Adults on Spawning Grounds ${ }^{\text {a }}$ |  | Smolts per Spawner (at Prosser) | Adult Returns per Spawner (at Prosser) | Adult Returns per <br> Spawner (on <br> Spawning <br> Grounds) |  |
| Upper Yakima Tributaries | 21,268 | 397 | 333 | 109 | 1.9\% | 85 | 1.8 | 1.62 | 3\% |
| Yakima Mainstem above Roza Dam | 6,016 | 108 | 95 | 31 | 1.8\% | 77 | 1.6 | 1.37 | 22\% |
| Yakima Mainstem between Roza Dam and Ahtanum Cr. | 0 | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 5\% |
| Wilson/Naneum Cr. Watershed | 0 | 0 | , | 0 | N.A. | 0 | 0.0 | 0 | 0\% |
| Wenas Cr. Watershed | 0 | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0\% |
| Ahranum/Wide Hollow Watershed | 0 | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0\% |
|  |  |  | HISTO | ORICAL/NORMATIVE | CONDITIONS |  |  |  |  |
|  |  | Equilibriu | Abundance |  |  |  | Productivity |  |  |
| Diagnostic Area | Smolts at Prosser | Adults at Prosser | Adults on Spawning Grounds | Adjusted Adults on Spawning Grounds ${ }^{\text {a }}$ | Smolt-to-adult Survival (at Prosser Dam) | Smolts per Spawner (at Prosser) | Adult Returns per Spawner (at Prosser) | Adult Returns per <br> Spawner (on <br> Spawning <br> Grounds) | Life History Diversity |
| Upper Yakima Tributaries | 280,375 | 24,492 | 18,713 | 18,713 | 8.7\% | 196 | 20.2 | 19.3 | 81\% |
| Yakima Mainstem above Roza Dam | 66.561 | 5.266 | 5.125 | 5.125 | 7.9\% | 185 | 17.7 | 16.6 | 100\% |
| Yakima Mainstem between Roza Dam and Ahtanum Cr. | 29,114 | 2.050 | 2.152 | 2.152 | 7.0\% | 234 | 20.2 | 19.1 | 100\% |
| Wilson/Naneum Cr. Watershed | 44,064 | 3.667 | 4.218 | 4.218 | 8.3\% | 142 | 14.8 | 14.3 | 100\% |
| Wenas Cr. Watershed | 70,103 | 5,330 | 5,457 | 5,457 | 7.6\% | 172 | 17.2 | 16.6 | 96\% |
| Ahranum/Wide Hollow Watershed | 10,975 | 994 | 1,087 | 1,087 | 9.1\% | 125 | 13.8 | 13.3 | 78\% |
| a. The adjusted spawning escapement is the result of a provisional solution to the resident/anadromous equilibrium. |  |  |  |  |  |  |  |  |  |



Figure 5. Summary of the relative importance of habitat (geographic area) restoration and preservation to steelhead in the upper Yakima basin.

Table 6. Summary of Naches basin steelhead performance (by diagnostic area) in terms of equilibrium abundance, productivity and life history diversity for current and historical conditions.

| CURRENT CONDITIONS |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Equilibrium Abundance |  |  |  | Smolt-to-adult <br> Survival (at Prosser Dam) | Productivity |  |  | Life History Diversity |
| Diagnostic Area | Smolts at Prosser | Adults at Prosser | Adults on Spawning Grounds | Adjusted Adults on Spawning Grounds ${ }^{\text {a }}$ |  | Smolts per Spawner (at Prosser) | Adult Returns per Spawner (at Prosser) | Adult Returns per Spawner (on Spawning Grounds) |  |
| AmericanSthd | 6.810 | 99 | 90 | 90 | 1.5\% | 156 | 2.6 | 2.4 | 27\% |
| Bumping above dam | 0 | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0\% |
| Bumping below dam | 824 | 21 | 19 | 19 | 2.5\% | 47 | 1.3 | 1.2 | 3\% |
| Cowiche drainage | 0 | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0\% |
| L Naches drainage | 14,002 | 226 | 204 | 204 | 1.6\% | 105 | 1.9 | 1.7 | 19\% |
| Lower Tieton drainage | 5,721 | 87 | 77 | 77 | 1.5\% | 102 | 1.7 | 1.6 | 2\% |
| Naches above Tieton | 5,816 | 93 | 85 | 85 | 1.6\% | 86 | 1.5 | 1.3 | 16\% |
| Naches below Tieton | 7.229 | 113 | 104 | 104 | 1.6\% | 92 | 1.6 | 1.5 | 4\% |
| Nile drainage | 0 | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0\% |
| Rattlesnake drainage | 7,325 | 115 | 105 | 105 | 1.6\% | 108 | 1.8 | 1.7 | 25\% |
| Upper Tieton drainage | 0 | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0\% |
|  |  |  |  | ISTORICAL/NO | ORMATIVE COND | DITIONS |  |  |  |
|  |  | Equilibriu | m Abundance |  |  |  | Productivity |  |  |
| Diagnostic Area | Smolts at Prosser | Adults at Prosser | Adults on Spawning Grounds | Adjusted Adults on Spawning Grounds ${ }^{\text {a }}$ | Smolt-to-adult <br> Survival (at <br> Prosser Dam) | Smolts per Spawner (at Prosser) | Adult Returns per Spawner (at Prosser) | Adult Returns per Spawner (on Spawning Grounds) | Life History Diversity |
| AmericanSthd | 36,905 | 3.068 | 2.908 | 2,908 | 8.3\% | 251 | 25.0 | 24 | 63\% |
| Bumping above dam | 5.501 | 475 | 454 | 454 | 8.6\% | 212 | 19.0 | 18 | 35\% |
| Bumping below dam | 34,846 | 2,706 | 2,604 | 2,604 | 7.8\% | 185 | 19.0 | 18 | 97\% |
| Cowiche drainage | 17,237 | 1,501 | 1,438 | 1,438 | 8.7\% | 198 | 21.0 | 20 | 96\% |
| L Naches drainage | 77,389 | 6,756 | 6,438 | 6,438 | 8.7\% | 199 | 20.0 | 20 | 89\% |
| Lower Tieton drainage | 72,313 | 4.969 | 4.752 | 4,752 | 6.9\% | 214 | 21.0 | 20 | 98\% |
| Naches above Tieton | 89,627 | 7.703 | 7.318 | 7.318 | 8.6\% | 232 | 25.0 | 24 | 1\% |
| Naches below Tieton | 178,410 | 14,014 | 13,196 | 13,196 | 7.9\% | 268 | 27.0 | 26 | 95\% |
| Nile drainage | 5,795 | 701 | 670 | 670 | 12.1\% | 195 | 23.0 | 22 | 98\% |
| Rattlesnake drainage | 22,064 | 2,277 | 2,183 | 2,183 | 10.3\% | 186 | 21.0 | 20 | 90\% |
| Upper Tieton drainage | 68,446 | 5,872 | 5,519 | 5,519 | 8.6\% | 170 | 18.0 | 17 | 80\% |

## Naches Drainage (including American River) Summer Steelhead

Relative Importance Of Geographic Areas For Protection and Restoration Measures

| Geographic Area | Protection <br> Category/rank |  | Restoration Category/rank |  | Change in Abundance with |  |  |  | Change in Productivity with |  |  |  | Change in Diversity Index with |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Degradation | Restoration |  |  | Degradation | Restoration |  |  | Degradation |  | Restoration |  |
| Columbia above Estuary | B | 3 |  |  | A | 1 |  |  |  |  |  |  |  |  |  |  |  |  |
| Naches Cowiche to Tieton | B | 4 | A | 2 |  |  |  |  |  |  |  |  |  |  |  |  |
| Naches Nile to L Naches/Bumping | B | 4 | B | 3 |  |  |  |  |  |  |  |  |  |  |  |  |
| Columbia Estuary | B | 6 | B | 5 |  |  |  |  |  |  |  |  |  |  |  |  |
| Yakima Prosser Dam to Satus | A | 2 | C | 9 |  |  |  |  |  |  |  |  |  |  |  |  |
| Yakima Satus to Toppenish | A | 1 | D | 18 |  |  |  |  |  |  |  |  |  |  |  |  |
| Tieton below Rimrock Dam | D | 16 | B | 4 |  |  |  |  |  |  |  |  |  |  |  |  |
| Bumping below Bumping Dam | D | 15 | B | 6 |  |  |  |  |  |  |  |  |  |  |  |  |
| L. Naches above Salmon Falls | B | 7 | C | 14 |  |  |  |  |  |  |  |  |  |  |  |  |
| Naches Tieton to Rattlesnake | C | 10 | C | 11 |  |  |  |  |  |  |  |  |  |  |  |  |
| L. Naches mouth to Salmon Falls | C | 12 | C | 10 |  |  |  |  |  |  |  |  |  |  |  |  |
| Naches mouth to Cowiche | D | 17 | B | 6 |  |  |  |  |  |  |  |  |  |  |  |  |
| Rattlesnake Drainage | B | 7 | D | 17 |  |  |  |  |  |  |  |  |  |  |  |  |
| Naches Rattlesnake to Nile | C | 13 | C | 12 |  |  |  |  |  |  |  |  |  |  |  |  |
| Cowiche Drainage | E | 21 | B | 6 |  |  |  |  |  |  |  |  |  |  |  |  |
| Yakima Toppenish to Sunnyside Dam | C | 9 | D | 18 |  |  |  |  |  |  |  |  |  |  |  |  |
| American R. and tribs | C | 10 | E | 24 |  |  |  |  |  |  |  |  |  |  |  |  |
| Yakima Ahtanum to Naches | C | 13 | E | 21 |  |  |  |  |  |  |  |  |  |  |  |  |
| Yakima Horn Dam to Benton | D | 19 | D | 16 |  |  |  |  |  |  |  |  |  |  |  |  |
| Yakima SSide Dam to Ahtanum Cr | E | 22 | C | 13 |  |  |  |  |  |  |  |  |  |  |  |  |
| Yakima delta to Horn Dam | D | 18 | E | 20 |  |  |  |  |  |  |  |  |  |  |  |  |
| Tieton drainage above Rimrock Dam | E | 25 | D | 15 |  |  |  |  |  |  |  |  |  |  |  |  |
| Yakima Chandler Bypass Reach | D | 20 | E | 21 |  |  |  |  |  |  |  |  |  |  |  |  |
| Bumping above Bumping Dam | E | 25 | E | 23 |  |  |  |  |  |  |  |  |  |  |  |  |
| Yakima Benton to Powerplant | E | 23 | E | 25 |  |  |  |  |  |  |  |  |  |  |  |  |
| Yakima delta | E | 24 | E | 26 |  |  |  |  |  |  |  |  |  |  |  |  |
| Coastal Zone | E | 25 | E | 27 |  |  |  |  |  |  |  |  |  |  |  |  |
| Offshore Marine | E | 25 | E | 27 |  |  |  |  |  |  |  |  |  |  |  |  |
| $\begin{array}{ccc} -130 \% \quad 0 \% & +130 \% \\ \text { Percentage change } \end{array}$ |  |  |  |  |  |  |  |  | $-130 \% \quad 0 \% \quad+130 \%$Percentage change |  |  |  | $\begin{array}{ccc} \hline-130 \% \quad 0 \% \quad+130 \% \\ \text { Percentage change } \end{array}$ |  |  |  |

Figure 6. Summary of the relative importance of habitat (geographic area) restoration and preservation to steelhead in the Naches basin.

Table 7. Summary of Toppenish basin spring chinook performance (by diagnostic area) in terms of equilibrium abundance, productivity and life history diversity for current and historical conditions.

## Baseline Performance Analysis, Toppenish Creek Summer Steelhead

"Upper Toppenish Creek" consists of all of those reaches above the Toppenish Lateral Canal and the Simcoe Feeder Canal.
These irrigation diversions dewater Toppenish and Simcoe Creeks, respectively, from June through December.
Therefore, the progeny of fish that spawned above these diversions are effectively isolated and incapable of "normal" juvenile movement patterns.
Marion Drain, Wanity Slough and Harrah Drain are considered a part of Toppenish Creek because of hydraulic connections between Toppenish
reek and Marion Drain, and because of the high probability steelhead spawning in Marion Drain are Toppenish Creek strays.

| CURRENT CONDITIONS |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Equilibrium Abundance |  |  | Smolt-to-adult <br> Survival (at <br> Prosser Dam) | Productivity |  |  | Life History Diversity |
| Diagnostic Area | Smolts at Prosser | Adults at Prosser | Adults on Spawning Grounds |  | Smolts per Spawner (at Prosser) | Adult Returns per Spawner (at Prosser) | Adult Returns per Spawner (on Spawning Grounds) |  |
| Upper Toppenish Creek | 16.480 | 247 | 217 | 1.5\% | 131 | 2.4 | 2.1 | 45\% |
| Lower Toppenish Creek | 2,373 | 39 | 33 | 1.6\% | 87 | 1.5 | 1.3 | 1\% |
| Marion Drain Complex | 0 | 0 | 0 | N.A. | 0 | 0.0 | 0 | 0\% |
| HISTORICAL/NORMATIVE CONDITIONS |  |  |  |  |  |  |  |  |
|  | Equilibrium Abundance |  |  | Smolt-to-adult Survival (at Prosser Dam) | Productivity |  |  | Life History Diversity |
| Diagnostic Area | Smolts at Prosser | Adults at Prosser | Adults on Spawning Grounds |  | Smolts per Spawner (at Prosser) | Adult Returns per Spawner (at Prosser) | Adult Returns per Spawner (on Spawning Grounds) |  |
| Upper Toppenish Creek | 65.208 | 5,242 | 5.016 | 8.0\% | 183 | 16.9 | 16.3 | 97\% |
| Lower Toppenish Creek | 66,062 | 5.316 | 5,108 | 8.0\% | 189 | 16.9 | 16.3 | 95\% |
| Marion Drain Complex | N.A. | N.A. | N.A. | N.A. | N.A. | N.A. | N.A. | N.A. |



Figure 7. Summary of the relative importance of habitat (geographic area) restoration and preservation to steelhead in the Toppenish basin.

Table 8. Summary of Satus basin spring chinook performance (by diagnostic area) in terms of equilibrium abundance, productivity and life history diversity for current and historical conditions.

| Baseline Performance Analysis, Satus Creek Summer Steelhead |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CURRENT CONDITIONS |  |  |  |  |  |  |  |  |
|  | Equilibrium Abundance |  |  | Smolt-to-adult <br> Survival (at Prosser Dam) | Productivity |  |  | Life History Diversity |
| Diagnostic Area | Smolts at Prosser | Adults at Prosser | Adults on Spawning Grounds |  | Smolts per Spawner (at Prosser) | Adult Returns per Spawner (at Prosser) | Adult Returns per Spawner (on Spawning Grounds) |  |
| Satus Creek Watershed | 53,568 | 799 | 747 | 1.5\% | 134 | 2.3 | 2.2 | 42\% |
| HISTORICAL/NORMATIVE CONDITIONS |  |  |  |  |  |  |  |  |
|  | Equilibrium Abundance |  |  | Smolt-to-adult Survival (at Prosser Dam) | Productivity |  |  | Life History Diversity |
| Diagnostic Area | Smolts at Prosser | Adults at Prosser | Adults on Spawning Grounds |  | Smolts per Spawner (at Prosser) | Adult Returns per Spawner (at Prosser) | Adult Returns per Spawner (on Spawning Grounds) |  |
| Satus Creek Watershed | 180,388 | 14,460 | 13,671 | 8.0\% | 231 | 21.1 | 20.4 | 98\% |

## Satus Creek Summer Steelhead

Relative Importance Of Geographic Areas For Protection and Restoration Measures

| Geographic Area | Protection benefit | Restoration benefit | Change in Abundance with |  | Change in Productivity with |  | Change in Diversity Index with |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Category/rank | Category/rank | Degradation | Restoration | Degradation | Restoration | Degradation | Restoration |
| Columbia above Estuary | 4 | 1 |  |  |  |  |  |  |
| Satus Cr.-2 | 4 | 4 |  |  |  |  |  |  |
| Logy Cr . | 6 | 3 |  |  |  |  |  |  |
| Satus Cr.-4 | 2 | 7 |  |  |  |  |  |  |
| Columbia Estuary | 8 | 2 |  | $\square$ |  | , |  |  |
| Satus Cr.-1 | 1 | 9 |  |  |  |  |  |  |
| Satus Cr.-3 | 9 | 5 |  |  |  | , |  |  |
| Yakima Prosser Dam to Satus | 3 | 13 |  |  |  |  |  |  |
| Dry Cr. (Satus)-2 | 7 | 11 |  | ] |  |  |  |  |
| Satus Cr.-7 | 10 | 9 |  |  |  |  |  |  |
| Dry Cr. (Satus)-1 | 14 | 8 |  |  |  | , |  |  |
| Mule Dry Cr. | 19 | 5 |  |  |  |  |  |  |
| Satus Cr.-6 | 12 | 12 |  |  |  |  |  |  |
| Satus Cr.-5 | 11 | 14 |  |  |  |  |  |  |
| Kusshi Cr. | 13 | 15 |  |  |  |  |  |  |
| Yakima delta to Horn Dam | 15 | 16 |  |  |  |  |  |  |
| Bull Cr. | 15 | 19 |  |  |  |  |  |  |
| Yakima Horn Dam to Benton | 17 | 18 |  |  |  |  |  |  |
| Wilson Charlie Cr. | 19 | 17 |  |  |  |  |  |  |
| Yakima Chandler Bypass Reach | 18 | 19 |  |  |  |  |  |  |
| Yakima Benton to Powerplant | 21 | 21 |  |  |  |  |  |  |
| Yakima delta | 22 | 22 |  |  |  |  |  |  |
| Coastal Zone | 23 | 23 |  |  |  |  |  |  |
| Offshore Marine | 23 | 23 |  |  |  |  |  |  |
| $\begin{array}{ccc} -100 \% & 0 \% & +100 \% \\ \text { Percentage change } \end{array}$ |  |  |  |  | $\begin{gathered} -100 \% ~ 0 \% ~+100 \% \\ \text { Percentage change } \end{gathered}$ |  | $\begin{aligned} & -100 \% \quad 0 \% \quad+100 \% \\ & \text { Percentage change } \end{aligned}$ |  |

Figure 8. Summary of the relative importance of habitat (geographic area) restoration and preservation to steelhead in the Satus basin.

Table 9. Summary of Yakima basin fall chinook performance (by diagnostic area) in terms of equilibrium abundance, productivity and life history diversity for current and historical conditions.

| Baseline Performance Analysis, Yakima Fall Chinook |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CURRENT CONDITIONS |  |  |  |  |  |  |  |  |
|  | Equilibrium Abundance |  |  | Smolt-to-adult Survival (at Yakima mouth) | Productivity |  |  | Life History Diversity |
| Diagnostic Area | Adults on Spawning Grounds | Adults at Yakima mouth | Smolts at Yakima mouth |  | Smolts per Spawner (at Yakima mouth) | Adult Returns per Spawner (at Yakima mouth) | Adult Returns per Spawner (on Spawning Grounds) |  |
| Lower Yakima Mainstem (mouth to Sunnyside Dam): Natural Spawners | 3.150 | 4.309 | 617.533 | 0.7\% | 224 | 2.3 | 1.7 | 9.0\% |
| Lower Yakima Mainstem: F1 Hatchery Fish | 1.863 | 2.469 | 791,196 | 0.3\% | 851 | 2.8 | 2.1 | 31.0\% |
| Marion Drain | 104 | 155 | 4.169 | 3.7\% | 54.2 | 2.1 | 1.7 | 8.0\% |
| Combined Hatchery/Natural Mainstem Population: Natural-origin spawners only | 4.289 |  |  |  |  |  |  |  |
| Combined Hatchery/Natural Mainstem Population: Hatchery-origin spawners only | 1,887 |  |  |  |  |  |  |  |
| Supplemented Total (including Marion Drain | 6,280 |  |  |  |  |  |  |  |
|  |  |  | HISTORICAL | NORMATIVE CON | TIONS |  |  |  |
|  |  | ibrium Abu | dance |  |  | Productivi |  |  |
| Diagnostic Area | Adults on Spawning Grounds | Adults at Yakima mouth | Smolts at Yakima mouth | Survival (at Prosser Dam) | Smolts per Spawner (at Yakima mouth) | Adult Returns per Spawner (at Yakima mouth) | Adult Returns per Spawner (on Spawning Grounds) | Life History Diversity |
| Lower Yakima Mainstem (mouth to current location of Sunnyside Dam) | 75,744 | 81,201 | 8,216,607 | 1.0\% | 396 | 22.5 | 21.8 | 100\% |

## Marion Drain Fall Chinook

Relative Importance Of Geographic Areas For Protection and Restoration Measures


Figure 9. Summary of the relative importance of habitat (geographic area) restoration and preservation to fall chinook in Marion Drain.

## Lower Yakima Mainstem Fall Chinook

Relative Importance Of Geographic Areas For Protection and Restoration Measures


Figure 10. Summary of the relative importance of habitat (geographic area) restoration and preservation to fall chinook in lower Yakima River.

## Klickitat

Work initiated in the last quarter of FY01 to delineate additional reaches/ obstructions and input of Level 2 attributes was continued in FY 02 . A biologist was hired in FY 02 to specifically construct the Klickitat EDT model. Through FY 02 the hydrography stream routing layer had been defined for the model and hydrological data was being summarized for input of the Level 2 hydrological attributes to the EDT model. All information is simultaneously being input into the GIS for future use in producing graphical outputs of the EDT products, which is considered an improvement over past methods of displaying model outputs using tables and figures.

Personnel Acknowledgements: Chris Frederiksen, Joel Hubble and William Sharp YN biologists are handling this task for Y akima and Klickitat basins.

## Task 1b Yakima River Fall Chinook Fry Survival Study

Rationale: To determine the optimal locations within the lower Yakima basin where fall chinook production is feasible, and to guide location of future acclimation/ release sites.

Methods: The feasibility of beach seining for juvenile fall chinook was initiated in 2001, with the longterm objective of initiating a PIT tag study to evaluate smolt-smolt survival between different reaches of the Y akima River. In 2001 beach seine sites were established at Toppenish, Granger and Benton City.

Progress: Growth profiles of naturally rearing fall chinook juveniles in the lower Y akima River were successfully monitored via beach seining during the month of May in 2002. High river discharge precluded any sampling during the month of April. Beach seining areas were located in nine sections of the Yakima River, at Van Giesen Street Bridge (Rm 8.4), West Richland (Rm 9), Horn Rapids Dam (Rm 18), Benton City (Rm 29.8), below Granger (Rm 6983), above Granger (Rm 83-90), Toppenish (Rm 90), Union Gap (Rm106.9116, and Sundown Ranch (Rm 123.5). Seining was conducted using a 60 ft beach seine. Areas were seined until 100 fork lengths of juvenile fall chinook salmon were gathered. Any additional fish were enumerated, identified to species and released back into the river.

The data set indicates a continued large spatial distribution of spawning fall chinook throughout the middle and lower Y akima River. Juvenile fall chinook,
were found rearing from Rm 106 at Union Gap down to the mouth of the Y akima River. The rearing juveniles throughout the river showed faster growth down river, possibly due to the warmer water.

A freshet resulting in the loss of two temperature loggers caused the 2002 temperature data set to be incomplete. However, the water temperatures observed in 2001 (Figure 11) are likely an appropriate representation of typical water temperatures at Benton City, Granger and Toppenish under which juvenile fall chinook reared in 2002.

As observed in 2001 the mean fork length of juvenile fall chinook generally increased in size from Union $G$ ap to West Richland (Table 10). The mean fork length for sites sampled upstream of G ranger was 47 mm compared to 61 mm for sites below Granger. There was also a greater range in fork lengths for sites sampled above Granger (23, 43 and 28 mm ) compared to site below Granger ( 75,60 and 58 mm ). This difference observed in the range of size could be attributed to either greater variance in fry emergence timing or from the emigration of selectively larger fish from upstream reaches.


Figure 11. Temperature comparisons of the Yakima River.

Table 10. 2002 Growth profiles of naturally rearing juvenile fall chinook salmon.

|  |  |  | Total Counts |  | Average FL |  | Max FL |  | Min FL |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | RM | April | May | April | May | April | May | April | May |  |
| Sundown | $\mathbf{1 2 3 . 5}$ |  |  |  |  |  |  |  |  |  |
| Union Gap | $\mathbf{1 0 6 . 9 - 1 1 6}$ |  | 83 |  | 46 |  | 60 |  | 37 |  |
| Toppenish | $\mathbf{9 0}$ |  | 164 |  | 52 |  | 78 |  | 35 |  |
| Above Granger | $\mathbf{8 3 - 9 0}$ |  | 94 |  | 43 |  | 61 |  | 33 |  |
| Below Granger | $\mathbf{6 9 - 8 3}$ |  | 307 |  | 48 |  | 106 |  | 31 |  |
| Benton City | $\mathbf{2 9 . 8}$ |  | 248 |  | 69 |  | 100 |  | 40 |  |
| Horn Rapids | $\mathbf{1 8}$ |  |  |  |  |  |  |  |  |  |
| West Richland | $\mathbf{9}$ |  | 178 |  | 67 |  | 98 |  | 40 |  |
| Van Giesen | $\mathbf{8 . 4}$ |  |  |  |  |  |  |  |  |  |

Of interest were the 414 wild juvenile coho, presumably all age- 0 that were captured. Their mean fork length was 48 mm (standard deviation was 7.4 ) and ranged in length from 33 to 75 mm (Figure 12). The Age-0, wild coho were


Figure 12. Length (fork) histogram of age-0 wild coho captured in the fall chinook beach seining surveys on the Yakima River, spring 2002.
most prevalent in the Toppenish reach (205 fish); followed by the Granger reach (119 fish) and the Union Gap reach ( 90 fish). No wild juveniles were captured in the Benton City reach. This distribution follows closely the distribution of spawning in the mainstem Yakima River based on current coho spawner radio telemetry studies.

Personnel Acknowledgements: Todd Newsome is the project biologist for this task. Technicians Linda Lamebull, Joe Jay Pinkham, Jason Allen, Conan Northwind and Wilda Watlamet conducted all field activities.

## Task 1.c Yakima River Juvenile Spring Chinook Micro-habitat Utilization

The WDFW annual report for this task can be located on the BPA website: http:/ / www.efw.bpa.gov/ cgi-bin/ FW/ publications.cgi

Pearsons, T.N., B. James, C. L. Johnson, A. L. Fritts, and G. M. Temple. 2003. Spring chinook salmon interactions indices and residual/ precocial monitoring in the upper Yakima River. Annual Report FY 2001-2002 submitted to Bonneville Power Administration, Portland, Oregon. DOE/BP-00004666-14.

## Task 1.d Yakima River Juvenile Spring Chinook Marking

Rationale: Estimate hatchery spring chinook smolt-to-smolt survival at CJMF and Columbia River projects, and smolt-to-adult survival at Bonneville (PIT tags) and Roza (PIT and CWT) dams.

Method: To estimate smolt-to-smolt survival by rearing treatment (OCT/ SNT), acclimation location and raceway, we PIT tagged and adipose clipped the minimum number to determine statistically meaningful differences detected at CJMF and lower Columbia River projects. The remaining fish will be adipose fin clipped and tagged with multiple body placement coded wire tags unique for rearing treatment, acclimation location, and raceway. Returning adults that are adipose clipped at Roza Dam Broodstock Collection Facility (RDBCF) will be interrogated using a hand-held CWT detector to determine the presence/ absence of body tags. We will recover CWT during spawning
ground surveys. We will use ANOVA to determine significant differences between groups for both smolt-to-smolt and smolt-to-adult survival.

Progress: Tagging of brood year 2001 fish began at the Cle Elum hatchery on O ctober 14, 2002 and was completed on November 15, 2002. Marking results are summarized in Table 11. This brood suffered a significant BKD mortality at pre-spawn/ spawn time, so had fewer fish in the rearing ponds for marking. As in prior years, all fish were adipose fin-clipped. Approximately 4,000 fish ( $9.3 \%$ to $10.2 \%$ of the fish) in each of 8 raceways were CWT tagged in the snout and then PIT tagged. For the Predator Avoidance Training raceways (PAT), a total of 8,000 fish (about 20\% of these fish) were CWT tagged in the snout and then PIT tagged. The remainder of the fish $(334,360)$ had a CWT placed in their body (i.e. left/ right cheek, anterior/ posterior dorsal fin, caudal fin and adipose fin) and a colored elastomer dye placed into the adipose eyelid. The three colors of elastomer dye in the adipose eyelid corresponded to the three acclimation sites (red = Clark Flat, green $=$ Jack Creek and orange $=$ Easton). Fish with the elastomer dye in the left eyelid corresponded to the OCT treatment and the right eyelid to the SNT treatment. The six different CWT body tags corresponded to the rearing raceway (numbers 1-6, 7-12 and 13-18) at the Cle Elum Hatchery. A final quality control check by YN staff took place in late D ecember, 2002.

Table 11. Summary of 2001 brood year marking activities at the Cle Elum Supplementation and Research Facility.

| CE RW ID | Treatment | Acclld | Comment | Est. Number | Elastomer Eye Site/Color | CWT <br> Body site | Total No. Tagged | Total No. PIT-Tagged | Grand total Tagged | Start <br> Date | Finish Date |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CLE01 | OCT | CFJ04 | BKD 6 | 41727 | Right/Red | Anterior dorsal fin | 38809 | 4000 | 42809 | 10/31/2002 | 11/6/2002 |
| CLE02 | SNT | CFJ03 | BKD 6 | 41733 | Left/Red | Posterior dorsal fin | 38496 | 4000 | 42496 | 11/6/2002 | 11/12/2002 |
| CLE03 |  |  |  |  |  |  |  |  |  |  |  |
| CLE04 |  |  |  |  |  |  |  |  |  |  |  |
| CLE05 | SNT | JCJ03 | Production | 40174 | Left/Green | Posterior dorsal fin | 37765 | 4017 | 41782 | 10/14/2002 | 10/17/2002 |
| CLE06 | OCT | JCJ04 | Production | 41127 | Right/Green | Anterior dorsal fin | 36700 | 4000 | 40700 | 10/17/2002 | 10/23/02 |
| CLE07 | SNT | CFJ01 | Production | 42638 | Left/Red | Right cheek | 39081 | 4000 | 43081 | 10/23/2002 | 10/29/02 |
| CLE08 | OCT | CFJ02 | Production | 41829 | Right/Red | Left cheek | 39048 | 4000 | 43048 | 10/29/2002 | 11/04/02 |
| CLE09 | SNT | CFJ05 | Production | 41001 | Left/Red | Caudal peduncle | 37655 | 4001 | 41656 | 11/5/2002 | 11/08/02 |
| CLE10 | OCT | CFJ06 | Production | 40126 | Right/Red | Adipose fin | 35321 | 4000 | 39321 | 11/12/2002 | 11/15/02 |
| CLE11 |  |  |  |  |  |  |  |  |  |  |  |
| CLE12 |  |  |  |  |  |  |  |  |  |  |  |
| CLE13 | PAT | ESJ01 | Control | 6427 | Left/Orange | Right cheek | 3618 | 1333 | 6788 | 10/31/2002 | 10/31/02 |
| CLE14 | PAT | ESJ02 | Treatment | 6427 | Right/Orange | Left cheek | 3587 | 1333 | 6585 | 10/30/2002 | 10/30/02 |
| CLE15 | PAT | ESJ03 | Control | 6427 | Left/Orange | Posterior dorsal fin | 3280 | 1336 | 6314 | 10/30/2002 | 10/30/02 |
| CLE16 | PAT | ESJ04 | Control | 6427 | Left/Orange | Anterior dorsal fin | 3248 | 1333 | 6493 | 10/29/2002 | 10/29/2002 |
| CLE17 | PAT | ESJ05 | Treatment | 6427 | Right/Orange | Caudal peduncle | 3452 | 1334 | 6678 | 10/29/2002 | 10/29/02 |
| CLE18 | PAT | ESJ06 | Treatment | 6427 | Right/Orange | Adipose fin | 3218 | 1333 | 6627 | 10/28/2002 | 10/28/02 |

## Task 1e Roza Juvenile Wild/ H atchery Spring Chinook Smolt PIT Tagging

Rationale: To capture and PIT tag wild and hatchery spring chinook to estimate, 1) wild and hatchery smolt-to-smolt survival to CJMF and the lower Columbia River projects, and 2) to estimate differential smolt-adult survival between winter versus a spring migrant fish.

Methods: The Roza Dam juvenile fish bypass trap was used to capture wild and hatchery spring chinook pre-smolts. The trap was operated from December 19, 2001 and ended on April 26, 2002. The trap was fished five days per week, 24 hours per day. Fish were removed from the trap each morning and PIT tagged on site and released the following day after recovery.

Progress: A total of 10,221 ( 8,718 wild and 1,503 hatchery) juvenile spring chinook were PIT tagged from fish collected at the Roza juvenile fish bypass trap. A maximum of 250 fish were tagged per day. Wild fish were tagged from December 19, 2001 through April 26, 2002; and hatchery fish March 18 through April 26, 2002.

## Task 1f Yakima River Wild/ H atchery Salmonid Survival and Enumeration (CJMF)

Rationale: As referenced in the Y K FP Monitoring Plan (Busack et al. 1997), CJMF is a vital aspect of the overall M\&E for YKFP. The baseline data collected at CJMF includes: stock composition of smolts, outmigration timing, egg-to-smolt and/ or smolt-to-smolt survival rates, hatchery-v-wild and hatchery optimum conventional treatment (OCT) reared fish-v-hatchery seminatural treatment (SNT) reared fish survival rates (spring chinook). Monitoring of these parameters is essential to determine whether post-supplementation changes are consistent with increased natural production. This data can be gathered for all anadromous salmonids within the basin.

In addition, the ongoing fish entrainment study is used to refine smolt, both present and historic, as adjustments are made to the CJMF fish entrainment to river discharge logistical relationship.

The facility also collects steelhead kelts for the kelt reconditioning project, and conducts trap and haul operations when conditions in the lower Y akima are not favorable to smolt survival.

Methods: The CJMF is operated on an annual basis, with smolt enumeration efforts conducted from late winter through early summer corresponding with salmonid smolt out-migrations. A sub-sample of salmonid outmigrants is biosampled on a daily basis and all PIT tagged fish interrogated.

Replicate releases of PIT tagged smolts were made in order to estimate the fish entrainment and canal survival rates in relation to river conditions. The entrainment rate estimates were used in concert with a suite of independent environmental variables to generate a multi-variate smolt passage relationship used to develop current, future and passage estimates with confidence intervals.

Hand held CWT detectors were used to scan for body-tags on hatchery spring chinook smolts. This is a monitoring and evaluation protocol is built in as a backup in the event that the corresponding PIT tagged fish from each treatment group (OCT/ SNT) failed to be accurately detected by the PIT detectors stationed at the CJMF. Fortunately, there was good correspondence between the detection rates between the two mark groups.

Progress: The 2002 smolt passage estimates were as follows: wild spring chinook-367,006; OCT spring chinook- 193,430 (Easton: 52,835; Jack Creek: 66,224 and Clark Flat: 74,371); SNT spring chinook- 132,232 430 (Easton: 26,469; Jack Creek: 57,502 and Clark Flat: 48,261); wild fall chinook- 41,571; Marion Drain hatchery fall chinook- 0; wild coho- 19,793; hatchery coho30,006; and wild steelhead- 38,509. These estimates are provisional and subject to change as better entrainment estimates are developed.

Personnel Acknowledgements: Biologists Mark Johnston and D avid Lind; and Fisheries Technician Leroy Senator are, respectively, the project supervisors and on-site supervisor of CJMF operations. Other Technicians that assisted are Sy Billy, Wayne Smartlowit, Morales Ganuelas, Pharamond Johnson, Steve Salinas, Shiela D ecoteau, Jimmy Joe O lney and Tammy Swan.

## Task 1.g Yakima River Fall Chinook Monitoring \& Evaluation

Rationale: To determine the optimal release timing (April vs. May) to increase overall smolt and smolt-to-adult survival

Method: Approximately 365,409 fall chinook smolts were produced from fish spawned during the fall of 2002. These smolts were divided into two equal groups. One group was reared using conventional methods using ambient river
temperature incubation and rearing profiles. The other group was incubated and reared using warmer well water to accelerate emergence and rearing and ultimately smoltification. Both groups of fish were spawned, incubated and reared at the Prosser Hatchery. Fish from both groups were 100\% marked using ventral fin clips (pelvic fins), and approximately 2,000 fish from each group were PIT tagged to evaluate survival and migration timing to the lower Columbia River. Approximately 1,000 PIT tagged Marion Drain hatchery fall chinook juveniles were released to estimate survival from Marion Drain Hatchery to CJMF and McNary D am.

Progress: Yakama Nation collected a total of 130 fall chinook broodstock between Prosser D am D enil ladder and from fish taken from Chandler canal at Prosser. This resulted in 365,409 smolts that were split into two groups of approximately 165,000 accelerated incubation and rearing, and 100,000 incubated and reared on ambient river water (conventional group). All fish were ventral clipped, either left (conventional group) or right (accelerated group), to distinguish treatment groups as returning adults at Prosser Dam (video monitoring) and from carcasses recovered by WDFW during their fall chinook redd surveys conducted downstream of Prosser D am. A total of 1,000 PIT tagged fish were marked from each of the two treatment groups (nonaccelerated and accelerated) in order to estimate smolt-smolt survival to the lower Columbia River. There was no significant difference in the smolt-atCJMF to smolt-at-McNary D am survivalindex between the accelerated (0.22) and conventional (030) groups (Neeley, 2003).
The survival indice for the Marion Drain conventional group was 0.30 and was not significantly different to either of the two Prosser released groups.

## Task 1h Yakima River Coho Optimal Stock, Temporal, and Geographic Study

Rationale: To determine the optimal location, date, and stock of release to maximize the feasibility of coho re-introduction into the Yakima Basin, and to determine the spawning distribution of returning adults.

Method: A nested factorial experimental design was intended to be used to test for survival differences between out of basin hatchery and Prosser Hatchery stocks; release location (upper Yakima and Naches subbasins); and release date (May 7 and May 31). Each release date had two replicates per sub-basin). Within each replicate 2,500 coho smolts were PIT tagged (1,250
out of basin stock and Prosser Hatchery stock were intended to be PIT tagged) to evaluate survival to CJMF and lower Columbia projects. In addition to PIT tags to monitor juvenile survival, a portion of the smolts were CWT'ed in order to assess the survival of returning adult to Prosser Dam. Beginning with the 1997 broodyear $100 \%$ of the locally produced and out of basin smolts have been CWT in order to monitor smolt-adult survival, and relative wild contribution of both smolt and adult coho production. The 2000 returning adults was the first year where wild and hatchery smolt-adult return rates could be compared. In order to determine the relative abundance of hatchery coho smolt residuals, we conducted surveys in the upper Yakima and Naches rivers to enumerate coho that did not migrate during the spring. Since 1999 about 98 spawners have been radio tagged at Prosser Dam to evaluate spawning distribution. In 2000105 fish were tagged and 75 fish were successfully tracked until spawning.

Progress: The first hatchery smolt release under the auspices of Phase I of the coho feasibility study occurred in 1998. Completion of Phase I will occur in the fall of 2003 with the adult returns from the 2002 smolt release (BY 2000). A complete summary of Phase I results will be reported in next year's FY 2003 YKFP M\&E annual report. The experimental design for Phase II of the coho feasibility study is near completion at the time of this writing and will be reported on in the YK FP Yakima Coho Master Plan to be made public in the summer of 2003 (2003)

The Yakima stock, late-release (pooled across release sites) survival index (to McNary Dam) was significantly higher than that of the early-released group; however, there was no difference between releases for the Willard stock. For the Stiles Pond site the late-release groups (pooled across stocks) had a significantly higher survival index value to that of the early-release groups. The Yakima stock (pooled across pooled across Naches sites) late-release survival index was significantly greater than for the Willard late-release groups.

The Yakima stock late-release had the highest mean survival index (0.405) followed by Willard late-release (0.2482), Willard early-release (0.200) and Y akima early-release (0.1713) (Table 12). As observed last year smolt releases from Stiles pond on the Naches had the highest survival index (0.3541), followed by Lost Creek pond (0.2395) and Easton pond (0.0580, $6.1 \times$ lower than for Stiles pond).

Table 12. Summary of release-to-McNary survival index by stock, timing and location ${ }^{1}$.

| Survival Index |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Site | Willard <br> Early | Willard <br> Late | Yakima <br> Early | Yakima <br> Late | Mean |  |
|  |  |  |  |  |  |  |
| Easton | 0.0580 | 0.1971 | 0.0154 | 0.0100 | 0.0580 |  |
| Lost Creek | 0.2492 | 0.1317 | 0.2297 | 0.4002 | 0.2395 |  |
| Stiles Pond | 0.2928 | 0.4153 | 0.2688 | 0.8059 | 0.3541 |  |
| Mean | 0.2000 | 0.2482 | 0.1713 | 0.405 |  |  |

- We estimated that wild smolt-to-adult survival rate for 40,605 natural origin coho smolts (counted at CJMF) in 2001 was $.87 \%$, which was 42 times greater than that observed for hatchery smolts.
- We estimated that hatchery smolt-to-adult survival rate for 442,249 hatchery coho smolts (counted at CJMF) released in the Naches and upper Y akima rivers in 2001 was $.04 \%$.
- The 2002 adult coho run was comprised of $65 \%$ (352fish) naturally produced fish and $35 \%$ hatchery fish. This was the second year where this distinction could be made due the $100 \%$ CWT'ing of smolts beginning with the 2000 release.
- Smolt-smolt survival (CJMF to McNary Dam) was higher for the Y akima stock (mean $=43.5 \%$ ) than for the Willard stock (mean=27\%) in 2001. Reasons for this are not readily understood at this time.
- There was no significant difference in smolt-smolt survival (CJMF to McNary Dam ) between the early and late release groups for either basin. The lack of a differential survival difference between the two groups is most likely due to the extremely poor outmigration conditions, which persisted the entire smolt outmigration period in 2001.

[^0]- A total of 105 coho spawners were radio tagged at Prosser Dam in the fall of 2001, of which 75 were subsequently successfully tracked. The spawner distribution throughout the Yakima basin was as follows: Prosser Dam (rm 47.1)-G ranger (m 83.0)- 6.7\%, G ranger-Sunnyside D am reach (rm 103.8)- 37.1\%, Sunnyside Dam-Naches River (rm 116.3)$5.7 \%$, mid-Yakima River tributaries- $4.8 \%$, lower Naches River- $3.8 \%$, Naches River above Cowiche Dam (rm 2.7)- 13.3\%, Naches River confluence to above Roza D am (rm 127.9)- 9.5\%.
- Since 1999 all smolts have been released in the Naches and the upper Yakima rivers, and in 1998 a portion of the smolts were released from Lost Creek in the upper Naches River. Despite this, the majority of spawning appears to occur in the Yakima River downstream to the Naches River confluence. It's believed that three factors are contributing to this, 1) lack of stamina primarily by females to reach their areas of release located further upstream, 2) straying and delay due to false attraction from irrigation return flow and 3) from natural production occurring in the Y akima River above G ranger. Nevertheless, with the exception of 2002 (9\%), the percentage of spawners returning to the Naches River has steadily increased from $8.2 \%$ in 1999 to $26.7 \%$ in 2001. Correspondingly the percentage of fish spawning in the Granger to Sunnyside Dam reach has decreased from 61.6\% in 1999 to $37.1 \%$ in 2001, to $19 \%$ in 2002. In addition, nearly $13 \%$ of radio tagged coho spent various amounts of time in Sulfur D rain.
- Residual coho smolt survey sites on the upper Yakima River (Easton reach) were from the Easton acclimation site (Rkm 325.4) to the confluence of Cle Elum River (Rkm 294.6). The Naches River (Lost Cr. reach) surveys were done from the Lost Creek acclimation site (Rkm 61.8) to the confluence with Rock Creek (Rkm 53.9). In 2002, residual coho were generally absent from all snorkel surveys. One residual coho was seen in the Lost Cr. Reach, which equates to less than 0.25 fish per river kilometer. No residuals were observed in the upper Y akima River reach. Sub-yearling coho were generally absent from index areas, however, there were some small numbers of sub yearling coho found in adjacent areas, indicating continued natural production. Results in 2002 are consistent with those the past two years, where relatively low densities of residuals and sub-yearlings were observed in both subbasins.

Personnel Acknowledgements: They are the same as for Task 1.i with the following additions. PIT tagging occurred at Prosser Hatchery with assistance
from Biologist Mark Johnston and Fisheries Technicians Leroy Senator, Tammy Swan, Sy Billy, Joe Hoptowit, and Gerry Lewis.

## Task 1.i Yakima Spring Chinook Juvenile Behavior

Rationale: This Three year study (1999-2001) is part of an effort to evaluate the rearing of spring chinook salmon (0 norhynchus tshawytscha), at the Cle Elum Supplementation and Research Facility. Yearling spring chinook (0 norhynchus tshawytscha) smolts from two hatchery treatment groups, conventional and semi natural rearing treatments, were compared to wild smolts in an experiment designed to assess differences in cover utilization, and survival to a predation (pikeminnow, Ptychocheilus oregonensis) threat.

Methods: Groups of five smolts from each of the three treatment groups, (Wild, OCT \& SNT), were placed sequentially into an aquarium. Cover and a predation threat were present in the aquarium. Typically, upon introduction, smolts will dive for cover and remain hidden for several minutes before emerging to explore or school with other smolts. Observers recorded the amount of time smolts spent in cover and made qualitative assessments of the smolt's cover utilization. Northern Pikeminnows, Ptoccheilus oregonensis, were then allowed to feed on the smolts until approximately one-half were consumed. Surviving smolts were then counted and measured by treatment group.

Progress: Yearling Spring Chinook smolts, 0 nchorhynchus tshawytsha, from two hatchery treatment groups, conventional, (OCT), and semi natural (SNT), rearing treatments were compared to wild smolts in an experiment designed to assess differences in cover utilization, and survival to a predation threat. Survival to Northern Pikeminnows was seen to be size dependant for hatchery fish, ( $\mathrm{p}=0.001$ ), with the largest fish $(141-158 \mathrm{~mm})$ surviving at over twice the rate as the smallest fish $(90-120 \mathrm{~mm})$. Survival was not size dependant for wild fish however, ( $\mathrm{p}=0.713$ ). O verall survival rates were similar between the three groups, although wild smolts tended to be smaller. Among the smaller smolts ( $<=130 \mathrm{~mm}$ ), wild smolts survived at higher rates, and rate was significantly different from the OCT group, ( $\mathrm{p}=0.033$ ).

The order of introduction did not significantly affect the time any of the three groups of smolts remained in cover, indicating that the presents or absence of other smolts did not influence a newly introduced smolts decision on how long to remain in cover. No significant difference was found between the two
hatchery treatments in time spent in cover. The semi-naturally reared smolts spent the least time in cover, and the difference from the wild was significant, ( $p=0.023$ ). Qualitative observations also revealed little difference between the conventional and semi-naturally reared smolts. In comparison to wild smolts, hatchery smolts appeared less adept at finding and concealing themselves in cover. Wild smolts also tended to swim less, i.e. in cover they appeared nearly motionless, whereas hatchery fish were almost always swimming.

Personnel Acknowledgements: John McConnaughey, (YKFP Research Center) and Dr. Terry DeVietti, (CWU Psychology Dept).

## Task 1j Yakima Spring Chinook Juvenile Morphometric/ Coloration

The WDFW annual report for this task can be located on the BPA website: http:/ / www.efw.bpa.gov/ cgi-bin/ FW/ publications.cgi

Schroder, S.L., C.M. Knudsen, B. Watson, T. Pearsons, S. Young and J. Rau. 2003. Comparing the reproductive success of Yakima River hatcheryand wild-origin spring chinook. Annual Report 2002, Project Number 1995-064-24.

## Task 11 Adult Salmonid Enumeration at Prosser Dam

Rationale: To estimate the total number of adult salmonids returning to the Yakima Basin by species (spring and fall chinook, coho and steelhead), including the estimated return of externally marked fish (i.e., adipose clipped fish). In addition, biotic and abiotic data is recorded for each fish run.

Methods: Monitoring was accomplished through use of timelapse video recorders (VHS) and a video camera located at each of the three fishways. The videotapes were played back and various types of information/ data were recorded for each fish that migrated past, and data was entered into the Y KFP database.

## Progress: <br> Spring Chinook (2002 run)

Estimated 14,771 spring chinooks were counted past Prosser Dam. The total adult count was 14,054 (95.1\%) fish, while the jack count was 717 (4.9\%) fish. Of the adult count, 7,762 were identified as hatchery origin. Returning hatchery adults this year comprised 4 and 5 year olds (brood years 1997 and 1998). The ratio of wild jacks to hatchery jacks was $48.5 \%$ to $51.5 \%$, respectively.

The $25 \%, 50 \%$ and $75 \%$ dates of cumulative passage were May 11, May 15 and May 21, respectively.

The estimated mean fork length for adults (wild and hatchery) and jacks (wild and hatchery) was 67.6 cm and 51.8 cm , respectively. The estimated video fork length for adults was 3 cm smaller than that measured "hands-on" at Roza in the broodstock collection. The difference between jacks was 4.7 cm bigger than those collected at Roza. This suggests that video based fork lengths at Prosser are not a reliable measurement to estimate true fork length. It's believed this is a result of a "mismatch" in the applied multiplier value (video length x multiplier value $=$ true length) relative to the horizontal passage trajectory of the fish as it passes by the viewing window.

## Fall Run (coho and fall chinook)

## Coho (2002)

The estimated coho run was 818 fish. It should be mentioned that an undetermined number of fish "dropped out" below Prosser Dam and are not reflected in this count. Some fish were harvested while others were falsely attracted into tributaries such as Spring Creek. Adults comprised 58.1\% and jacks $41.9 \%$ of the run. A total of 118 adipose clipped fish were counted, 60 were adults and 58 were jacks. Of the estimated run, $41.8 \%$ were processed at the Denil.

The $25 \%, 50 \%$ and $75 \%$ dates of cumulative passage were October 5, O ctober 19 and November 8, respectively.

The estimated mean adult and jack fork length was 64.7 cm and 35.2 cm , respectively, which is smaller than measured fish collected for broodstock. This indicates a size bias (underestimate) of the true fork length for fish measured from the videotapes. This same bias has been observed in past years for all salmonid species at Prosser D am.

## Fall Chinook (2002 run)

Estimated fall chinook passage at Prosser Dam was 6,241 fish. Adults comprised $98.5 \%$ of the run, and jacks 1.5\%. Of the total number of fish, 681 were adipose clipped, 669 fish were adults and 12 fish were jacks. The median passage date was October 13, while the $25 \%$ and $75 \%$ dates of cumulative passage were September 15 and October 23, respectively. Of the total fish estimate, $15.5 \%$ were counted at the D enil.

The mean adult and jack fork length was 75 cm and 49 cm , respectively.
Steelhead (2001-02 run)
The estimated steelhead run was 4,525 fish. Of the total, 34 adipose clipped fish, which were all out-of-basin strays since no hatchery returns were expected to the Yakima River. The median passage date was November 19th, 2001, while the $25 \%$ and $75 \%$ cumulative dates of passage were October 30th, 2001 and January 15th, 2002 respectively.

The mean fork length was 56.7 cm , and fish ranged in size from 39.9 cm to 85 cm.

Personnel Acknowledgements: Biologists, Melinda D avis and Joel Hubble, and Fisheries Technicians Winna Switzler, Florence Wallahee and Sara Sohappy.

## Task 1m Adult Salmonid Enumeration and Broodstock Collection at Roza/ Cowiche Dams.

Rationale: The purpose is to estimate the total number of adult salmonids returning to the upper Yakima Basin for spring and fall chinook, coho and steelhead) at Roza D am, and for coho only into the Naches Basin at Cowiche Dam . This includes the count of externally marked fish (i.e., adipose clipped). In addition, biotic and abiotic data is recorded for each fish run.

Methods: Monitoring was accomplished through use of timelapse video recorders (VHS) and a video camera located at each fishway. The videotapes were played back and various types of information/ data are recorded for each fish that passes. Spring chinook passing Roza Dam are virtually entirely enumerated through the Cle Elum Supplementation and Research Facility broodstock activity.

## Progress:

Roza Dam

## Steelhead

A total of 216 steelhead were counted past Roza D am for the 2001-02 run. As shown in Figure 13, most steelhead migrated past Roza D am from late March through early May of 2002.

## Spring Chinook

At Roza D am 8,922 (98\% adults and 2\% jacks) spring chinook were counted at the adult facility between May 3 and September 7, 2002. The adult return was comprised of natural ( $28 \%$ ) and CESRF-origin (72\%) fish. The jack return was comprised of natural (61\%) and CESRF-origin (39\%) fish. Figure 14 shows passage and wild brood collection timing at Roza in 2002.

## Coho

A total of 5 adult and 1 jack coho were observed passing Roza Dam from November 18, 2002 through January 7, 2003. Of the total, 2 adults and 1 jack were observed to have a CWT in the snout (hatchery-origin).

Cowiche Dam

## Coho

The persistence of moderate turbidity levels resulting primarily from Tieton River water releases through the most of the coho upmigration period negated the opportunity to video monitor adult counts for spawning coho in 2002.

## Task 1.n Spawning Ground Surveys (Redd Counts)

Rationale: To enumerate the temporalspatial distribution of spring chinook, fall chinook, steelhead and coho redd deposition in the Klickitat and Yakima basins. To collect biological information from spawned out carcasses.

Methods: Regular foot and/ or boat surveys were conducted within the established geographic range for each species (this is increasing for coho as acclimation sites are located upriver and as the run increases in size). Redds were individually marked during each survey and carcasses are sampled to collect-egg retention, scale sample, sex, body length and to check for possible experimental marks.


Figure 13. Daily steelhead passage at Roza Dam, 2001-02.


Figure 14. Daily spring chinook passage for CESRF-origin, natural, and broodstock collected at Roza Dam, 2002.

Progress: A summary of the spawning ground surveys by species are as follows-

Steelhead: Steelhead surveys in Satus and Toppenish basins and Ahtanum Creek began in mid-March and end in late April. Total redd counts by subbasin were as follows: Satus basin- 172, Toppenish basin- 354, and Ahtanum Creek- 8. For all three basins a total of 534 redds were counted.

Spring Chinook: Redd counts began in hate July in the American River and ended in early October in the upper Yakima River. Total counts for the American, Bumping, Little Naches, Naches, and Rattlesnake rivers were, respectively, 366, 262, 89, 203, and 23 redds. Redds counts in the upper Yakima, Teanaway and the Cle Elum rivers were, 2,441, 110 and 275, respectively. The entire Yakima basin had a total of 3,769 redds (Naches- 943 redds, upper Y akima- 2,826).

Fall Chinook: Marion Drain fall chinook surveys were conducted three times in 2001. A total of 34 redds were counted. The number of redds located for each survey was as follows: October 31-15 redds, November 10-15 redds, and November 24- 4 redds.

Coho: Surveys began in early November and ended in late December in the Y akima River basin. A total of 151 redds were located in the Y akima Basin. Surveys were concentrated where radio telemetry fish were located to maximize survey effort. Due to untimely winter freshets, river conditions prevented accurate enumeration of coho redds. Nearly all the redds were located before the first winter freshet. The redd distribution was as follows:

Y akima R.- 27 redds. Most redds were located between the Zillah Bridge and Roza D am. Two redds were located in the upper Y akima Canyon.

Naches R.- 124 redds. Most redds were located from the confluence to below the Tieton River confluence.
Ahtanum Cr.- 37 redds.
Cowiche Cr.- 10 redds.
Buckskin Cr.- 29 redds.
Teanaway R.- 0 redds.

## Task 1.p Yakima Spring Chinook Residuals/ Precocials Studies

The WDFW annual report for this task can be located on the BPA website: http:/ / www.efw.bpa.gov/ cgi-bin/ FW/ publications.cgi

Pearsons, T.N., B. James, C. L. Johnson, A. L. Fritts, and G. M. Temple. 2003. Spring chinook salmon interactions indices and residual/ precocial monitoring in the upper Y akima River. Annual Report FY 2001-2002 submitted to Bonneville Power Administration, Portland, Oregon. D OE/ BP-00004666-14.

## Task 1.q Yakima River Relative Hatchery/ Wild Spring Chinook Reproductive Success

The WDFW annual report for this task can be located on the BPA website: http:/ / www.efw.bpa.gov/ cgi-bin/ FW/ publications.cgi

Schroder, S.L., C.M. Knudsen, B. Watson, T. Pearsons, S. Young and J. Rau. 2003. Comparing the reproductive success of Y akima River hatcheryand wild-origin spring chinook. Annual Report 2002, Project Number 1995-064-24.

## Task 1.r Yakima Spring Chinook Gamete Quality Monitoring

Refer to WDFW report:
C.M. K nudsen, S.L. Schroder, T.N. Pearsons, J.A. Rau, A.L. Fritts, and C.R. Strom. 2003. Monitoring Phenotypic and Demographic Traits of upper Y akima River Hatchery and Wild Spring Chinook: G ametic and juvenile Traits. YKFP Annual Report 2002.

## Task 1.s Scale Analysis

Rationale: To determine age/ length and stock (hatchery vs. wild) composition of adult salmonids in the Y akima Basin.

Methods: Random scale samples are collected at broodstock collection sites (Prosser and Roza dams and Chandler Canal) and from spawner surveys. Acetate impressions are made from scale samples and then are read for age and
stock type using a microfiche reader. D ata is entered into the Y KFP database maintained by the D ata Management staff.

Progress: Adult scale sample results are summarized in Table 13 by species and sampling method.

Table 13. The 2002 adult scale sample data summary for salmonids in the Yakima Basin.

|  | Age 2 |  | Age 3 |  | Age 4 |  | Age 5 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Count | Length | Count | Length | Count | Length | Count | Length |
| Yakima R. Spring Chinook |  |  |  |  |  |  |  |  |
| Roza Dam Samples |  |  |  |  |  |  |  |  |
| Upper Yakima Supplementation | 475 | 18.4 | 26 | 46.8 | 1535 | 70.4 | 34 | 80.8 |
| Upper Yakima Wild/Natural |  |  | 45 | 50.7 | 525 | 72.3 | 29 | 85.1 |
| Spawner Survey Samples |  |  |  |  |  |  |  |  |
| Upper Yakima Supplementation |  |  | 6 | 56.7 | 187 | 73.7 | 24 | 84.0 |
| Upper Yakima Wild/Natural |  |  | 5 | 56.2 | 73 | 76.7 | 21 | 84.5 |
| American River Wild/Natural |  |  | 1 | 51.0 | 93 | 80.4 | 74 | 94.6 |
| Naches River Wild/Natural |  |  | 1 | 50.0 | 99 | 78.0 | 48 | 90.7 |
| Yakima R. Fall Chinook |  |  |  |  |  |  |  |  |
| Hatchery | 12 | 47.6 | 69 | 68.6 | 42 | 85.0 | 3 | 97.0 |
| Wild/Natural | 48 | 45.2 | 455 | 70.5 | 265 | 86.7 | 27 | 100.6 |
| Yakima R. Coho |  |  |  |  |  |  |  |  |
| Hatchery | 54 | 36.9 | 45 | 66.7 | 1 | 83.0 |  |  |
| Wild/Natural | 133 | 39.5 | 62 | 70.2 |  |  |  |  |

Note: Length is average fork length.

## Task l.u H abitat Monitoring Flights and Ground Truthing

Rationale: To record an aerial video record of a particular subbasin that can be used to aid in the EDT Level 2 data input to the model.

Methods: A Piper Cub 180 airplane was used in combination with a hand held digital video camera to record habitat conditions for all the major subbasins of the Klickitat Basin. The flight was conducted in late February 2003 over a two day period prior to spring leaf out. The survey was conducted at approximately 300 to 400 feet above river level. The goal was to record habitat conditions for the bankfull conditions, as well as, record habitat conditions across that portion of the floodplain inundated by moderate flood events.

The video tape was captured and stored in a digital format on a Fisheries Resource Management computer. The images are being used to calculate or estimate the best input rank value for various Level 2 EDT attributes.

Progress: Flight survey data has been used to estimate the area of habitat types and other physical attributes in tributaries where access is difficult or non existent. Small sections of these tributaries with access have been used to ground truth aerial estimates from the flight survey. Mainstem aerial survey data has been examined for spatial variances within habitat types for defined reaches. This is currently being utilized for appropriate habitat sampling points on the ground, reflecting the diversity of habitat within a given reach. Aerial surveys are also being examined in conjunction with past flight surveys to document and track geomorphic changes in the riverine system over space and time, enabling us to better understand the rivers natural tendencies.

## Task 1.w Sediment Impacts on Habitat

Rationale: To monitor stream sediment loads associated with the operation of dams and other anthropogenic factors (e.g. logging, agriculture and road building), which can increase sediment loads in stream utilized by all salmonids in the Yakima Basin.

Methods: Representative gravel samples were collected from the upper Yakima River (upstream of the Cle Elum River) and the Naches Basin in the fall of 2002. Each sample was analyzed to estimate the percentage of fine or small particles present ( $<0.85 \mathrm{~mm}$ ). The Washington State TFW program guidelines on sediments were used to specify the impacts estimated sedimentation levels have had on salmonid egg-to-smolt survival. These impacts will be incorporated in analyses of impacts of "extrinsic" factors on natural production.

Progress: A complete summary of the field data for sampled collected in the upper Yakima and Naches basins can be obtained from Jim Mathews, fisheries biologist for the Yakama Nation.

## Upper Yakima

Sixty samples were collected; with the control reach located above Lake Easton (Stampede Pass) and the treatment reaches extending from Easton to the Cle Elum River confluence. Mean percent fines ( $<0.85 \mathrm{~mm}$ ) by sample reach wereStampede Pass (control): 6.3\%, upper Easton: 11.4\%, lower Easton: 11.6\%, Elk Meadows: $10.2 \%$ and Cle Elum: 17.5.

## Naches

Thirteen sites were sampled in the Naches Basin in 2002. The mean percent fines ( $<0.85 \mathrm{~mm}$ ) in the Little Naches River (mainstem) was $14.0 \%$; North

Fork- 11.6\%; South Fork-13.1\%; Bear Creek- 12.7\% and Pyramid Creek- 11.8; Rattlesnake Creek- 12.5\% and in the Tieton South Fork-17.3\%.

## Task 1.x Predator Avoidance Training

Rationale: Hatchery fish have been shown to be more susceptible to predation than wild counterparts and it has been suggested that hatchery fish lack skills required to avoid predators (Wiley et al. 1993; Olla et al. 1994; Maynard et al. 1995).

Method: Predator avoidance training will consist of introducing a hungry hooded or red-breasted merganser into a cage submerged in a raceway three times per week for three weeks prior to release. The predator will be allowed to feed for 30 minutes. The design will consist of SNT fish randomly divided into control and treatment PIT tagged groups. Survival both groups will be estimated at CJMF and McNary and John D ay dams.

Progress: Initial predator avoidance training took place at the Cle Elum Hatchery in August, 2002. Three training sessions took place where experimental raceways were exposed to a hooded merganser for 30 minutes. The training resumed at the Easton acclimation site in February, 2003, for a total of 14 separate sessions using red-breasted mergansers. Direct observations were taken during all sessions from behind tall, camouflage platforms set next to experimental ponds. Distance observations of fish from the predator were recorded three times per minute for 30 minutes, made possible by 1 foot square green and white checkered grid patterns painted on the avian cage surfaces, as well as a description of predator activity. Fish from all ponds were force-out released on March 28, 2003. Provisional survival indexes will be available in a thesis/ report format late fall, 2003.

## Task 1.y Biometrical Support:

Doug Neeley of IntSTATS was contracted by the YKFP to conduct the following statistical analyses-

- Annual Report: Smolt Survival to McNary of Year-2002 Coho and Fall Chinook Releases into the Y akima Basin
- 2002 Annual Report OCT-SNT Survival
- 2002 Annual Report, Wild and Hatchery Smolt Survival of Roza Spring Chinook Releases
- 2002 Annual Report, Indirect Predation

All for reports are attached to the YKFP, M\&E annual report as appendices, and results have been incorporated within the appropriate M\&E task.

## HARVEST

## Task 2.a Yakima and Klickitat Subbasin Harvest Monitoring

Rationale: To develop a database to track the contribution of target stocks to in-basin fisheries.

Method: The two co-managers, Yakama Nation and WDFW, are responsible for monitoring their respective fisheries in both the Klickitat and Yakima rivers. Each agency employs fish monitors dedicated to creel surveys and/ or fisher interviews at the most utilized fishing locations and/ or boat ramps. From these surveys, standard techniques are employed to expand fishery sample data for total effort and open areas and times to derive total harvest estimates. Fish are interrogated for various marks. This information is used along with other adult contribution data (i.e. broodstock, dam counts, spawner ground surveys) to determine overall project success.

Progress: Yakima and Klickitat River in-basin Tribal harvest for salmon and steelhead are presented in Table 14.

Personnel Acknowledgements: Biologist Bill Bosch, Mark Johnston and Fisheries Technicians Russ Olney and Arnold Barney.

## GENETICS

Overall Objective: D evelop methods of detecting significant PAPS genetic changes in extinction risk, within-stock genetic variability, between-stock variability and domestication selection.

Table 14. A summary of Yakama Nation tributary estimated harvest in the Yakima and Klickitat subbasins, 2002.

| River | Dates | Weekly Schedule | Notes | Chinook | Jacks | Steelhead | Coho |
| :--- | :---: | :--- | :--- | ---: | ---: | ---: | ---: |
| Klickitat River | $4 / 2-6 / 1$ | Noon Tues to 6 PM Saturday | 1 | 225 | 29 | 48 | 0 |
| Klickitat River | $6 / 4-8 / 3$ | Noon Tues to 6 PM Saturday | 2 | 189 | 16 | 394 | 0 |
| Klickitat River | $8 / 6-12 / 28$ | Noon Tues to 6 PM Saturday | 3,4 | 701 | 73 | 732 | 2,623 |
| Klickitat Total | $\mathbf{4 / 2 - 1 1 / 9}$ | Noon Tues to 6 PM Saturday |  | $\mathbf{1 , 1 1 5}$ | $\mathbf{1 1 8}$ | $\mathbf{1 , 1 7 4}$ | $\mathbf{2 , 6 2 3}$ |
|  |  |  |  |  |  |  |  |
| Yakima River | $4 / 9-7 / 27$ | Noon Tues to 6 PM Saturday | 5 | 2,507 | 73 | 11 | 0 |
| Yakima River | $9 / 24-11 / 23$ | Noon Tues to 6 PM Saturday | 6 | 0 | 0 | 0 | 0 |

1. Commericail Sale allowed during Spring Zone 6 Commercial Sale Periods.
2. Summer Fishery extended through June, and considered to be addition to Spring Fishery
3. Commercial ticket landings not fully included.
4. Commercial Sale allowed for Chinook and Coho from 10/15 to 12/14
5. YKFP Staff collected Data and Bill Bosch did Harvest Estimate.
6. No Observed Effort or Catch.

Progress: All Tasks within this Section are assigned to WDFW and are reported in written progress reports submitted to BPA. These tasks are the following:

- Task 3.a Allozyme/ DNA data collection and analysis.
- Task 3.b Stray recovery on Naches and American river spawning grounds.
- Task 3.c Yakima spring chinook domestication.

The WDFW annual report for this task can be located on the BPA website: http:/ / www.efw.bpa.gov/ cgi-bin/ FW/ publications.cgi

Busack, Craig. F. Sewall, Anthony Fritts, Janet Loxterman, James Shaklee, Steven Schroder, Curtis K nud sen, Jason Rau. 2003. Genetic studies in the Y akima River Basin, Y akima/ Klickitat Fisheries Project, Monitoring and Evaluation, Annual Report 2002. Project No. 1995-06424; BPA Report D OE/ BP-00004666-13.

## ECOLOGICAL INTERACTIONS

Overall Objective: To develop monitoring methods to determine if supplementation and enhancement efforts keep ecological interactions on nontarget taxa of concern within prescribed limits and to determine if ecological interactions limit supplementation or enhancement success.

## Task 4.a Avian Predation Index

Rationale: To assess the annual impact of avian predation upon juvenile salmonid populations in the Y akima Basin.

## Method:

## Hotspot Survey- Spring

In 2002, hotspot surveys were conducted systematically, on Mondays, Wednesdays, and Fridays at Horn Rapids and Chandler Pipe, with two additional survey days at Horn Rapids during four of the survey weeks. D uring these four weeks at Horn Rapids, three different survey methods were used. These additional surveys were conducted to make comparisons between current and past survey methods. The data from the other survey methods are not included as part of this report. A total of 32 surveys were conducted at Chandler Pipe and a total of 41 surveys were conducted at Horn Rapids for the 2002 field season, which occurred between April 11 and June 28. Both sites were surveyed simultaneously by different personnel. Observations on survey days began on the nearest 15 -minute interval after sunrise and ran for eights hours, or began at midday, eight hours after the nearest 15-minute interval after sunrise, and ended on the nearest 15 -minute interval before sunset. This allowed for observations during all periods of the day, to account for the diurnal patterns of avian piscivores. Regionally calibrated tables obtained from the National O ceanic and Atmospheric Administration were used to determine sunrise and sunset times. Depending upon the length of day and start time, between seven and eight 2-hour periods existed within a single day.

The survey area for Horn Rapids Dam included 50 meters of river above the dam and 150 meters below the dam. Since the buoy located above the dam was not included within the survey area the birds resting upon the buoy were not included in abundance counts. The survey area for the Chandler Canal Bypass outfall included 50 meters of river above the outfall pipe and 150 meters of river below the outfall pipe. All birds resting upon the shoreline lateral to the specified 50 meters of river above and 150 of river meters below both hotspots were included in abundance counts.

Observations at both sites were made from shore stations. At Horn Rapids Dam observations were made from either inside or outside an automobile. At Chandler Canal Bypass observations were made from a blind, to avoid disrupting normal bird activity. The bird blind at Chandler was used intermittently due to high water conditions. Binoculars (Leica, 10x42) were used to aid in identification. At Horn Rapids Dam, survey personnel stationed
themselves on the windward bank of the river such that the preferred orientation of feeding birds, primarily gulls, was towards the observer. At the Chandler Canal Bypass outfall, altering the side of the river from which observations were made was not feasible. However, the distance from one side of the river to the other was considerably less than at Horn Rapids D am, which improved the observer's ability to accurately monitor bird behavior.

The hotspot survey design for 2002 followed the method used in 2001. Each day was divided into 2-hour survey 'windows', consisting of three, 15 -minute abundance/ feeding 'blocks'. Each of these blocks was divided by a 15 -minute period of no observation, unless a feeding interval was still being measured, in which case the observation period was extended into the next 15 minutes. This 75 -minute cycle of 'blocks' was followed by a 45 -minute rest period before beginning a new 2-hour 'window'. Within the 15 -minute survey 'blocks', abundance of all piscivorous birds, foraging ratios, the number feeding to total number present, and foraging rates, fish consumed/min, of gulls were determined. Gulls flying within the study area were considered foraging. Gulls within the study area foraging on terrestrial prey items-such as insects, seeds, plants-were not considered feeding, but were included in total abundance counts. Gulls sitting or standing on rocks emerging from the river or along the river edge were not counted as part of the foraging fraction. Although gulls sometimes utilized such rocks as fishing platforms, more frequently such platforms were used for loafing and other non-foraging activities. In addition, it was not feasible to distinguish foraging gulls standing on rocks from those loafing.

The gull chosen to be observed for foraging rate was the first individual observed consuming a fish within the study area. Once a gull was chosen it was followed continuously until a second successful capture occurred or a maximum of 30 minutes had passed. Initial successful feeding attempts were those in which a foraging bird captured a fish by plunging from the air into the water. Second takes were counted regardless of the means of capture. This accounted for the rare instance in which the second successful take by a gull was accomplished by stealing from another bird or jumping from an exposed rock or $\log$ into the water to catch a fish.

## River Reach Surveys- Spring and Summer

Spring river surveys included six river reaches. Each reach was surveyed approximately once every 2 weeks, from April 15 through June 28. These reaches included Benton, Vangie, Zillah, the Canyon, Cle Elum and Easton. During the summer, river surveys included only the Canyon, Cle Elum and

Easton reaches, which were surveyed every week from July 1 through August 28. The Canyon was an additional drift in the summer in 2002, compared with previous years, when only Cle Elum and Easton were surveyed during this time of year. All reaches surveyed in both the spring and summer were identical in length and location to those conducted in previous years.

All river reach surveys were conducted by a two-person survey team from either a 5.2 m aluminum drift boat or a two-person raft, depending upon water conditions. Most surveys began between 0800 and 0900 and lasted between 2.5 to 5.5 hours, depending upon length of reach, water flow and wind speed. All surveys were preformed while actively rowing the drift boat or raft down stream to decrease the interval of time required to traverse the reach.

Of the two-person survey team, one person was responsible for navigation while the other was responsible for identifying and recording birds. Team members alternated between rowing and bird identification duties approximately every hour. All piscivorous birds detected visually or aurally were recorded, including time of observation, species, and sex and age if they were distinguishable. Binoculars (Leica, 10x42) were used to aid in identification. All birds positively identified by the navigator were included, although the team member responsible for bird identification at the time of the encounter made final decisions for uncertain or potential repeat identifications, that is, double counting.

All piscivorous birds encountered on the river by survey personnel were recorded at the point of initial observation. Most birds observed were only slightly disturbed by the presence of the survey boat and were quickly passed. Navigation of the survey boat to the opposite side of the river away from encountered birds minimized escape behaviors. If subsequent to the encounter the bird attempted to escape from the survey boat by moving down river a note was made that the bird was being pushed. Birds being pushed were usually kept in sight until passed by the survey boat. Passage usually occurred when the river widened sufficiently to let the pushed bird pass to the side of the survey boat.

If the bird being pushed down river moved out of sight of the survey personnel, a note was made, and the next bird of the same species/ age/ sex to be encountered within the next 1000 meters of river was assumed to be the pushed bird. If a bird of the same species/ age/ sex was not encountered in the subsequent 1000 meters, the bird was assumed to have departed the river or
passed the survey boat without detection, and the next identification of a bird of the same species/ age/ sex was recorded as a new observation.

## Acclimation Site Surveys- Spring

Beginning February 1 and continuing until May 29, YN hatchery personnel at the Clark Flat, Jack Creek and Easton acclimation sites conducted piscivorous bird surveys. Jack Creek was surveyed from February 22 to May 23, Easton from March 1 to May 17, and Clark Flat from February 1 to May 29. In addition, a few observations were made at the Cle Elum Hatchery site from February 13 to April 3. Surveys were conducted at various times throughout the day. In general, each site had at least three surveys conducted, one in the morning, one around noon, and one later in the afternoon. All piscivorous birds within the acclimation facility, along the length of the artificial acclimation stream, and 50 meters above and 150 meters below the acclimation stream outlet, into the main stem of the Yakima River or N. Fork Teanaway, were identified and recorded within their respective zones. Surveys were conducted on foot by hatchery personnel.

## North Fork Teanaway River Surveys- Spring and Summer

The survey reach included the river and its banks from the Jungle Creek/ North Fork Teanaway confluence down river past the Jack Creek acclimation site continuing downstream for approximately 3.5 km . One surveyor moved down from Jungle Creek, noting the presence of piscivorous birds. If navigation of the river-bank was not possible, the river was crossed and surveys were continued on the opposite bank. If it was not possible to cross the river, detours were taken away from the river-bank, down stream, and paths through the underbrush were located to enable periodic return to the river-bank. Once there, a visual search up and down the stream was conducted. All piscivorous birds detected visually were recorded including time of observation, species of bird, and sex and age if distinguishable. A pair of Leica (10x42) binoculars was utilized to aid in identification. This area was surveyed nine times between May 2 and August 20, 2002, approximately once every two weeks.

## Secondary H otspot Surveys- Spring

Additional surveys were conducted in 2002 at four dam sites along the Yakima River. These surveys were conducted to ensure that potential hotspot sites were not being overlooked. These sites, in addition to others, were initially identified by Phinney et al. (1998) as areas for potential heavy predation and were also surveyed in 2000, but not in 2001. Sites surveyed in 2002 included Prosser Dam, Sunnyside Dam, Wapato Dam and Roza Dam. Each site was visited approximately nine times, once every one to two weeks between April

16 and June 25. Wapato Dam was only visited seven times due to high water conditions, which made the road to one part of the dam inaccessible. Observations were made for one hour at each site, with birds present noted every 15 minutes. Bird species, time, number and location, either above or below the dam, or at the canal intake at Prosser D am, were all noted.

In addition, checks were made at Prosser Dam when time permitted, to determine if there were a significant number of birds feeding at the head of the canal, where fish are susceptible to predation due to upwelling.

Progress: Avian predation of fish is suspected to contribute to the loss of outmigrating juvenile salmonids in the Yakima River Basin, potentially constraining natural and artificial production. In 1997 and 1998, the Yakima/ Klickitat Fisheries Project (YKFP), whose goal is to increase the natural production of salmonids within the Yakima River, initiated investigations to assess the feasibility of developing an index to avian predation of juvenile salmon within the river. This research, conducted by Dr. Steve Mathews and David Phinney of the University of Washington and the Washington Department of Fish and Wildlife (WDFW), confirmed that Ringbilled Gulls and Common Mergansers were the primary avian predators of juvenile salmon on the Yakima River (Phinney et al. 1998), and that under certain conditions could significantly impact migrating smolt populations.

Beginning in 1999, the Washington Cooperative Fish and Wildlife Research Unit (WACFWRU) was asked by the Y KFP to continue development of avian consumption indices. Monitoring methods developed by Phinney et al. (1998) were adopted with modifications and the monitoring of impacts to juvenile salmon along river reaches and at areas of high predator/ prey concentrations, referred to here as "hotspots", has continued each year through 2002. Beginning in 2002, the YKFP Yakama Nation (YN) personnel joined the monitoring of avian predation, working cooperatively with the WACFWRU.

In 2002, as in previous years, piscivorous birds were counted from river banks at hotspots and from a raft or drift boat along river reaches. Consumption by gulls at hotspots was based on direct observations of foraging success and modeled abundance while consumption by all other piscivorous birds was estimated using published dietary requirements and modeled abundance. Seasonal patterns of avian piscivore abundance were identified, diurnal patterns of gull abundance at hotspots were identified, and predation indices were calculated for hotspots and river reaches, for both the spring and summer.

General survey methods used in 2002 were the same as those used in 2001. Changes to the survey schedule in 2002 included the addition of surveys on the Easton reach during the early spring, and in the Canyon during the summer. Methods for measuring gull feeding rates at hotspots were the same as those used in 2001.

Primary avian predators in 2002 were again gulls, California and Ring-billed, at hotspots and Common Mergansers within the upper river reaches. Consumption on the lower reaches was distributed among a number of species. As in 2001, slightly more then half of all fish consumption in the lower reaches can be attributed to American White Pelicans. Estimated consumption by gulls at both hotspots combined, between April 11 to June 30, was 279,482 fish. Assuming a worst case scenario, that all fish taken were smolts, this represented approximately $10 \%$ of all smolts estimated passing or being released from the Prosser Dam area during the 2002 smolt migration season. Total gull abundances and estimates of consumption at the two hotspot sites showed an increase from that seen in 2001.

Total estimated take by Common Mergansers across all strata surveyed was $11,938 \mathrm{~kg}$ between April 8 and August 31, a decrease of $2,839 \mathrm{~kg}$ from 2001. Approximately 64 percent of that consumption was within the upper river reaches, where there is a known breeding population of mergansers.

## Task 4.b Fish Predation Index (Yakama Nation Portion Only)

Rationale: Develop an index of the mortality rate of upper Yakima spring chinook attributable to non-salmonid piscivorous fish in the lower Yakima. This index will be used to estimate the contribution of in-basin predation to fluctuations in hatchery and wild smolt-to-adult survival rate.

Methods: Monthly mark-recapture pikeminnow population estimates are attempted from March through June at Toppenish (RM 94-100), Sunnyside Dam (RM 103.2-103.8) and Granger (RM 80-83). In addition, stomach samples are collected from pikeminnows $200+\mathrm{cm}$ in fork length, which are collected primarily above and below the population estimate sites. Pikeminnow stomachs with fish present are further analyzed to determine what species and how many were consumed. This analysis is performed using diagnostic bones which allows determination of species (though for salmonids this is more difficult) and approximate body length. All new pikeminnows over $200+\mathrm{cm}$ are tagged with a PIT tag and subsequently all fish are scanned for the presence
of a PIT tag. If a PIT tag is found its code is recorded along with the fish's location (GPS) and its fork length recorded. An estimate of total salmonids consumed by the pikeminnow population on an annual basis is attempted based on the population estimates and the salmonid consumption rate measured from the pikeminnow stomach samples. The lack of valid population estimates over the years and across sites and months has made this last task difficult to achieve with precision.

Progress: Summarized in Table 15 are the population estimates for the Toppenish, Granger and Sunnyside Dam sample sites since 1999 when the project was initiated. In 2002 successful population estimates were made at Toppenish for April and May; at Granger for April, and no successful population estimates were made at Sunnyside Dam. Typically the lack of valid population estimates was a function of insufficient recaptures to validate the estimate.

Table 15. Summary of pikeminnow population estimates for the Toppenish, Granger and Sunnyside Dam sites, spring 2002.

| Year | Toppenish |  |  | Granger |  |  | Sunnyside Dam |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | April | May | June | April | May | June | April | May | June |
| 1999 | 933 | 1722 | 1220 | nv | 476 | nv | nv | 83 | nv |
| 2000 | nv | 2622 | 4811 | nv | nv | nv | nv | nv | nv |
| 2001 | 511 | 2167 | 1420 | nv | 568 | 828 | nv | nv | nv |
| 2002 | 2266 | 1432 | nv | 1627 | nv | 1149 | nv | nv | nv |

A summary of hatchery spring chinook and coho identified in the 2002 pikeminnow stomachs is presented in Table 16. Fish were identified as to their origin (when possible) from recovered CWT and PIT tags and colored elastomer fragments. A total of 59 fish were identified- 5 hatchery coho and 54 hatchery spring chinook. Of the 54 hatchery spring chinook smolts 21 were OCT, 20 SNT and 13 were unidentifiable. Of the four hatchery coho smolts, three were from the Naches releases and one from the upper Y akima releases.

A summary of pikeminnow stomachs collected at Toppenish, Sunnyside D am and Granger is presented in Table 17. A total of 805 stomachs (Toppenish452, Sunnyside Dam- 16 and Granger- 326) were collected during the spring 2002 field season. The mean percent of stomachs collected in March, April, May and June that contained fish at the Toppenish, Sunnyside Dam and Granger sites was $47 \%$ ( $25 \%-65 \%$ ), $59 \%$ ( $43 \%-78 \%$ ) and $36 \% ~(23 \%-$ $38 \%$ ), respectively. This represents the initial analysis. All stomachs with fish
present will be further analyzed to determine the species using diagnostic bones to identify them, which will be reported on in the FY 2003 annual report.

Within the sampling period from March through June of 2002 the pikeminnow population displayed fidelity within the reach they were initially marked. A total of 78 (Granger- 22, Toppenish- 49 and Sunnyside Dam- 7 ) pikeminnows were tagged and subsequently recaptured during the course of the spring sampling period. Of those fish tagged in the G ranger reach, three fish were found in the Toppenish reach and later had returned to the upper end of the G ranger reach. Within the Toppenish reach six fish out of 49 were re-sampled outside the reach on at least one occasion. Of these six occurrences, five fish were subsequently captured within the Toppenish reach by the end of the season. Fish were found moving both up and down stream out of their "home" reach. In the Sunnyside Dam reach one fish was sampled in the Toppenish reach and then later was found in its original reach.

## Task 4.c Indirect Predation (and environmental analysis)

Rationale: The release of hatchery salmonids may enhance or decrease the survival of randomly commingled wild salmonid smolts by altering the functional or numerical response of predators. For example, predators may increase consumption of wild fish by switching prey preferences from invertebrates to fish, or may be attracted to areas where hatchery fish are released. Conversely, large numbers of hatchery fish may confuse or satiate predators, resulting in enhanced survival of wild fish.

Methods: Survival from Prosser Dam to McNary Dam was estimated for separate releases of PIT-tagged spring chinook made in 2002 (coho and fall chinook releases were not analyzed because McNary detection rates had not yet been developed at the time of the analysis). All releases were "self-selected": made up of tagged fish released at various points above Prosser Dam and detected at the main PIT-tag detector at the Chandler trap over a one to twoday period. Fish detected at the secondary Chandler detector were excluded from analysis because the detector is located at the exit of the live-box, and fish detected at this point might have incurred stresses or injuries attributable solely to handling.

Survival was estimated from the main detector at Chandler trap at Prosser D am to McNary Dam on the Columbia River. The method of estimating survival consisted of dividing daily McNary tag detections by the estimated McNary
detection rate for the appropriate time period. McNary detection rates were estimated by Dr. Doug Neeley, and were based on the ratio of joint John D ayl McNary detections to John-D ay-only detections on a given day:

Detection rate (day i) = (number joint detections McNary and John D ay)/ (number detections at John D ay).

Dr. Neeley developed statistical techniques to determine appropriate intervals over the outmigration season during which it is most reasonable to use a mean detection rate as the interval-specific estimate.

Multiple logistic regression was used to detect a survival impact attributable to a number of factors acting both just below Prosser D am and in the McNary fore bay. The variables that were examined were: flow (below Prosser and in McNary fore bay); water temperature (Prosser and McNary); and smolt density (daily passage estimate at Prosser and Smolt Passage Index at McNary). Unlike analyses in earlier years, turbidity could not be included in this analysis because the turbidity detector at Prosser Dam malfunctioned. Similarly, the mean size of smolts in the self-selected releases could not be used because none of the fish used in these releases were subsampled.

This procedure assumes a number of factors affect smolt survival and that if there is a real indirect predation effect on survival, it should be statistically apparent after the effects of the other factors have been accounted for. Accordingly, a statistical test of developed by Dr. D oug Neeley, a Y N biometrical consultant, was developed which determines the significance of one of two independently significant independent variables when both are acting simultaneously. The only fadtors considered to ex ert a real effect on survival were those whose impad remained significant after the affect of other (independent significant) factors had been aocounted for (by Dr. N eeley's analysis).

Again, the 2002 analysis included only spring chinook and only "unhandled" spring chinook- those detected only at the main PIT-tag detector at Chandler.

Progress: There is evidence from 2002 outmigrants that survival to McNary Dam of hatchery-produced spring chinook smolt increases with an increase in the number of fish volitionally exiting acclimation ponds. Even though survival for these 2002 outmigrants also increased with increased release-site stream flow, the relation of survival to fish number appears to be independent of stream flow's effect.

Table 16. Summary of hatchery coho and spring chinook found in pikeminnows in 2002.

| Reach | Species | CWT code | Pit tag \# | Acclimation release site | Rearing treatment | Total \# salmon in stomach |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Granger | hat spck | 63-12-96 |  | Jack Cr | OCT | 1 |
| Granger | hat spck | 63-12-96 |  | Jack Cr | OCT | 1 |
| Granger | hat spck | 63-12-96 |  | Jack Cr | OCT | 1 |
| Granger | hat spck | 63-05-83 |  | Clark Flat | OCT | 1 |
| Granger | hat spck | 63-09-74 |  | Easton | OCT | 1 |
| Granger | hat spck | 63-09-79 |  | Easton | OCT | 1 |
| Granger | hat spck | 63-13-63 |  | Jack Cr | OCT | 1 |
| Granger | hat spck | 63-12-99 |  | Easton | OCT | 1 |
| Sunnyside Dam | hat spck | 63-09-79 |  | Easton | OCT | 1 |
| Sunnyside Dam | hat spck | 63-09-79 |  | Easton | OCT | 1 |
| Sunnyside Dam | hat spck | 63-09-79 |  | Easton | OCT | 1 |
| Toppenish | hat spck | 63-09-80 |  | Clark Flat | OCT | 1 |
| Toppenish | hat spck | 63-09-72 |  | Jack Cr | OCT | 1 |
| Toppenish | hat spck | 63-13-65 |  | Clark Flat | OCT | 1 |
| Toppenish | hat spck | 63-13-63 |  | Jack Cr | OCT | 1 |
| Toppenish | hat spck | 63-13-63 |  | Jack Cr | OCT | 1 |
| Toppenish | hat spck | 63-05-83 |  | Clark Flat | OCT | 1 |
| Toppenish | hat spck | 63-13-65 |  | Clark Flat | OCT | 1 |
| Toppenish | hat spck | 63-13-65 |  | Clark Flat | OCT | 1 |
| Toppenish | hat spck | 63-12-99 |  | Easton | OCT | 1 |
| Toppenish | hat spck | 63-12-99 |  | Easton | OCT | 1 |
| Granger | hat spck |  | 3D9.1BF1302576 |  | SNT | 1 |
| Granger | hat spck | 63-13-64 |  | Clark Flat | SNT | 1 |
| Granger | hat spck | 63-12-98 |  | Easton | SNT | 1 |
| Granger | hat spck | 63-12-98 |  | Easton | SNT | 1 |
| Granger | hat spck | 63-11-76 |  | Easton | SNT | 1 |
| Sunnyside Dam | hat spck | 63-13-60 |  | Jack Cr | SNT | 1 |
| Sunnyside Dam | hat spck | 63-13-60 |  | Jack Cr | SNT | 1 |
| Sunnyside Dam | hat spck | 63-13-60 |  | Jack Cr | SNT | 1 |
| Toppenish | hat spck | 63-13-60 |  | Jack Cr | SNT | 1 |
| Toppenish | hat spck | 63-11-76 |  | Easton | SNT | 1 |
| Toppenish | hat spck | 63-11-76 |  | Easton | SNT | 1 |
| Toppenish | hat spck | 63-12-97 |  | Jack Cr | SNT | 1 |
| Toppenish | hat spck | 63-13-60 |  | Jack Cr | SNT | 1 |
| Toppenish | hat spck | 63-13-64 |  | Clark Flat | SNT | 1 |
| Toppenish | hat spck | 63-13-64 |  | Clark Flat | SNT | 1 |
| Toppenish | hat spck | 63-09-78 |  | Easton | SNT | 1 |
| Toppenish | hat spck | 63-12-97 |  | Jack Cr | SNT | 1 |
| Toppenish | hat spck | 63-09-81 |  | Clark Flat | SNT | 1 |
| Toppenish | hat spck | 63-09-81 |  | Clark Flat | SNT | 1 |
| Toppenish | hat spck | 63-09-81 |  | Clark Flat | SNT | 1 |
| Granger | hat spck |  |  |  |  | 1 |
| Sunnyside Dam | hat spck |  |  |  |  | 1 |
| Sunnyside Dam | hat spck |  |  |  |  | 1 |
| Sunnyside Dam | hat spck |  |  |  |  | 1 |
| Toppenish | hat spck |  |  |  |  | 1 |
| Toppenish | hat spck |  |  |  |  | 1 |
| Toppenish | hat coho | 5-43-13 |  | lost Cr Yakima Early |  | 1 |
| Toppenish | hat spck |  |  |  |  | 1 |
| Toppenish | hat spck |  |  |  |  | 1 |
| Toppenish | hat spck |  |  |  |  | 1 |
| Toppenish | hat spck |  |  |  |  | 1 |
| Toppenish | hat spck |  |  |  |  | 1 |
| Toppenish | hat spck |  |  |  |  | 1 |
| Toppenish | hat spck |  | 3D9.1BF12F4938 |  |  | 1 |
| Toppenish | hat coho | 5-43-15 |  | Stiles Willard Early |  | 1 |
| Toppenish | hat coho | 5-44-45 |  | Stiles Willard Early |  | 1 |
| Toppenish | hat coho | 5-43-11 |  | Easton Willard Late |  | 2 |
| Total |  |  |  |  |  | 59 |

Table 17. Summary of pikeminnow stomach samples for the Toppenish, Sunnyside Dam and Granger sample sites, spring 2002.

|  | Toppenish |  |  |  |  | Sunnyside |  |  |  |  | Granger |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date | No. of fish sacrificed | No. of empty stomaches | No. stomachs with biomass | No. stomachs with fish | \% stomachs with fish | No. of fish sacrificed | No. of empty stomaches | No. stomachs with biomass | No. stomachs with fish | \% stomachs with fish | No. of fish sacrificed | No. of empty stomaches | No. stomachs with biomass | No. stomachs with fish | \% stomachs with fish |
| 3/21 | 17 | 9 | 8 | 7 |  | 5 | 52 | 3 | 3 |  |  |  |  |  |  |
| 3/22 |  |  |  |  |  |  |  |  |  |  | 11 | 6 | 5 | 4 |  |
| 3/28 | 66 | 30 | 36 | 27 |  | 6 | $6 \quad 1$ | 5 | 3 |  |  |  |  |  |  |
| 3/29 |  |  |  |  |  |  |  |  |  |  | 16 | 6 | 10 | 4 |  |
| Monthly Total | 83 | 39 | 44 | 34 | 41\% | 11 | - 3 | 8 | 6 | 55\% | 27 | 12 | 15 | 8 | 30\% |
| 4/4 | 36 | 6 | 30 | 25 |  | 3 | 32 | 1 | 1 |  |  |  |  |  |  |
| 4/5 | 41 | 15 | 26 | 16 |  | 3 | 3 | 1 | 1 |  |  |  |  |  |  |
| 4/11 |  |  |  |  |  |  |  |  |  |  | 58 | 21 | 37 | 16 |  |
| 4/12 |  |  |  |  |  |  |  |  |  |  | 42 | 10 | 32 | 15 |  |
| 4/18 | 42 | 2 | 40 | 38 |  | 1 | 10 | 1 | 1 |  | 26 | 5 | 21 | 17 |  |
| 4/19 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 4/22 | 22 | 3 | 18 | 13 |  |  |  |  |  |  |  |  |  |  |  |
| Monthly Total | 141 | 26 | 114 | 92 | 65\% | 7 | 7 | 3 | 3 | 43\% | 126 | 36 | 90 | 48 | 38\% |
| 5/2 | 40 | 6 | 34 | 16 |  | 7 | 70 | 7 | 5 |  |  |  |  |  |  |
| 5/3 | 35 | 8 | 37 | 16 |  | 2 | 20 | 2 | 2 |  |  |  |  |  |  |
| 5/9 |  |  |  |  |  |  |  |  |  |  | 30 | 12 | 18 | 6 |  |
| 5/13 |  |  |  |  |  |  |  |  |  |  | 37 | 10 | 27 | 8 |  |
| 5/14 | 25 | 3 | 22 | 14 |  |  |  |  |  |  | 10 | 2 | 8 | 3 |  |
| 5/20 | 25 | 5 | 20 | 9 |  |  |  |  |  |  |  |  |  |  |  |
| 5/21 | 23 | 11 | 12 | 10 |  |  |  |  |  |  |  |  |  |  |  |
| 5/22 |  |  |  |  |  |  |  |  |  |  | 22 | 8 | 14 | 6 |  |
| Monthly Total | 148 | 33 | 125 | 65 | 44\% | 9 | 0 | 9 | 9 7 | 78\% | 99 | 32 | 67 | 23 | 23\% |
| 6/6 | 24 | 9 | 15 | 4 |  | 0 | 00 | 0 | 0 |  |  |  |  |  |  |
| 6/7 | 15 | 5 | 10 | 2 |  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |
| 6/13 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 6/14 |  |  |  |  |  |  |  |  |  |  | 33 | 5 | 28 | 11 |  |
| 6/17 | 25 | 13 | 12 | 7 |  |  |  |  |  |  | 10 | 1 | 9 | 3 |  |
| 6/18 | 16 | 7 | 9 | 7 |  |  |  |  |  |  | 31 | 6 | 25 | 7 |  |
| Monthly Total | 80 | 34 | 46 | 20 | 25\% | 0 | 0 | 0 | 0 | ---- | 74 | 12 | 62 | 21 | 28\% |
| Seasonal Totals | 452 | 132 | 329 | 211 | 47\% | 27 | 7 | 20 | -16 | 59\% | 326 | 92 | 234 | 100 | 31\% |

There was no evidence of a change in survival with a change in volitional release number for 2000 and 2001 outmigrants.
A complete read of this study is presented in Appendix D.

## Task 4.d Yakima River Spring Chinook Competition/ Prey Index

This task is assigned to WDFW and they will report on its status in their annual progress report to BPA.

Task 4.e Upper Yakima Spring Chinook NTTOC Monitoring
The WDFW annual report for this task can be located on the BPA website: http:// www.efw.bpa.gov/ cgi-bin/ FW/ publications.cgi

Pearsons, Todd, Brenda James, Christopher Johnson, Anthony Fritts, G abriel Temple. 2003. Spring chinook salmon interactions indices and residual/ precocial monitoring in the upper Y akima Basin, Y akima/ Klickitat Fisheries Project, Annual Report 2002. Project No. 1995-06424, BPA Report D OE/ BP-00004666-14.

## Task 4.f Pathogen Sampling

The WDFW annual report for this task can be located on the BPA website: http:/ / www.efw.bpa.gov/ cgi-bin/ FW/ publications.cgi

Pearsons, Todd, Joan Thomas. 2003. Pathogen screening of naturally produced Yakima River spring chinook smolts, Annual Report 2001. Project No. 1995-06424, BPA Report D OE/ BP-00004666-8.

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## APPENDICES

## A through $\mathbf{H}$

A. Task 1.a. Modeling (CWU subcontract)
B. Task 1.e. Roza juvenile wild/ hatchery spring chinook smolt PIT tagging
C. Task 1.f. Yakima River wild/ hatchery salmonid survival and enumeration (Chandler)
D. Task 1.g. Yakima River fall chinook M\&E
E. Task 1.i. Yakima Spring Chinook Juvenile Behavior
F. Task 4.c. Indirect Predation (and environmental analysis)
G. M\&E Financial Report
H.

M\&E Equipment Inventory List

## Appendix A

Protocols to Measure and Assess Select Geomorphic and Habitat Correlates For the YKFP EDT Model (draft report)

# Protocols to Measure and Assess Select Geomorphic and Habitat Correlates 

 For the YKFP EDT ModelDraft Report

Central Washington University Department of Geography and Land Studies

August 2003

The purpose of this work is to develop a biophysical classification scheme and measurement protocols that will increase the precision of environmental data for the Yakima Klickitat Fisheries Project (YKFP) Ecosystem Diagnosis and Treatment (EDT) model. The first goal focuses on establishing a classification scheme that systematically defines and identifies ecologically homogenous stream reaches within a river basin. The second goal is the development of protocols used to measure select level two correlates within the EDT model; they include the following: gradient, natural and anthropogenic confinement, minimum and maximum channel width, habitat type composition, riparian function, and measurements of woody debris. The third goal applies the results of protocols for the level two correlates described above to the EDT reaches found within the Easton and Cle Elum floodplains of the Yakima River basin. Utilizing data generated from the previous three goals, the fourth goal is the initial development of a preservation-restoration scheme for the Easton floodplain.

All of the methods described below were developed or chosen for their level of precision relative to assessment scale and the expenditure of both time and money needed to implement them. This is a key point, since the range of index values associated with each of the level two correlates within EDT does not necessitate absolute precision. By making this statement, we are not implying that inherent inaccuracies exist within the EDT model itself, but are merely discounting many of the criticisms that have surfaced surrounding the amount of time and money needed to populate an EDT database. This is not to say that generating data for all level two correlates in a chosen river basin will necessarily be without difficulty; nevertheless, we do believe that the following protocols will serve to expedite the EDT process wherever it is implemented.

## Using Stream Gradient and Confinement to Derive a Geomorphic Channel Classification Scheme

An inherent and crucial step in the development of regional watershed classification schemes is the systematic definition and identification of ecologically homogenous stream reaches within a river basin. Classification permits stream reaches to be identified and inventoried within an objective, quantifiable, hierarchical, and communicable framework (Kondolf 1995). Many geomorphically-based classification systems have been developed over the last 100 years (see reviews by Bauer and Ralph 1999; Hawkes 1975; Kondolf 1995; Montgomery and Buffington 1996; Mosley 1987; Naiman et al. 1992), each as a varied as the stream morphologies they try to represent (Montgomery and Buffington 1996). General classifications of stream channels have been developed based on stream order (Horton 1945; Strahler 1957), relationships between slope and discharge (e.g. Leopold and Wolman 1957), modes of sediment transport (e.g. Schumm 1977), and longitudinal zonation (e.g. Palmer 1976). Classifications have become increasingly more descriptive and complex, emphasizing differences in channel patterns based on additional factors such as landform setting and degree of confinement (e.g. Galay et al. 1973), sediment supply and channel stability (e.g. Kellerhaus et al. 1976; Church 1992), island and bar types (e.g. Galay et al. 1973; Kellerhaus et al. 1976; Church 1992), valley stability and characteristics (e.g. Galay et al. 1973; Cupp 1989), as well as floodplain energy and sediment characteristics (e.g. Nanson and Croke 1992). As an extreme example of complexity, a channel reach classification developed by Rosgen (1994) includes 7 major and 42 minor channel types based on variables including channel pattern, entrenchment, width-to-depth ratios, sinuosity, gradient, and bed material size.

Each of these channel classification has advantages and disadvantages for biophysical, engineering and ecological applications (Kondolf 1995), while no classification can address all possible channel types (Montgomery and Buffington 1996). In addition, most of these classifications are largely descriptive characterizations of channel patterns, and are not processbased. One notable exception is Whiting and Bradley's (1993) process-based classification for headwater channels that links patterns based on factors such as gradient, channel widths and depth, valley-channel width ratios, and substrate size to potentials for debris flow impacts and sediment transport rate and processes.

We have chosen to use a geomorphic, process-based classification system developed within the Pacific Northwest by Montgomery and Buffington to classify the EDT reaches of the Yakima River Basin. Based on classifying bedforms, it is a comprehensive classification scheme used regularly by agencies such as the United States Forest Service (USFS) (Arend 1999). Lacking the complexity of Rosgen's (1994) hierarchical classification system, the MontgomeryBuffington approach integrates well with channel geomorphic unit classifications (e.g. Hawkins et al. 1993), thereby providing a useful tool for classifying aquatic habitats at intermediate landscape scales (Bisson and Montgomery 1996). Being process-based also allows for better analysis of the relationships between geomorphic and habitat correlates/variables used within the EDT model.

The Montgomery-Buffington classification approach focuses on the physical relationships between three internal forcing mechanisms, including variations in transport capacity, sediment supply, and large woody debris (LWD) within a stream reach. The interrelated processes between these variables ultimately determine a channel's morphology. It is important to note that any alteration to one or more of these three mechanisms, whether natural or anthropogenic, can elicit a concomitant response in channel form. For instance, fluxes in discharge, sediment supply, and riparian vegetation that result from external mechanisms such as climate change, masswasting, dam and reservoir construction, confinement, and clearing of riparian vegetation will be reflected by alterations to a channel's width, depth, slope, grain size, bedform, and pattern (Montgomery and Buffington 1998). The dynamic relationship between internal and external forcing mechanisms determines, at least in part, the condition of the aquatic habitat present within a watershed (Montgomery et al. 1999).

A channel's morphology within the Montgomery-Buffington classification scheme is initially based on information from topographic maps and aerial photographs, though site visits are necessary to verify reach boundaries and their classifications (Arend 1999). The first step in classifying a channel reach within this system is to derive slope gradient. Once generated, these data allow the initial empirical association of gradient to channel form. Frequency distributions of surveyed gradients for Pacific Northwest rivers west of the Cascade crest within the state of Washington and their associated forms are presented in Table 1. Bedrock, and forced alluvial reaches (forced pool-riffle and forced step-pool) can occur across the various range of slope
gradients listed above; however, bedrock and forced step-pool reaches are commonly found on slopes with higher gradients, while forced poolriffle reaches will more often be found on lower gradient slopes. Because each channel form is more or less sensitive to the external influences that effect one or more of the three internal forcing mechanisms, we should expect discontinuities between slope gradient measurements and the present channel form predicted by valley slope alone. For example, cascade, bedrock, and step-pool channels show little response to perturbations. Conversely, colluvial and plane-bed channels exhibit a moderate response, while pool-riffle and dune-ripple channels are most sensitive to fluxes (Montgomery and MacDonald 2002).

Table 1. Slope gradient and associated channel forms (Montgomery and Buffington, 1998).

|  | Slope <br> Gradient | $\mathbf{< 1 \%}$ | $\mathbf{2 \% - 4 \%}$ | $\mathbf{4 \% - 8 \%}$ | $\mathbf{8 \% - 2 0 \%}$ | $\mathbf{> 2 0 \%}$ |
| :---: | :--- | :--- | :--- | :--- | :--- | :---: |
| Channel <br> Form |  | Dune-Ripple <br> \& Pool-Riffle | Plane-Bed | Step-Pool | Cascade | Colluvial |

Since the effects of confinement can significantly transform a channel's morphology, determining the degree of confinement, whether natural or anthropogenic, is the second step in classifying a channel and its potential response to perturbations within this system. Together, valley slope and the degree of channel confinement more accurately describe the channel form of a given reach (Montgomery and MacDonald 2002). Field verification of slope gradient measurements for various channels within the Yakima River basin shows that a low gradient pool-riffle reach can exhibit characteristics of a plane-bed channel form and is most likely the result of an increase in transport capacity, sediment supply, or both (Figures 1-3). Slope gradient and the degree of channel confinement for large order streams ( $4^{\text {th }}$ order and up) are easily derived using contemporary geo-spatial software; nevertheless, field verification of channel form remains an essential step in assessing the accuracy of measured results. There is simply no substitute for on-the-ground observations, which become essential when working on smaller order streams ( $3^{\text {rd }}$ to $1^{\text {st }}$ order). For example, field verification of slope gradient measurements that projected a plane-bed channel revealed a forced step-pool morphology may result from woody debris inputs (Figure 4), while slope gradient measurements that projected a pool-riffle channel revealed a plane-bed morphology in areas of increased confinement.

## Gradient

## EDT Definition:

The average gradient of the main channel of the reach over its entire length.
Note: Categorical levels are shown here but values are required to be input as point estimates for each reach.

EDT Categories:

| Index 0 | Index 1 | Index 2 | Index 3 | Index 4 |
| :---: | :---: | :---: | :---: | :---: |
| $0-0.1 \%$ | $>0.10 \%$ and $<0.5 \%$ | $>0.5 \%$ and $<1 \%$ | $>1 \%$ and $<2 \%$ | $>2 \%$ and $<4 \%$ |

## Measurement Techniques

Gradient measurements may be conducted via field surveys, a combination of digital elevation models (DEMs) and Geographic Information System (GIS) hydrology layers, or topographic map interpretation using either a map wheel, piece of string or software such as MAPTECH Terrain Navigator. In all cases, the percent gradient is then calculated by dividing the change in elevation (i.e. rise) by the distance between the points (i.e. run), and multiplying by 100 . Gradient should be recorded to the nearest $0.1 \%$ (Johnston and Slaney 1996; Overton et al. 1997).

## Field survey techniques

Slopes may be measured on the ground using a stadia rod and either line of sight or level line methods (Murdoch et al. 2001), a hand held clinometer (Hogan et al. 1996), or a hand level and surveying rod (Fitzpatrick et al. 1998). Overton et al. (1997) caution against using clinometers due to the high variability in results when applied to stream gradients, while Fitzpatrick et al (1998) suggest that accurate slope measurements on low-gradient streams requires using a surveyor's level on a tripod and a surveyor's rod. Bauer and Ralph (1999) contend that the most accurate method of measuring gradient is by creating a longitudinal profile of an entire stream reach using surveying equipment, though this may not be cost-effective for most EDT modeling exercises.

In terms of measurement spacing, Hogan et al. (1996) propose calculating average channel gradient from five evenly spaced measurements along a reach, each taken over similar distances, generally over the longest length of channel visible between field surveyors, with a minimum
length of several channel widths. Harrelson et al. (1994) argue for longitudinal profile lengths of approximately 20 times the bankfull channel width, while Overton et al. (1997) suggest measurement distances of 200-300 m, taken along relatively straight sections of river at least 2030 m in length, and between similar morphological features (e.g. from one riffle crest to the next). Measurements may also be taken between transects used to calculate channel width (Fitzpatrick et al. 1997).

Measurements are taken at the water's edge or surface, relative to semi-permanent markers with either known or assumed elevations, and are corrected for the height of the measur ing instrument (see Fitzpatrick et al. 1998, Harrelson et al. 1994, Murdoch et al. 2001, or Overton et al. 1997 for more detail). The distance should also be taken between measurement points using a tape measure or line held taut between the points. Note: Arend and Bain (1999) make a further distinction between measuring the "energy gradient" (i.e. the surface of the stream), which is generally assumed to be synonymous with stream gradient by other authors, and their definition of stream gradient (measured along the thalweg).

## Topographic map interpretation

Using large scale maps (scales 1:24,000 or greater), the main channel length of larger streams may be obtained by subtracting the river mile estimate for the upstream boundary of the reach from the river mile estimate for the downstream reach, while the elevations for the two boundary points may be estimated from the contour lines (USFS 2001). A more exact estimate of stream channel length (i.e. the blue line distance) may be measured by tracing the main channel using either a map wheel or string, and converting the distance by the scale of the map (Allen and Guenther 1996; Murdock et al. 2001; Overton et al. 1997; Watershed Professional Network 1999). One can also calculate slope using a gradient template printed on a clear piece of Mylar (Pleus and Schuett-Hames 1998), though the latter should only be used for relatively straight channels (Watershed Professional Network 1999). Software such as MAPTECH Terrain Navigator may also be used to measure gradient from topographic maps.

## GIS tools

If available, gradients may be calculated using a combination of digital elevation models and a GIS hydrology layer (Allen and Guenther 1996; Johnston and Slaney 1996), using the GIS
system to calculate channel gradients between individual contours or "smooth" them by a running average (Watershed Professional Network 1999).

## Comparison of Methods

We compared the precision of three different methods for measuring the average gradient of a reach over its entire length. The three methods assessed include using ArcGIS Geographic Information System (GIS) software, MAPTECH Terrain Navigator software, and a handheld map planimeter. The advantages and disadvantages for each method are presented in Table 2.

Table 2. Advantages and disadvantages for methods used to determine slope gradient.

|  | Method | GIS | MAPTECH <br> Terrain Navigator | 1:24000 USGS <br> Map \& Map <br> Planimeter |
| :--- | :--- | :--- | :--- | :--- |
| Advantages/Disadvantages |  | High Precision | High Precision | Moderate Precision |
|  |  | Retrievable Spatial <br> Dig ital Data | Retrievable Spatial <br> Digital Data | No Digital Data <br> Generated |
|  |  | Data easily merged <br> with EDT Database | Data entry into <br> EDT Database <br> required | Data entry into <br> EDT Database <br> required |
|  |  | Expensive | Inexpensive | Inexpensive |

Results of the two-tailed Wilcoxon signed-rank test for slope gradient measurements on $\sim 490$ river miles within the Yakima River basin were not significant at the .05 confidence level when comparing ArcGIS and MAPTECH Terrain Navigator software. Likewise, comparing results of the two-tailed Wilcoxon signed-rank test for slope gradient on 45 river miles using a map planimeter with United States Geological Survey (USGS) 1:24000 Quadrangles, ArcGIS, and MAPTECH Terrain Navigator were also not significant at the .05 confidence level.

Given that there is no significant difference in the level of precision for measuring slope gradient by any of the three methods tested, the decision to use one method over another becomes more subjective. If for example, a GIS is already in place and the skill le vel of the technician operating the software is advanced, then generating slope gradient measurements for an entire river basin is practical. However, if an up and running GIS does not exist, or a skilled operator is lacking, then we highly recommend implementing one of the two other methods explored here. Given its advantages, we recommend the use of MAPTECH Terrain Navigator software. Not only is Terrain Navigator somewhat more precise, but it also generates retrievable digital data in less
time. Furthermore, a cost comparison shows that Terrain Navigator software is by far less expensive than purchasing the large quantities of USGS 1:24000 Quadrangles required for use with a map planimeter over an entire drainage basin.

Results for slope gradient measurements for all EDT reaches within the Yakima River basin are presented in appendices A-1 through A-3. Only those cells colored green have been field verified. In those cases where two channel forms are listed, the first form predominates within the reach and the second occurs where the channel is influenced by one or more of the internal and/or external variables mentioned above. Given these results, we believe the MontgomeryBuffington classification system, while initially developed for use in forested watersheds on the west side of the Cascades, may be readily used to classify reaches in eastern Washington.

## Confinement - Natural

## EDT Definition:

The extent that the valley floodplain of the reach is confined by natural features -determined as the ratio between the width of the valley floodplain and the bankfull channel width.

Note: this attribute addresses the natural (pristine) state of valley confinement only. The extent that reaches are confined by hydromodifications (e.g., diking) is addressed under a separate attribute.

## EDT Categories

| Index 0 | Index 1 | Index 2 | Index 3 | Index 4 |
| :---: | :---: | :---: | :---: | :---: |
| Reach mostly unconfined by natural features Average valley width > 4 channel widths. | Reach comprised approximately equally of unconfined and moderately confined sections. | Reach mostly moderately confined by natural features -Average valley width 2-4 channel widths. | Reach comprised approximately equally of moderately confined and unconfined sections. | Reach mostly confined by natural features Average valley width < 2 channel widths. |

## Confinement - Hydromodifications

## EDT Definition:

The extent that man-made structures within or adjacent to the stream channel constrict flow (as at bridges) or restrict flow access to the stream's floodplain (due to streamside roads, revetments, diking or levees) or the extent that the channel has been ditched or channelized, or has undergone significant streambed degradation caused by channel incision/entrenchment (associated with the process called "headcutting"). Flow access to the floodplain can be partially or wholly cut off due to channel incision.

Note: Setback levees are to be treated differently than narrow-channel or riverfront levees-consider the extent of the setback and its effect on flow and bed dynamics and micro-habitat features along the stream margin in reach to arrive at rating conclusion. Reference condition for this attribute is the natural, undeveloped state.

EDT Categories

| Index 0 | Index 1 | Index 2 | Index 3 | Index 4 |
| :---: | :---: | :---: | :---: | :---: |
| The stream channel within the reach is essentially fully connected to its floodplain. Very minor structures may exist in the reach that do not result in flow constriction or restriction. Note: this describes both a natural condition within a naturally unconfined channel as well as the natural condition within a canyon. | Some portion of the stream channel, though less than $10 \%$ (of the sum of lengths of both banks), is disconnected from its floodplain along one or both banks due to man-made structures or ditching. | More than $10 \%$ and less than $40 \%$ of the entire length of the stream channel (sum of lengths of both banks) within the reach is disconnected from its floodplain along one or both banks due to man-made structures or ditching. | More than $40 \%$ and less than $80 \%$ of the entire length of the stream channel (sum of lengths of both banks) within the reach is disconnected from its floodplain along one or both banks due to man-made structures or ditching. | Greater than $80 \%$ of the entire length of the stream channel (sum of lengths of both banks) within the reach is disconnected from its floodplain along one or both banks due to man-made structures or ditching. |

## Measurement Techniques

While many authors agree that natural confinement may be calculated as a ratio of bankfull width to its floodplain (e.g. Bauer and Ralph 1999; Overton et al. 1997), little guidance is given in the literature on how to best calculate this correlate. Bauer and Ralph (1999) suggest calculating confinement as a ratio between bankfull width (often correlated with the 1.5 year recurrence interval flood) and either the 100 year floodplain or the channel migration zone. The Watershed Professionals Network (1999) use a similar ratio, though they define the modern floodplain as the flood-prone area, which may not correspond to the 100-year floodplain. For example, USFS (2001) defines the flood-prone are as the width of the valley floor inundated during the 50 -year flood, which may be estimated by doubling the maximum bankfull depth and extending the resulting flood-prone elevation across the floodplain. Moore et al. (2002), who define confinement as a ratio of active channel width to valley width, similarly distinguish confinement using the flood-prone elevation, defining "constrained" valleys as those with terrace heights greater than flood-prone elevations. Both floodplain and bankfull channel widths can be estimated using topographic maps and aerial photographs, though measurements should be verified in the field along evenly spaced intervals along the longitudinal profile of the reach (Pleus and Schuett-Hames 1998; Watershed Professionals Network 1999). All width measurements should be measured perpendicular to their feature's corresponding centerlines (Pleus and Schuett-Hames 1998).

We believe that the degree of natural and anthropogenic confinement may be derived via a simple two-step process. Using aerial photography with Mylar overlays, or enlarged copies of USGS 1:24000 quadrangles, the investigator performs a field reconnaissance of the entire reach. All natural and cultural features that confine the channel from its 100-year floodplain are drawn onto the aerial photographs or map copies (Fitzpatrick et al. 1998; Pleus and Schuett-Hames 1998). Once complete, this data is then entered into either a GIS or MAPTECH's Terrain Navigator. Again, the investigator's choice should be based on practicability, since either of the two technologies will return the same level of precision. Figures 5-6 depict examples of both natural and anthropogenic confinement along portions of the Easton reach.

Definitions for determining the degree of both natural and anthropogenic confinement as described within the EDT model should be closely followed. According to the definition for confinement induced by hydro-modifications, anthropogenic confinement is measured along that portion of the reach where cultural features are present either on one or both banks of the channel. Once all measurements are made, the length of the confined channel is then divided by the total length of the entire reach. Multiplying the resulting quotient by 100 produces the percentage of the reach that is confined. This percentage determines the index value for confinement within the EDT model.

With the exception of differing index values, deriving the degree of natural confinement is similar to the method described above. Using a combination of topographic maps, Federal Emergency Management Agency (FEMA) floodplain maps, and aerial photographs, measure the width of the 100-year floodplain at systematic intervals along a reach (e.g. 500 m ), and compare it to 2 and 4 times the average bankfull channel width for the reach. For larger streams lacking 100-year floodplain information, the flood-prone area may be estimated from aerial photographs using the extent of riparian vegetation as a proxy indicator.

The degree of confinement for smaller streams must be measured along systematically spaced intervals in the field, using the 50-year floodplain estimated by doubling the maximum bankfull depth and extending the resulting flood-prone elevation across the floodplain (USFS 2001). To determine flood-prone width, pieces of flagging are temporarily tied to vegetation corresponding
to the flood-prone elevation, and a measuring tape is stretched level at that elevation to determine the extent of the flood-prone elevation. Flood-prone widths are measured to the nearest foot if less than 4 times the bankfull width at that transect.

Results for measures of confinement within the Cle Eum and Easton reaches show the effects of anthropogenic features on channel form within these two alluvial floodplains. For instance, of the total $\sim 10.03$ river miles with the Cle Elum reach, 3.4 river miles, or $\sim 33.86$ percent of its length is confined by human-induced features, yet a much greater proportion of the channel's length seems to be affected. As already mentioned, gradient measurements projected that a pool riffle channel form should predominate throughout this reach; however, habitat unit measurements show that nearly 83 percent of the channel exhibits a plane-bed morphology. Similarly, anthropogenic features confine 4.95 river miles, or $\sim 26.29$ percent of the total $\sim 18.82$ river miles within the Easton reach; an additional 2.68 percent is naturally confined. Even though habitat unit measurements have not been completed for the entire reach at this time, preliminary surveys show that the majority of the reach exhibits a plane-bed channel morphology rather than a pool-riffle form.

## Channel Widths

## EDT Definitions:

## Month Maximum Width:

Average width of the wetted channel during high flow month (average monthly conditions). If the stream is braided or contains multiple channels, then the width would represent the sum of the wetted widths along a transect that extends across all channels. Channel width -month maximum width ( ft ) is to be rated for the month when average flow tends to be highest. This month will typically be during some part of March-June east of the Cascade crest and during December or January on the west side of the crest.

## Month Minimum Width:

Average width of the wetted channel during low flow month (average monthly conditions). If the stream is braided or contains multiple channels, then the width would represent the sum of the wetted widths along a transect that extends across all channels.

## EDT Categories

| Index 0 | Index 1 | Index 2 | Index 3 | Index 4 |
| :---: | :---: | :---: | :---: | :---: |
| $<15 \mathrm{ft}$ | $>15 \mathrm{ft}$ and $<60 \mathrm{ft}$ | $>60 \mathrm{ft}$ and $<100 \mathrm{ft}$ | $>100 \mathrm{ft}$ and 360 ft | $>360 \mathrm{ft}$ |

Note: categorical index levels are presented because they are used in some bio-rules. However, it is now required that Level 2 attribute values for this attribute be input as non-categorical estimates, i.e., point estimates.

## Measurement Techniques

Measuring maximum and minimum channel widths was conducted using a combination of aerial photo interpretation, field survey, and estimation methods.

## Aerial photo interpretation

For the aerial photo interpretation, we compared three spatial sampling approaches commonly used to sample streams and rivers: random, systematic, and stratified (Conquest and Ralph 1998) (Figure 7).

Simple random samples were generated by dividing each EDT reach into 100 m segments, numbering each segment, and choosing a representative sample of 30 using a random number
table (see Note A). This method allowed channel measurements to be randomly distributed, though the samples tended to be more time consuming to generate. In addition, a random sample may not provide uniform, representative coverage with limited numbers of samples. Systematic samples were also taken by dividing each EDT reach into 30 regular sampling intervals, providing uniform coverage of each reach. The systematic sampling method is easier than the random method to implement, and generally allows for a representative sample as long as there is no underlying periodicity corresponding with the spatial sampling intervals. Finally, we established stratified samples for each reach by classifying the reaches into homogeneous subsets based on a combination of: 1) number of channels (single or multiple); and 2) degree of confinement (unconfined, naturally confined, or $100 \%$ anthropogenically confined) (Table 3; Figures 8-9).

Table 3. Stratified sample categories and proportions for the Easton and Cle Elum EDT reaches.

| Category | Cle Elum | Easton |
| :--- | :---: | :---: |
| Unconfined single channel | $68.8 \%$ | $53.8 \%$ |
| Confined single channel | $13.8 \%$ | $8.2 \%$ |
| Unconfined multiple channel | $17.4 \%$ | $36.3 \%$ |
| Confined multiple channel | $0 \%$ | $1.7 \%$ |

The stratified approach allowed smaller samples to be taken while still ensuring smaller subsets were represented in the sample. As all three EDT reaches in the Easton reach were classified as one geomorphic unit using Montgomery and Buffington's classification scheme, we combined all three reaches into a single sampling unit, making sure to generate 10 random and systematic sampling points for each EDT. This process of "lumping" stream segments (Pleus and SchuettHames 1998) for the purpose of sampling is justifiable for EDT reaches that have homogenous geomorphic categories, and no substantial differences in flow due to the confluence of significant tributaries.

Each stratified sample consisted of 4-5 channel measurements (Hogan et al. 1996), which were spaced approximately 5-7 channel widths apart. Several authors suggest determining sample stream segments based on 20 average channel widths, with the smallest possible segment being $300 \mathrm{ft}(100 \mathrm{~m})$ (Fitzpatrick et al. 1998; MacDonald et al. 1991; Pleus and Schuett-Hames 1998). The 20-channel width criteria attempts to encompass at least one complete meander wavelength, based on the classic pool-riffle channel system proposed by Leopold et al. (1964), with pools
spaced every 5-7 channel widths. Having a minimum of 3-4 repeated habitat association patterns ensures that all habitat types are represented in the stream segment (Fitzpatrick et al. 1998), and allows for more effective statistical analysis and confidence (Pleus and Schuett-Hames 1998). Fitzpatrick et al. (1998) further contend that a minimum sample reach length is necessary for representative samples, while a maximum length is needed to prevent a reduction in sampling efficiency. They suggest that the minimum and maximum sample reach lengths for wadeable streams are 150 and 300 m , respectively, while the recommended minimum and maximum lengths for nonwadeable streams are 500 and 100 m , respectively. Fitzpatrick et al. (1998) also suggest that each stream segment be further divided into 10 equal parts for sampling, and that habitat units such as channel width be measured at 11 equal points (i.e. approximately 2 channel widths apart). However, given the similarity in channel measurements we found at each location, such sampling intensity might not be warranted in most cases.

Wetted channel widths were measured using ArcGIS and rectified aerial photos taken during flow conditions approximating month-maximum and month-minimum flows. The width of gravel bars and vegetated islands were excluded from the overall channel width, while multiple channels were summed for a total width measurement. Measurements were recorded to the nearest 1.0 ft , given the degree of precision required by EDT model. The measurements were averaged for each subset category, and applied proportionally by the percentage of the reach with a similar classification to obtain an overall average channel width for the EDT reach.

## Field survey

For the field assessment, we used the stratified sampling approach to obtain channel width measurements, as obstructions to navigation and limited public access prohibited application of random and systematic sampling in the field. Public access locations representing each of the stratified sample categories were located using a combination of USGS topographic maps, aerial photographs, and field reconnaissance. Each stratified sample consisted of 4-5 regularly spaced channel measurements (Hogan et al. 1996). Arend and Bain (1999) suggest that transects should be spaced approximately 5-7 channel widths and up to 40 channel widths apart, depending on research objectives. Schuett-Hames et al. (1999) recommend adjusting transect intervals to stream segment lengths, with transect intervals varying from $10 \%$ of segment lengths for segments less than 100 m , to 100 m intervals for stream segments over 2500 m .

Measurements were taken at representative locations (Johnston and Slaney 1996; Overton et al. 1997), usually in riffle areas (USFS 2001) or other straight sections with no signs of water stacking or piling (such as the outside of a bend or near channel obstructions) (Allen and Guenther 1996; Harrelson et al. 1994; USFS 2001). In addition, transects were located at locations with clear bankfull indicators (Allen and Guenther 1996; Harrelson et al. 1994; USFS 2001). Areas with undercut banks or actively eroding banks are to be avoided, since bank slumping tends to obscure true bankfull conditions (USFS 2001).

Both the minimum and maximum channel widths were measured at the same transect locations during the low flow month, based on average monthly conditions. Using a Bushnell Yardage Pro 400 laser range finder, the minimum channel width was estimated by measuring the wetted width from one side of the stream to the other, perpendicular to the flow or channel axis (i.e. thalweg). The wetted edge was defined as the point where sediment particles are no longer surrounded by water (Johnston and Slaney 1996). Cross-channel measurements excluded any dry channel bars (Johnston and Slaney 1996; Fitzpatrick et al. 1998), and multiple channel widths were summed for a total width. Channel widths less than 17 feet (the lower threshold of the laser range finder) were measured using a measuring tape stretched tight from one wetted edge to the other. Measurements were recorded to the nearest 1.0 ft , given the degree of precision required by EDT model. Maximum channel widths were measured from the top of one stream bank to the other, again perpendicular to stream flow. The height/extent of bankfull flow was estimated using a variety of indicators widely used in the literature, including: 1) change in bank morphology (e.g. slope changes, top of point bar deposits and undercut banks); 2) change in sediment composition (e.g. sand to pebbles); 3) vegetative indicators (e.g. beginning of perennial terrestrial vegetation, lower limit of lichens and mosses; 4) scour lines (e.g. exposed roots,); and 5) defined water marks (e.g. stain lines, line of organic debris on the ground)(Allen and Guenther 1996; Arend and Bain 1999; Fitzpatrick, et al. 1998; Harrelson et al. 1994; Hogan et al. 1996; Johnston and Slaney 1996; Moore et al. 2002; Pleus and Schuett-Hames 1998; USFS 2001). The width of vegetated islands with perennial terrestrial vegetation > 1 m in height were not included in the overall width measurement, and multiple channel widths were summed for a total width (Hogan et al. 1996; Johnston and Slaney 1996; Moore et al. 2002).

Again, the stratified approach allowed smaller samples to be taken while ensuring that smaller subsets were still represented in the sample. Each stratified sample consisted of 4-5 channel measurements, which were spaced approximately 5-7 channel widths apart. The measurements were averaged for each subset category, and applied proportionally by the percentage of the reach with a similar classification to obtain an overall average channel width for the EDT reach.

## Estimation techniques

The EDT primer suggests that if empirical width data are not available for the reach of interest, reasonable conclusions can usually be based on personal knowledge of the area. In some cases, a better characterization of flow may exist than channel width. Here, an estimate of width (in feet) for larger streams might be obtained from flow data (cfs) using an equation formulated for streams on the east side of the Cascade crest using an equation given in Johnson et al. (1988) as follows:

$$
\text { Width }=a * C F S^{b}
$$

Where $a=4.5789$ and $b=0.5660$

## Comparison of Methods:

## Aerial photo interpretation vs. Field survey

In comparing 30 replicate channel width measurements, no significant difference was found between the GIS and field based measuring methods for either minimum or maximum flow widths (two sample t -test, $\mathrm{p}<0.05_{\text {two-tailed }}$ (Tables 4-5).

Table 4. Comparison of maximum channel width measurements.

|  | Random | Systematic | Stratified GIS <br> Measures | Stratified <br> Field <br> Measures | Estimation <br> Technique |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Easton |  |  |  |  |  |
| Mean (ft.) | 123 | 110 | 103 | 115 | 181 |
| SE Mean (ft.) | 5.9 | 5.8 | 3.0 | 5.0 |  |
| EDT Index Value | 3 | 3 | 3 | 3 | 3 |
| Cle Elum |  |  |  |  |  |
| Mean (ft.) | 185 | 185 | 232 | 247 | 463 |
| SE Mean(ft.) | 6.6 | 6.1 | 8.9 | 13 |  |
| EDT Index Value | 3 | 3 | 3 | 3 | 4 |

Table 5. Comparison of minimum channel width measurements.

|  | Random | Systematic | Stratified GIS <br> Measures | Stratified <br> Field <br> Measures | Estimation <br> Technique |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Easton |  |  |  |  |  |
| Mean (ft.) | 89 | 93 | 93 | 93 | 94 |
| SE Mean (ft.) | 3.5 | 5.0 | 4.9 | 3.9 |  |
| EDT Index Value | 2 | 2 | 2 | 2 | 2 |
| Cle Elum |  |  |  |  |  |
| Mean (ft.) | 129 | 142 | 134 | 159 | 158 |
| SE Mean (ft.) | 4.9 | 7.5 | 7.8 | 13 |  |
| EDT Index Value | 3 | 3 | 3 | 3 | 3 |

## Sample designs

No significant difference was found between the minimum channel widths determined through the various sampling methods (random, systematic, stratified) (KruskalWallis test, $\mathrm{p}<0.05_{\text {two- }}$ tailed) (Table 5). In addition, no significant difference was found between the maximum channel widths determined through the random and systematic sampling methods (two sample $t$-test, $\mathrm{p}<0.05_{\text {two-tailed }}$ ) (Table 4). However, we did find a significant difference between the maximum channel widths determined through the various sampling methods for the Cle Elum reach (Kruskal-Wallis test, $\mathrm{p}<0.05_{\text {two-tailed) , though no significant difference was found for the Easton }}$ reach (Table 4). The estimation technique using the equation derived by Johnson et al. (1988) seems to be consistent with the minimum channel width measurements derived through the other methods. However, the formula greatly overestimated the maximum channel widths, illustrating the problem of using a generalized equation on a regulated river with sustained high flows and a relatively high degree of confinement.

## Habitat Type

Using Hawkins et al. (1993), the EDT model distinguishes between three different categories of habitat types: 1) slow water (e.g. primary pools, pool-tailouts/glides, beaver ponds, and backwater pools); 2) fast water habitat types (e.g. small cobble/gravel riffles and large cobble/boulder riffles); and 3) off-channel habitat.

## EDT Definitions:

## Slow Water Habitat Types

## Backwater pools:

Percentage of the wetted channel surface area comprising backwater pools. Backwater pools are habitat units located along the channel margins but are otherwise enclosed-though still connected to the main channel (or side channel).
Note: backwater pools as defined here include "alcoves" as described by Nickleson et al. (1992).

## Beaver ponds:

Percentage of the wetted channel surface area comprising beaver ponds.
Note: this includes only those sites associated with the main channel or its side channels. Offchannel sites are addressed through the Off-Channel Habitat Factor.

## Pool tailouts:

Percentage of the wetted channel surface area comprising pool tailouts.

## Glides:

Percentage of the wetted channel surface area comprising glides.
Note: There is a general lack of consensus regarding the definition of glides (Hawkins et al. 1993), despite a a commonly held view that it remains important to recognize a habitat type that is intermediate between pool and riffle. The definition applied here is from the ODFW habitat survey manual (Moore et al. 1997): an area with generally uniform depth and flow with no surface turbulence, generally in reaches of $<1 \%$ gradient. Glides may have some small scour areas but are distinguished from pools by their overall homogeneity and lack of structure. They are generally deeper than riffles with few major flow obstructions and low habitat complexity.

## Primary pools:

Percentage of the wetted channel surface area comprising pools, excluding beaver ponds

## Fast Water Habitat Types

## Large cobble/boulder riffles:

Percentage of the wetted channel surface area comprising large cobble/boulder riffles. Particle sizes of substrate modified from Platts et al. (1983) based on information in Gordon et al. (1991): gravel ( 0.2 to 2.9 inch diameter), small cobble ( 2.9 to 5 inch diameter), large cobble ( 5 to 11.9 inch diameter), boulder (>11.9 inch diameter).

## Small cobble/gravel riffles:

Percentage of the wetted channel surface area comprising small cobble/gravel riffles. Particle sizes of substrate modified from Platts et al. (1983) based on information in Gordon et a. (1991): gravel ( 0.2 to 2.9 inch diameter), small cobble ( 2.9 to 5 inch diameter), large cobble ( 5 to 11.9 inch diameter), boulder (>11.9 inch diameter).

## EDT Categories

| Index 0 | Index 1 | Index 2 | Index 3 | Index 4 |
| :--- | :--- | :--- | :--- | :--- |
| $0-<0.25 \%$ of wetted <br> surface area <br> encompasses this <br> habitat type | $>0.25 \%$ and $<5 \%$ of <br> wetted surface area <br> encompasses this <br> habitat type | $>5 \%$ and $<25 \%$ of <br> wetted surface area <br> encompasses this <br> habitat type | $>25 \%$ and $<50 \%$ of <br> wetted surface area <br> encompasses this <br> habitat type | $>50 \%$ of wetted <br> surface area <br> encompasses this <br> habitat type |

Note: Where an index value is associated with a range, the integer value is assumed for modeling to be the midpoint. Index values can be identified as non-integers to represent the lower or upper ends of a range.

## Off-Channel Habitat Factor

## EDT Definition

A multiplier used to estimate the amount of off-channel habitat based on the wetted surface area of the all combined in-channel habitat. Off-channel habitats consist of oxbows, backswamps, riverine ponds, and the channels that connect them to the main channel or its side channels.

## EDT Categories

| Index 0 | Index 1 | Index 2 | Index 3 | Index 4 |
| :---: | :---: | :---: | :---: | :---: |
| No off-channel <br> habitat present | $>0 \mathrm{X}$ and $<0.05 \mathrm{X}$ | $>0.05 \mathrm{X}$ and $<0.25 \mathrm{X}$ | $>0.25 \mathrm{X}$ and $<0.5 \mathrm{X}$ | $>0.5 \mathrm{X}$ |

Note: Where an index value is associated with a range, the integer value is assumed for modeling to be the midpoint. Index values can be identified as non-integers to represent the lower or upper ends of a range.

## Identifying Habitat Types

The EDT habitat types are based on a variety of hierarchical classification schemes focused primarily on water velocity, channel morphology, turbulence, substrate characteristics, and water depth (e.g. Armantrout 1996; Flosi and Reynolds 1994; Hawkins et al. 1993). These classification frameworks are more complex, expanding on the primary habitat units used by the EDT model. Several authors provide good descriptions, cross-sectional diagrams, and/or ground photos of each classification type (e.g. Arend 1999; Fitzpatrick et al. 1998; Johnston and Slaney 1996; Moore et al. 2002; Overton et al. 1997; Pleus et al. 1999; USFS 2001; Watershed Professionals Network 1999).

Habitat types are distinguished by fluvial hydraulic and geomorphic descriptors, including water speed and depth, surface turbulence, substrate characteristics, bed roughness and uniformity, as well as bed and water surface slopes (Moore et al. 2002; Overton et al. 1997; Pleus et al. 1999). Boundaries between these discrete channel units are based on identifying changes in stream channel slopes along the thalweg of the channel bottom, such as the riffle crest (i.e. the high point in channel bed below a pool) (Overton et al. 1997; Pleus et al. 1999).

Pools are geomorphic channel units where water is impounded within a scour depression associated with a channel obstruction (i.e. hydraulic control). These features are characterized by reduced velocity, little surface turbulence (with the exception of eddies), and deeper water. In sharp contrast, riffles are relatively shallow, occur in straight stretches of the river, and have relatively fast flows over completely or partially submerged obstructions, leading to surface turbulence. Riffles also have coarser substrates. Glides (i.e. runs), typically found in the transition zone between pool tail-outs and riffles and in low-gradient reaches with no flow obstructions, have moderate depth, moderate to high flows, and no apparent surface turbulence. Glide cross-sections are U-shaped, with uniform substrates.

## Measurement Techniques

## Field survey

Several authors (e.g. Arend 1999; Pleus et al. 1999) suggest that habitat surveys be conducted during moderate to low flow conditions, preferably during the late summer/early fall when
discharge conditions are generally the most stable and allow repeat surveys to be conducted at similar discharges. While repeat surveys can be conducted during higher flows, Pleus et al. (1999) contend that such flows generally increase data variability because of decreased visibility and access due to increased turbidity, turbulence and water depths. Moderate or most frequent flows may also aid proper identification of the habitat types most commonly found in a stream segment throughout the year, as both higher and lower lows can change habitat classifications and sizes (Roper and Scarnecchia 1995). For example, a riffle might resemble a glide during high flow conditions, while a glide may become a riffle during low flow conditions (Fitzpatrick et al. 1998). In addition to the influence of stream discharge, Roper and Scarnecchia (1995) have noted variability in classifying habitat types can be related to differences in: 1) the level of distinction required in classification (e.g. pools in general vs. several specific subtypes of pools); 2) the level and uniformity of observer training; as well as 3) differences in other stream characteristics (e.g. gradient).

Bisson and Montgomery (1996) state that habitat unit inventories of small to mid-size streams are often time consuming, typically requiring teams of 2-3 people to cover $1-5 \mathrm{~km}$ per day. In addition, other factors such as reach length, available time, and access may make the study of an entire reach impractical. They suggest studying representative sections of a reach, providing that the sections include examples of each type of habitat unit present in the whole reach. As obstructions to navigation and limited public access prohibited measuring all the habitat units in the reach, we established stratified samples for each reach by classifying the reaches into homogeneous subsets based on a combination of: 1) number of channels (single or multiple); and 2) degree of confinement (unconfined, natural confined, or $100 \%$ anthropogenically confined) (Table 1). The stratified approach allowed smaller samples to be taken while ensuring smaller subsets were still represented in the sample.

Each stratified sample consisted of stream segments representing each reach category. Several authors suggest determining sample stream segments based on 20-50 average channel widths, with the smallest possible segment being $300 \mathrm{ft}(100 \mathrm{~m})$ (Bisson and Montgomery 1996;
Fitzpatrick et al.1998; MacDonald et al. 1991; Pleus and Schuett-Hames 1998). The 20-channel width criteria attempts to encompass at least one complete meander wavelength, based on the classic pool-riffle channel system proposed by Leopold et al. (1964), with pools spaced every 5-7
channel widths. Having a minimum of 3-4 repeated habitat association patterns ensures that all habitat types are represented in the stream segment (Fitzpatrick et al. 1998), and allows for more effective statistical analysis and confidence (Pleus and Schuett-Hames 1998).

Fitzpatrick et al. (1998) maintain that a minimum sample reach length is necessary for representative samples, while a maximum length is needed to prevent a reduction in sampling efficiency. They suggest that the minimum and maximum sample reach lengths for wadeable streams are 150 and 300 m , respectively, while the recommended minimum and maximum lengths for nonwadeable streams are 500 and 100 m , respectively. In addition, each stream segment should include at least two examples of each type of habitat unit, and only habitat units greater than $50 \%$ of the channel width should be measured and recorded. Similarly, Pleus et al. (1999) require that riffle and pool habitat units meet certain minimum surface size criteria based on the stream segment's mean bankfull width (Table 2). In order to be considered a habitat unit, its length has to be equal to or greater than the wetted width. The USFS (2001) requires that the sampling frequency must ensure that at least 10 pools and 10 riffles as well as $10 \%$ of all pools and riffles are measured for each stream.

Habitat type surveys typically begin at the downstream portion of a stream segment, moving systematically upstream (Bisson and Montgomery 1996; Fitzpatrick et al. 1998; Moore et al. 2002; Overton et al. 1997; Pleus et al. 1999; USFS 2001). Using a combination of field notes and a mylar sheet superimposed over a large-scale aerial photograph, each habitat unit was given a unique unit number, increasing sequentially upstream. Where multiple channels were present, geomorphic habitat units were first assessed and numbered in the main channel from the downstream outlet of the secondary channel to the upstream inlet, returning to the downstream portion of the secondary channel and continuing the sequential numbering of habitat units until the main channel was reached. Geomorphic habitat units were further labeled according to the EDT habitat types, distinguishing primary pools (PP), pool-tailouts (PT), glides (GL), beaver ponds (BV), backwater pools (BP), small cobble/gravel riffles (SCR) and large cobble/boulder riffles (LCR).

The two riffle habitat types are further distinguished on the relative size of cobbles. Dominant cobble sizes can be estimated using the Wolman Pebble count (Wolman 1954), or an ocular
estimate method (Overton 1997). The Wolman Pebble count method collects pebbles and cobbles along a transect, moving one step at a time from one stream bank at the bankfull elevation to the other. At each step, you pick up the first cobble one finger length from the toe of your boot, and measure its intermediate axis. To reduce sampling bias further, you must look across the channel rather than down at its bed. The transect is generally traversed several times to measure the recommended 100 pebbles/cobbles, though 25-50 may be enough in some instances. The ocular estimate method simply estimates the proportion of each sediment size class along an entire riffle, again from the bankfull elevation on each stream bank.

The length of each pool, riffle, and glide was measured along its thalweg using a Bushnell Yardage Pro 400 laser rangefinder, though a measuring tape, hip chain, or pacing technique could also have been used (Murdoch et al. 2001). Bisson and Montgomery (1996) suggest that rangefinders be calibrated at the beginning of each field trip by measuring the distance between two points with a tape and adjusting the readings on the rangefinder, which can become misaligned if dropped. GPS units have also been used for some reach surveys, though problems can occur in areas with heavy forest canopies or high topographic relief (Bisson and Montgomery 1996). Habitat units less than 17 feet in length or width (the lower threshold of the laser range finder) were measured using a measuring tape. Measurements were recorded to the nearest 1.0 ft , given the degree of precision required by EDT model. The maximum thalweg distance was recorded for each habitat unit, as well as one representative width measurement at the habitat unit's midpoint (Arend and Bain 1999). For sinuous habitats, the length was measured as the sum of straight line lengths along the thalweg (Johnston and Slaney 1996). Several authors suggest that habitat unit widths should be measured and averaged at a minimum of three points, especially when working within irregularly shaped habitat (Allen and Guenther 1996; Johnston and Slaney 1996; Overton et al. 1997), located one-quarter, one-half, and three-quarters of the habitat's length (Dolloff et al. 1997). Longer fast-water habitat units may also require several width measurements taken along the habitat unit (Overton et al. 1997).

During the stream inventory, the extent of off-channel habitat was observed and located on the aerial photographs to be measured later using ArcGIS. Where possible, the widths of side channels were measured entering and leaving the stream channel, and then averaged.

Area values for each habitat type (length x width) were summed to determine total length for the stream segment (Murdoch et al. 2001). The measurements were averaged for each subset category, and applied proportionally by the percentage of the reach with a similar classification to obtain overall average habitat unit proportions for the EDT reach.

As an alternative, Macdonald, et al. (1991) and Johnson and Slaney (1996) suggest visually estimating the area of each habitat unit for a reach based on the method developed by Hankin and Reeves (1988), as well as accurately measuring a systematic sample of each habitat type to develop calibration ratios. Moore et al. (2002) similarly propose estimating the size of each habitat unit, verifying every tenth unit with accurate measurements, while Dolloff et al. (1997) suggest verification of $20 \%$ of pools and $10 \%$ of riffles and cascades. Visual assessment methods and verification procedures are best applied when one has access and wishes to rapidly sample an entire EDT reach, either by foot or by boat. However, these methods are less appropriate when sampling stratified subsets, and then applying the less precise results to the overall reach.

Dolloff et al. (1997) have compared estimates of stream habitat using basin-wide visual estimation techniques versus extrapolating habitat information from 3-4 "representative reaches" of approximately 100 m in length. They found that the representative reach extrapolation technique tended to overestimate numbers of pools, while underestimating the number of cascades and the average area of all habitat types. However, representative reaches were chosen on the basis of professional judgment of whether they represented a stream or watershed as a whole, rather than stratifying the reaches further on the basis of differences in channel type, gradient and confinement, and only applying the results at a reach scale.

## Aerial photography

While aerial photographs may be used to classify stream types (Mollard 1973), identify channel disturbance (Grant 1988), and determine the size and shape of riparian areas (Platts et al. 1987), identification of channel geomorphic units such as pools and riffles may be difficult for many streams (Arend 1999). Aerial photographs at $1: 12,000$ scale may be used to measure the widths of streams and riparian areas as well as the extent of large woody debris (Ham 1996), though timely photos at this scale as well as at appropriate flows (e.g. minimum, maximum, and most frequent) are difficult and expensive to obtain, especially for all the reaches in a watershed.

Channel characteristics on small to medium size streams may also be difficult to detect because of steep slopes, dense riparian cover, and shading effects (Bisson and Montgomery 1996; Grant 1988). Large-scale, color aerial photographs ranging from 1:1,000 to $1: 5,000$ are needed to interpret more detailed information on stream habitats (Johnston and Slaney 1996; Platts et al. 1987).

## Aerial videography

Videography is being increasingly used to map linear features such as coastlines and fluvial environments, including assessment of pool-riffle spacing, large woody debris, and riparian vegetation (Ham 1996). However, the nature of these studies has been largely reconnaissance based, as detailed mapping and interpretation is limited by the difficulty of tying the imagery to known ground coordinates. While data can generally be transferred onto a map with an accuracy of plus or minus 100 m , detailed inventories with accuracy of 3-10 m require differentially corrected GPS coordinates, laser altimeters, and onboard compensation for aircraft movement.

## Riparian Function

## EDT Definition

The correlate "riparian function" is defined by EDT as a measure of riparian function that has been altered within the reach.

## EDT Categories

| Index 0 | Index 1 | Index 2 | Index 3 | Index 4 |
| :---: | :---: | :---: | :---: | :---: |
| Strong linkages with no anthropogenic influences. | $>75-90 \%$ of functional attributes present (overbank <br> flows, vegetated streambanks, groundwater interactions typically present). | 50-75\% functional attribute ratingsignificant loss of riparian functioningminor channel incision, diminished riparian vegetation structure and inputs etc. | $25-50 \%$ similarity to natural conditions in functional attributesmany linkages between the stream and its floodplain are severed. | < $25 \%$ functional attribute rating: complete severing of floodplain-stream linkages |

The riparian zone adjacent to a stream channel is important for proper stream functioning (Naiman et al 1998). A stream is connected to the riparian zone by pathways of water, organisms, and resources that involve the stream channel, shallow groundwater aquifer, flood plain, and adjacent uplands (Ward and Stanford 1995).

Riparian function is evaluated in terms of overbank flows, vegetated banks, and groundwater interactions. The quickest and most effective method to evaluate these functional attributes is by describing riparian vegetation, and using riparian vegetation as a proxy for the other two attributes, since riparian vegetation influences many fluvial geomorphic processes (Hickin 1984). In addition, riparian vegetation provides several critical habitat functions, including: 1) filtering surface runoff and promoting nutrient uptake, 2) channel shading, 3) streambank stability, 4) spawning habitat and cover for fish, and input of litter and woody debris (Quinn et al. 2001). Vegetation in the floodplain is dependent on over-bank flows and/or groundwater interactions (Decamps 1996). Therefore, a lack of vegetation indicates a significant loss of riparian function.

## Measurement Techniques

## Aerial photo interpretation

Aerial photographs may be used to describe riparian vegetation on mainstem rivers. Once the photos are scanned and rectified, GIS software may be used to measure loss of riparian vegetation. A FEMA 100-year flood plain digital coverage for the reach of interest is laid over the corresponding aerial photography, as it is assumed that in an unaltered state, riparian vegetation would occur within the 100-year flood plain (Knutson and Naef, 1997). This assumption can be verified by selecting a location along the reach that remains unaffected by human modification and determining the percentage of the 100-year floodplain that has riparian vegetation occurring within it. This percentage can then be used as a benchmark from which to adjust further measurements. If historic aerial photographs are available, they may be used as a measure of quality assurance since they demarcate the extent of riparian vegetation prior to human disturbance.

Riparian vegetation should be measured at 500 m intervals along a reach in terms of what percentage of the 100-year floodplain is vegetated. This percentage is then converted into a corresponding correlate score $(0-4)$, as shown in Table 6 . If there is great variation in riparian vegetation, measurements should be taken at shorter ( 250 m ) intervals. This methodology is only

Table 6. Percent of riparian vegetation lost and corresponding correlates.

| \% Riparian Vegetation Lost | EDT Correlate |
| :---: | :---: |
| $<10$ | 0 |
| $10-25$ | 1 |
| $25-50$ | 2 |
| $50-75$ | 3 |
| $75+$ | 4 |

useful for large mainstem rivers with adequate riparian coverage. For smaller streams, tributaries, and areas without extensive riparian coverage, a field assessment of riparian function will be necessary.

## Field survey

Several field survey techniques of riparian function focus on bank stability. MacDonald (1991) suggests visual estimation techniques (i.e. Platts et al.1987) using a multi- parameter approach that assigns values to the following streambank parameters as presented in Table 7.

Table 7. Streambank parameters and corresponding values for determining riparian function (Platts, 1987).

| Channel location | Parameter | Range of values |
| :---: | :---: | :---: |
| Upper bank | Side slope gradient | $0-8$ |
|  | Mass wasting potential | $0-12$ |
|  | Debris jam potential | $0-8$ |
|  | Lower bank | Vegetative cover |
|  | Channel capacity | $0-12$ |
|  | Bank rock content | $0-4$ |
|  | Obstructions and flow | $0-8$ |
|  | Deflectors | $0-8$ |
|  | Bank cutting | $0-16$ |
|  | Sediment deposition | $0-16$ |

Allen and Guenther (1996) suggest, at a minimum, measuring lineal distance of actively eroding bank along sides of stream above wetted edge/bankfull channel. For a more representative characterization they suggest separate measurements along both the upper and lower bank. Active eroding banks are characterized by the presence of one or more of these factors: bare exposed colluvial or alluvial substrates, exposed mineral soil, or evidence of tension cracks. This method is also suggested by the United States Department of Agriculture (USDA 2001).

Other field assessments of riparian function use a combination of bank stability and vegetation measurements. Platts et al. (1987) measure streambank stability by classifying the percent of streambank covered by vegetation or by boulders or rubble, using intervals of 0-24\%, 25-49\%, $50-74 \%$, and $75-100 \%$. Streamsides are also rated on a scale of 1 to 5 based on vegetation present (5-shrubs, 4 -trees, 3 -grass, 2 -forbs dominant streamside vegetation, and $1->50 \%$ of streambank transect line has no vegetation present). Murdoch et al. (2001) suggests the following: 1) classify vegetation on stream banks as abundant, moderately sparse, or nonexistent; 2) estimate the percent of banks covered by vegetation; 3 ) evaluate bank stability, noting specific areas which are eroding or have collapsed; 4) note bank steepness and effects of anthropogenic change; 5) describe types of vegetation present, and estimate width of riparian
zone. Plafkin et al. (1989) similarly measure bank stability in terms of bank failure, slopes, eroded areas, and potential for future erosion. Their method measures bank vegetation stability in terms of percent of stream bank surface covered by vegetation of boulders and cobbles, and describes dominant type of streamside cover as shrubs, trees, grass/forbs, or none. These parameters are classified as excellent, good, fair, or poor.
Several methods quantify riparian function using average scores or index values based on a number of criteria. For example, Bain and Stevenson (1999) use rating criteria for vegetative cover, rocky cover, and total cover for transects with boundaries evenly spaced 5 to 7 average channel widths apart. They compute mean bank scores by multiplying rated values by the number of observations and dividing the sum of products by the number of transect segments. Similarly, Fizpatrick et al. (1998) use a bank stability index based on criteria presented in Table 8.

Table 8. Vegetative and bank characteristics and their corresponding scores (Bain and Stevenson, 1999).

| Characteristic | Measurement | Score |
| :---: | :---: | :---: |
| Angle of bank (degrees) | $0-30$ | 1 |
|  | $31-60$ | 2 |
|  | Vegetation cover (\%) | $>60$ |
| 3 |  |  |
|  | $>80$ | 1 |
|  | $50-80$ | 2 |
| Bank height (m) | $<20$ | 3 |
|  | $0-1$ | 1 |
|  | $1.1-2$ | 2 |
|  | $2.1-3$ | 3 |
|  | $3.1-4$ | 4 |
|  | Substrate |  |
| (category) | $>4$ | 5 |
|  | Bedrock, artificial | 1 |
|  | Boulder, cobble | 3 |
|  | Silt | 5 |
|  | Sand | 8 |
|  | Gravel/sand | 10 |


| Total score |  |
| :--- | :--- |
| 4-7: stable | 11-15: unstable |
| 8-10: at risk | 16-22: very unstable |

## Large Woody Debris (LWD)

## EDT Definition:

The correlate "Large Woody Debris" (LWD) is defined by EDT as a measure of the amount of wood within a reach. Dimensions of what constitutes LWD are defined here as pieces $>0.1 \mathrm{~m}$ diameter and $>2 \mathrm{~m}$ in length. Numbers and volumes of LWD corresponding to index levels are
based on Peterson et al. (1992), May et al. (1997), Hyatt and Naiman (2001), and Collins et al. (2002).

Note: channel widths here refer to average wetted width during the high flow month (< bank full), consistent with the metric used to define high flow channel width. Ranges for index values are based on LWD pieces/CW and presence of jams (on larger channels). Reference to "large" pieces in index values uses the standard TFW definition as those $>50 \mathrm{~cm}$ diameter at midpoint.

## EDT Categories:

| Index 0 | Index 1 | Index 2 | Index 3 | Index 4 |
| :---: | :---: | :---: | :---: | :---: |
| A complex mixture of single large pieces and accumulations consisting of all sizes, decay classes, and species origins; crosschannel jams are present where appropriate vegetation and channel conditions facilitate their existence; large wood pieces are a dominant influence on channel diversity (e.g., pools, gravel bars, and mid-channel islands) where channel gradient and flow allow such influences. Density of LWD (pieces per channel width CW) consistent with the following: channel width <25 ft -- 3-10 pieces/CW, 25-50 ft -- 3-10 pieces/CW, 50-150 ft -- 7-30 pieces/CW , 150$400 \mathrm{ft}--20-50$ pieces/CW in conjunction with large jams in areas where accumulations might occur, >400 ft -- 15-37 pieces/CW in conjunction with large jams in areas where accumulations might occur. | Complex array of large wood pieces but fewer cross channel bars and fewer pieces of sound large wood due to less recruitment than index level 1 ; influences of large wood and jams are a prevalent influence on channel morphology where channel gradient and flow allow such influences. Density of LWD (pieces per channel width CW) consistent with the following: channel width < 25 ft -- 2-3 pieces/CW, 25-50 ft --2-4 pieces/CW, 50-150 ft -- 3-7 pieces/CW , $150-400 \mathrm{ft}-10-20$ pieces/CW (excluding large jams) in conjunction with large jams in areas where accumulations might occur, >400 ft -- 8-15 pieces/CW (excluding large jams) in conjunction with large jams in areas where accumulations might occur. | Few pieces of large wood and their lengths are reduced and decay classes older due to less recruitment than in index level 1 ; small debris jams poorly anchored in place; large wood habitat and channel features of large wood origin are uncommon where channel gradient and flow allow such influences. Density of LWD (pieces per channel width CW) consistent with the following: channel width < $25 \mathrm{ft}-\mathrm{t}^{-}$2 pieces/CW, 25-50 ft --$1-2$ pieces/CW, 50-150 ft -- 1-3 pieces/CW , 150400 ft -- 10-20 pieces/CW without large jams in areas where accumulations might occur, $>400 \mathrm{ft}$-- 8-15 pieces/CW without large jams in areas where accumulations might occur. | Large pieces of wood rare and the natural function of wood pieces limited due to diminished quantities, sizes, decay classes and the capacity of the riparian streambank vegetation to retain pieces where channel gradient and flow allow such influences. Density of LWD (pieces per channel width CW) consistent with the following: channel width <25 ft --0.33-1 pieces/CW, 25$50 \mathrm{ft}-\mathrm{-}$ 0.33-1 pieces/CW, 50-150 ft --0.33-1 pieces/CW, 150-400 ft -- 3-10 pieces/CW without large jams in areas where accumulations might occur, $>400 \mathrm{ft}$--2-8 pieces/CW without large jams in areas where accumulations might occur. | Pieces of LWD rare. Density of LWD (pieces per channel width CW) consistent with the following: channel width <25 ft -- <br> $<0.33$ pieces/CW, $25-50 \mathrm{ft}-\mathrm{-}<0.33$ <br> pieces/CW, 50-150 ft -- <0.33 <br> pieces/CW , 150- <br> 400 ft -- <3 <br> pieces/CW with <br> accumulations <br> where they might occur, >400 ft -- <2 pieces/CW with no accumulations where they might occur. |

Large woody debris can have a profound impact on streams, creating important habitat and cover for many fish species, trapping sediment, and providing food for invertebrates (Bilby and Bisson, 1998). While single logs can influence smaller streams, LWD must accumulate in jams to influence habitat in larger rivers (Montgomery and Buffington, 1998). LWD can armor banks, form pools, bars, and side channels (most often in combination with other processes), and can also foster channel avulsion or bank cutting in some cases (Montgomery and Buffington, 1998). LWD is also important in Montgomery and Buffington's channel reach morphology system, as LWD can force reaches into a type of habitat that would typically be expected on a steeper gradient (Montgomery and Buffington, 1998).

## Measurement Techniques

The original EDT assessment proposed a qualitative description of the effect of LWD on stream morphology. It has subsequently been expanded to include a quantitative description of LWD density. The former description may be provided by aerial photo interpretation. The latter description requires a field survey, which may be executed concurrently with the other field surveys described in this document.

## Aerial photo interpretation

LWD may be assessed by airphoto analysis (Johnston and Slaney 1996). LWD is evaluated in terms of its presence, extent, influence on channel morphology and diversity, and effect on cross channel jams. LWD pieces should be at least partially in the stream channel to be considered (USFS, 2001). "Functional LWD" is defined as those LWD pieces that are the primary cause of formation or geome try of a pool (Johnston and Slaney, 1996).

This assessment may be conducted at the same time as riparian function, at similar 250 m or 500 m intervals along the reach. At each measurement site (a cell consisting of the habitat area at the particular interval) evaluate the influence of woody debris on the environment. Note the appropriate environmental correlate for each measurement site. Characterize the entire reach by the most common correlate from all measurement sites along the reach. As with the aerial photo riparian function methodology, this methodology is only appropriate for large streams. Tributaries, streams covered by tree canopy, and streams without aerial photo coverage will require on-the-ground surveys.

## Field survey

Field techniques utilize a tally of LWD pieces, a set of measurements of LWD pieces in the stream channel, or a combination of the two. The USFS (2001) suggests the following (also in Allen and Guenther 1996): record number of pieces of LWD within the bankfull channel for each habitat unit, providing the trunk or root swell (the area between roots and trunk) interacts with stream flow at bankfull conditions. In addition, LWD is counted in small streams only if the tree's length is greater than two times the bankfull width.

Bain and Stevenson (eds.) (1999) suggest measuring LWD at transects across the channel. For each transect section, count and measure diameter (to nearest cm ) of all pieces of woody debris larger than 1 cm in diameter that intersect transect line. Record the number of wood pieces by diameter class $(1-5,6-10,11-50,>51 \mathrm{~cm})$ for each section. Sum number of pieces in each size class for all transect sections. Calculate length of all transect sections and average the number of pieces per section or length of transect. Platts et al. (1987) suggest that the amount of debris may be described as biomass (weight or volume), number of individual pieces, or percentage of stream area covered. For volume measure each end of LWD piece (d1 and d2) with calipers, and measure length with a meter stick or fiber tape $(\mathrm{L}) . \mathrm{V}=\left[\mathrm{p}\left(\mathrm{d} 1^{2}+\mathrm{d} 2^{2}\right) \mathrm{L}\right] / 8$. Weight $=$ volume $\times 0.5$. This method is time consuming. A quicker method is to take counts of individual pieces or accumulations (divide into size of accumulation and position in the stream). A third method is to measure percentage area of stream affected Overton et al. (1997) suggests measuring the length and diameter of all individual pieces in the bankfull channel that meet both of the following criteria: length must be $=3 \mathrm{~m}$ or $>2 / 3$ wetted width of stream. Diameter must be $0.1 \mathrm{~m} 1 / 3$ of the way up the base. Measure the percent of single pieces submerged. For aggregates, count, or estimate if counting is difficult. Measure with a stadia rod, or estimate. If estimation is used, occasionally check the estimate with a measurement.

Schuett-Hames et al. (1999) state that LWD logs, root wads, and jams must meet the following criteria:

LWD $\log$ criteria:

1. Dead
2. Root system (if present) no longer supports weight of stem/bole
3. Minimum diameter 10 cm along 2 m of length; and
4. Minimum 10 cm of length extending into bankfull channel

Note: Forked LWD are counted as 1 piece with diameter taken at midpoint.
$\underline{\text { LWD root wad criteria: }}$

1. Dead
2. Root system detached from original position
3. Minimum diameter of 0.2 m with total length $<2 \mathrm{~m}$ and
4. Minimum 0.1 m of length extending into bankfull channel

LWD jam criteria:

1. Minimum 10 qualifying pieces of LWD either physically touching at $1+$ points, or associated with jam structure
2. Minimum 0.1 m of 1 LWD piece's length extending into bankfull channel

Schuett-Hames et al. (1999) also describe several survey procedures, including a: level 1 survey where one categorizes LWD by diameter and location, including a tally of key pieces; a more detailed level 2 surve y; and a jam survey procedure.

## EDT ATTRIBUTE VALUES FOR THE CLE ELUM AND EASTON REACHES

| EDT Attribute | Cle Elum |  | Easton |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Correlate Value | Index Value | Correlate Value | Index Value |
| Gradient | . 28 | 1 | . 30 | 1 |
| Confinement <br> - Natural <br> - Hydromod | $\begin{gathered} 0 \% \\ 22.44 \% \end{gathered}$ | $\begin{aligned} & 0 \\ & 2 \end{aligned}$ | $\begin{gathered} 2.68 \% \\ 17.58 \% \end{gathered}$ | $\begin{aligned} & 0 \\ & 2 \end{aligned}$ |
| $\begin{aligned} & \text { Channel Width } \\ & \begin{aligned} \text { - } & \text { Maximum } \\ - & \text { Minimum } \end{aligned} \end{aligned}$ | $\begin{aligned} & 185-247 \mathrm{ft} . \\ & 129-159 \mathrm{ft} . \end{aligned}$ | $\begin{aligned} & 3 \\ & 3 \end{aligned}$ | $\begin{gathered} 103-123 \mathrm{ft} . \\ 89-93 \mathrm{ft} . \end{gathered}$ | $\begin{aligned} & 3 \\ & 2 \end{aligned}$ |
| Habitat Type <br> - riffle (s. cobble) <br> - riffle (1. cobble) <br> - glide <br> - pool <br> - pool tailout | $\begin{gathered} 1.98 \% \\ 8.97 \% \\ 84.55 \% \\ 1.25 \% \\ .36 \% \end{gathered}$ | $\begin{aligned} & 1 \\ & 2 \\ & 4 \\ & 1 \\ & 1 \end{aligned}$ |  |  |
| Off-Channel Habitat <br> - side channel <br> - beaver pond | $\begin{aligned} & .02 \mathrm{X} \\ & .007 \mathrm{X} \end{aligned}$ | $\begin{aligned} & 1 \\ & 1 \end{aligned}$ |  |  |
| Riparian Function |  | 3 |  | 1 |
| Large Woody Debris |  | 3 |  | 3 |

## APPENDIX A

RESULTS OF SLOPE GRADIENT ANALYSIS
FOR ALL EDT REACHES WITHIN THE YAKIMA RIVER BASIN
*Note, only those cells colored green have been field verified.
Appendix A1-Slope Gradient Analysis for WRIA 37

|  |  |  |  |  | Gradient/Slo | e Analysis | Cnannel Reach Form |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| WRIA NUMBER | STREAMNAME | WATERSHED | RNAME | DESCRIPTION | GRADIENT | \% SLOPE |  |
| 37 | Agency Cr | Toppenish | Agency Cr.-1 | Mouth to Job Corps site (RM 0 to 6.3). | 0.0142 | 1.4174 | Pool-Riffle |
| 37 | Agency Cr | Toppenish | Agency Cr.-2 | Job Corps site to impassible falls (RM 6.3 to 9.0) | 0.0310 | 3.1040 | Plane-Bed |
| 37 | Ahtanum Cr NF | Ahtanum | Ahtanum Cr. NF-1 | Mouth to historical spring chinook access limit (RM 0 to | 0.0165 | 1.6483 | Plane-Bed |
| 37 | Ahtanum Cr NF | Ahtanum | Ahtanum Cr. NF-2 | Spring chinook access limit to Nasty Cr. (Rm 2.0 to 5.3) | 0.0192 | 1.9162 | Plane-Bed |
| 37 | Ahtanum Cr NF | Ahtanum | Ahtanum Cr. NF-3 | Nasty Cr. to Foundation Cr. (RM 5.3 to 10.2) | 0.0199 | 1.9910 | Plane-Bed |
| 37 | Ahtanum Cr NF | Ahtanum | Ahtanum Cr. NF-4 | Foundation Cr. to MF Ahtanum Cr. (RM 10.2 to 11.6) | 0.0267 | 2.6735 | Plane-Bed |
| 37 | Ahtanum Cr NF | Ahtanum | Ahtanum Cr. NF-5 | MF Ahtanum Cr. to McLain Canyon (RM 11.6 to 13.1)(upper | 0.0351 | 3.5130 | Plane-Bed |
| 37 | Ahtanum Cr NF | Ahtanum | Ahtanum Cr. NF-6 | McLain Canyon to upper access limit for steelhead (RM 13 | 0.0397 | 3.9664 | Plane-Bed |
| 37 | Ahtanum Cr SF | Ahtanum | Ahtanum Cr. SF-1 | Mouth to historical spring chinook access limit (RM 0 to | 0.0174 | 1.7379 | Plane-Bed |
| 37 | Ahtanum Cr SF | Ahtanum | Ahtanum Cr. SF-2 | Spring chinook access limit to coho/steelhead access lim | 0.0282 | 2.8209 | Plane-Bed |
| 37 | Ahtanum Cr | Ahtanum | Ahtanum Cr.-1 | Mouth to Bachelor Cr. Return (RM 0 to 3.2 ) | 0.0084 | 0.8357 | Pool-Riffle |
| 37 | Ahtanum Cr | Ahtanum | Ahtanum Cr.-2 | Bachelor return to Hatton return (RM 3.2 to 8.5) | 0.0105 | 1.0472 | Pool-Riffle |
| 37 | Ahtanum Cr | Ahtanum | Ahtanum Cr.-3 | Hatton return to lower WIP diversion (RM 8.5 to 9.9) | 0.0102 | 1.0234 | Pool-Riffle |
| 37 | Ahtanum Cr | Ahtanum | Ahtanum Cr.-3A (Lower WIP Diversion Dam) | Lower WIP Diversion Dam (RM 9.9). |  |  |  |
| 37 | Ahtanum Cr | Ahtanum | Ahtanum Cr.-4 | Lower WIP Diversion Dam to American Fuit Rd. Bridge (RM | 0.0094 | 0.9357 | Pool-Riffle |
| 37 | Ahtanum Cr | Ahtanum | Ahtanum Cr.-5 | American Fruit Rd. Bridge to Bachelor/Hatton Diversion ( | 0.0112 | 1.1193 | Pool-Riffle |
| 37 | Ahtanum Cr | Ahtanum | Ahtanum Cr.-5A (Bachelor/Hatton Diversion | Bachelor/Hatton Diversion Dam (RM 18.9) |  |  |  |
| 37 | Ahtanum Cr | Ahtanum | Ahtanum Cr.-6 | Bachelor/Hatton Diversion to Upper WIP Diversion Dam (RM | 0.0131 | 1.3102 | Pool-Riffle |
| 37 | Ahtanum Cr | Ahtanum | Ahtanum Cr.-6A (Upper WIP Diversion Dam) | Upper WIP Diversion Dam (RM 19.6) |  |  |  |
| 37 | Ahtanum Cr | Ahtanum | Ahtanum Cr.-7 | Upper WIP Diversion Dam to confluence of NF and SF (RM 1 | 0.0130 | 1.3019 | Pool-Riffle |
| 37 | Ahtanum Cr | Ahtanum | Bachelor Cr.-1 | Bachelor Cr. Diversion at dam to diversion from Bachelor |  |  |  |
| 37 | Bachelor Cr . | Ahtanum | Bachelor Cr.-2 | Bachelor Cr. Re-entry point into Ahtanum (RM 0 ) to diver |  |  |  |
| 37 | Bull Cr | Satus | Bull Cr. | Mouth to headwaters (RM 0 to 1.5). | 0.0450 | 4.4989 | Step-Pool |
| 37 | Corral Canyon | Lower Yakima Trib | Corral Canyon Cr. | Mouth to steelhead access limit (RM 0 to 4.15) | 0.0279 | 2.7945 | Plane-Bed |
| 37 | Dry Cr (Satus | Satus | Dry Cr. (Satus)-1 | Mouth to Fortyday Cr. (Intermittent Zone) (RM 0 to 14.5) | 0.0108 | 1.0845 | Pool-Riffle |
| 37 | Dry Cr (Satus | Satus | Dry Cr. (Satus)-2 | Fortyday to SF Dry Cr. (RM 14.5 to 28.5). | 0.0152 | 1.5191 | Pool-Riffle |
| 37 | Foundation Cr | Ahtanum | Foundation Cr . | Mouth to steelhead/coho access limit (RM 0 to 0.8 ) | 0.0682 | 6.8220 | Step-Pool |
| 37 | Hatton Cr. | Ahtanum | Hatoon Cr. | Re-entry point into Ahtanum to diversion point from Bach | 0.0105 | 1.0514 | Pool-Riffle |
| 37 | Harrah Drain | Lower Yakima Trib | Harrah Drain | Mouth to Impassible RR Br. In Harrah (RM 0 to 5.3 ) |  |  |  |
| 37 | YakimaR | Lower Yakima | Horn Dam | Horn Dam |  |  |  |
| 37 | Kusshi Cr | Satus | Kusshi Cr. | Mouth to headwaters (RM 0 to 5). | 0.0161 | 1.6118 | Plane-Bed |
| 37 | Logy Cr | Satus | Logy Cr. | Mouth to falls (RM 0 to 14). | 0.0175 | 1.7464 | Plane-Bed |
| 37 | Marion Drain | Lower Yakima Trib | Marion Drain-1 | Mouth to Gate Structure (RM 0 to 1.5). | 0.0025 | 0.2463 | Pool-Riffle |
| 37 | Marion Drain | Lower Yakima Trib | Marion Drain-2 (Gate Structure) | Gate Structure (RM 1.5). |  |  |  |
| 37 | Marion Drain | Lower Yakima Trib | Marion Drain-3 | Gate Structure to Wanity Slough (RM 1.5 to 5.0 ). | 0.0001 | 0.0126 | Pool-Riffle |
| 37 | Marion Drain | Lower Yakima Trib | Marion Drain-4 | Wanity Slough to Harrah Drain (RM 5.0 to 17.8) | 0.0010 | 0.1009 | Pool-Riffle |
| 37 | MF Ahtanum Cr | Ahtanum | MF Ahtanum Cr. | Mouth to steelhead/coho access limit (RM 0 to 0.9 ) | 0.0398 | 3.9787 | Plane-Bed |
| 37 | Mule Dry Cr | Satus | Mule Dry Cr. | Mouth to limit of accessibilility (RM 0 to 17). | 0.0126 | 1.2571 | Pool-Riffle |
| 37 | Nasty Cr. | Ahtanum | Nasty Cr. | Mouth to steelhead/coho access limit (RM 0 to 3.7) | 0.0553 | 5.5343 | Step-Pool |
| 37 | Simcoe Cr NF | Toppenish | NF Simcoe Cr. | Mouth to access limit (RM 0 to 4). | 0.0268 | 2.6764 | Plane-Bed |
| 37 | Toppenish Cr | Toppenish | NF Toppenish Cr.-1 | Mouth to Tie Rd. (RM 0 to 2.9) | 0.0354 | 3.5392 | Plane-Bed |
| 37 | Toppenish Cr | Toppenish | NF Toppenish Cr.-2 | Tie Rd to impassible falls (RM 2.9 to 6.6) | 0.0417 | 4.1687 | Step-Pool |
| 37 | Satus Cr | Satus | Satus Cr.-1 | Mouth to Mule Dry (RM 0 to 8.5). | 0.0015 | 0.1455 | Pool-Riffle |
| 37 | Satus Cr | Satus | Satus Cr.-2 | Mule Dry Cr. to Dry Cr. (RM 8.5 to 18.7). | 0.0041 | 0.4143 | Pool-Riffle |
| 37 | Satus Cr | Satus | Satus Cr.-3 | Dry Cr. to Logy Cr. (RM 18.7 to 23.6). | 0.0059 | 0.5890 | Pool-Riffle |
| 37 | Satus Cr | Satus | Satus Cr.-4 | Logy Cr. to Bull Cr. (RM 23.6 to 36). | 0.0102 | 1.0190 | Pool-Riffle |
| 37 | Satus Cr | Satus | Satus Cr.-5 | Bull Cr. to Kusshi Cr. (RM 36 to 37.2) | 0.0159 | 1.5928 | Plane-Bed/Bedrock |
| 37 | Satus Cr | Satus | Satus Cr.-6 | Kusshi to Wilson Charlie Cr. (RM 37.2 to 39.3). | 0.0191 | 1.9149 | $\frac{\text { Pool-Riffle/Plane-Bed }}{}$ |
| 37 | Satus Cr | Satus | Satus Cr.-7 | Wilson Charlie Cr. to Falls (RM 39.3 to 45.0). | 0.0295 | 2.9469 | Plane-Bed |
| 37 | Simcoe Cr SF | Toppenish | SF Simcoe Cr. | Mouth to acess limit (RM 0 to 4). | 0.0276 | 2.7605 | Plane-Bed |
| 37 | Toppensish Cr | Toppenish | SF Toppenish Cr. | Mouth to access limit (RM 0 to 1). | 0.0587 | 5.8711 | Step-Pool |
| 37 | Simcoe CR | Toppenish | Simcoe Cr.-1 | Mouth to Stephenson Rd. (RM 0 to 5.9). | 0.0023 | 0.2344 | Pool-Riffle |
| 37 | Simcoe CR | Toppenish | Simcoe Cr.-2 | Stephenson R. to Agency Cr. (RM 5.9 to 9.5) | 0.0057 | 0.5741 | Pool-Riffle |
| 37 | Simcoe CR | Toppenish | Simcoe Cr.-3 | Agency Cr. to Wesley Rd. (RM 9.5 to 10.4) | 0.0076 | 0.7622 | Pool-Riffle |



Appendix A2-Slope Gradient Analysis for WRIA 38

| 38 | American R | American | American R.-1 | Mouth to Bumping Rd. Br. (RM 0 to 0.96) | 0.0201 | 2.0065 | Pool-Riffle/Step-Pool |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 38 | American R | American | American R.-2 | Bumping R. Br. to Hell's Crossing Br. (RM 0.96 to 5.4) | 0.0169 | 1.6858 | Pool-Riffle/Step-Pool |
| 38 | American R | American | American R.-3 | Hells' Crossing to Miner Cr. (RM 5.4 to 8.7) | 0.0032 | 0.3209 | Pool-Riffle/Step-Pool |
| 38 | American R | American | American R.-3A | Miner Cr. to Kettle Cr. (RM 8.7 to 9.7) | 0.0040 | 0.4028 | Pool-Riffle/Step-Pooi |
| 38 | American R | American | American R.-3B | Kettle Cr. to 4th Br. (RM 9.7 to 11.1) | 0.0068 | 0.6817 | Pool-Riffle/Step-Pool |
| 38 | American R | American | American R.-4 | 4th Br. to Union Cr. (RM 11.1 to 11.5) | 0.0039 | 0.3886 | Pool-Riffle/Step-Pooi |
| 38 | American R | American | American R.-4A | Union Cr. to Lodgepole Campground (RM 11.5 to 12.4) | 0.0080 | 0.7964 | Pool-Riffle/Step-Pool |
| 38 | American R | American | American R.-5 | Lodgepole Campground to Morse Cr. Footbridge (RM 12.4 | 0.0272 | 2.7230 | Pool-Riffle/Step-Pool |
| 38 | American R | American | American R.-6 | Morse Cr. Footbridge to Morse Cr. (RM 14.6 to 15.6) | 0.0037 | 0.3666 |  |
| 38 | American R | American | American R. 6 A | Morse Cr. to Rainier Fork (RM 15.6 to 16.9) | 0.0060 | 0.6044 | Pool-Riffle/Step-Pooi |
| 38 | American R | American | American R.-6B | Rainier Fork to braided section with impassible cascade | 0.0050 | 0.4962 | Pool-Riffle |
| 38 | BearCr. | Naches Trib | Bear Cr. (Lit. Nac | Mouth to low-flow section (RM 0 to 0.5) | 0.0123 | 1.2332 | Pool-Riffle |
| 38 | Blowout Cr. | Naches Trib | Blowout Cr. | Mouth to impassible road crossing at RM 0.6 | 0.0181 | 1.8071 | Step-Pool/Forced Step-Pool |
| 38 | Buckskin Slou | Naches Trib | Buckskin Slough | Mouth to upstream end at RM 5.5 | 0.0065 | 0.6516 | Pool-Riffle |
| 38 | Bumping R | Naches Trib | Bumping R.-1 | Mouth to American R. (RM 0 to 3.5 ). | 0.0105 | 1.0548 | Pool-Riffle |
| 38 | Bumping R | Naches Trib | Bumping R.-2a | American R. to dam (RM 3.5 to 17). | 0.0107 | 1.0730 | $\underline{\text { Pool-Riffle/Plane-Bed }}$ |
| 38 | Bumping R | Naches Trib | Bumping R.-2b(Bump | Bumping Dam |  |  |  |
| 38 | Bumping R | Naches Trib | Bumping R.-3a (Bum | Bumping Lake, Dam to Deep Cr. mouth (RM 17 - 18.9) |  |  |  |
| 38 | Bumping R | Naches Trib | Bumping R.-3b (Bum | Bumping Lake, Deep Cr. mouth to Bumping River lake int |  |  |  |
| 38 | Bumping R | Naches Trib | Bumping R.-4 | Bumping River lake inlet to Cougar Cr. (RM 21.1 to 22. | 0.0264 | 2.6394 | Plane-Bed/Step-Pool |
| 38 | Bumping R | Naches Trib | Bumping R.-5 | Bumping River, Cougar Cr. to falls (RM 22.3 - 25.8 ) | 0.0100 | 1.0034 | Pool-Riffle |
| 38 | Clear Cr | Naches Trib | Clear Cr. | Mouth to falls. (RM 0 to 4). Stopped at fork for routi | 0.0156 | 1.5602 | Plane-Bed |
| 38 | Clear Cr | Naches Trib | ClearCr.-2 | Mouth to falls. (RM 0 to 4). (fork to falls) | 0.1868 | 18.6824 | Cascade |
| 38 | Clear Lake Da | Naches Trib | ClearLake Dam | Clear Lake Dam |  |  |  |
| 38 | Cougar Cr | Naches Trib | Cougar Cr. | Mouth to accessibility limit (RM 0 to 4). | 0.0669 | 6.6924 | Step-Pool |
| 38 | Cowiche Cr | Naches Trib | Cowiche Cr. | Mouth to forks (RM 0 to 7.5 ). | 0.0108 | 1.0787 | Pool-Riffle |
| 38 | Crow Cr. | Naches Trib | Crow Cr. | Mouth to impassible slide at Falls Cr. (RM 0 to 9.8) | 0.0256 | 2.5638 | Plane-Bed/Step-Pool |
| 38 | Deep Cr | Naches Trib | Deep Cr. | Mouth to impassible falls (RM 0 to 4.9). | 0.0149 | 1.4886 | Pool-Riffle/Forced Step-Pool |
| 38 | Hindoo Cr. | Naches Trib | Hindoo Cr . | Mouth to falls (RM 0 to 0.8 ) | 0.0629 | 6.2943 | Step-Pool |
| 38 | Indian Cr | Naches Trib | Indian Cr. (NF Tie | Mouth to impassible falls (RM 0 to 5.1). | 0.0403 | 4.0259 | Step-Pool |
| 38 | American R | American | Kettle Cr. | Mouth to impassibly steep section (RM 0 to 0.4) | 0.0154 | 1.5400 | Plane-Bed |
| 38 | Little Naches | Naches Trib | Little Naches Fall | Little Naches Falls (RM4.5). |  |  |  |
| 38 | Little Naches | Naches Trib | Little Naches R.-1 | Mouth to Crow Cr. (RM 0 to 3.2). | 0.0094 | 0.9357 | Pool-Riffle |
| 38 | Little Naches | Naches Trib | Little Naches R.-2 | Crow Cr. to Quartz Cr. (RM 3.2 to 3.4). | 0.0117 | 1.1702 | Pool-Riffle |
| 38 | Little Naches | Naches Trib | Little Naches R.-3 | Quartz Cr. to Little Naches Falls (RM 3.4 to 4.5). | 0.0103 | 1.0333 | Pool-Riffle |
| 38 | Little Naches | Naches Trib | Little Naches R.-4 | Little Naches Falls to Matthew Cr. (RM 4.5 to 9.5) | 0.0101 | 1.0130 | Pool-Riffle |
| 38 | Little Naches | Naches Trib | Little Naches R.-5 | Mathew Cr. to SF Little Naches R. (RM 9.5 to 9.9). | 0.0066 | 0.6561 | Pool-Riffle |
| 38 | Little Naches | Naches Trib | Little Naches R.-6 | SF L. Naches to Bear Cr. (RM 9.9 to 10.9) | 0.0086 | 0.8625 | Pool-Riffle |
| 38 | Little Naches | Naches Trib | Little Naches R.-7 | Bear Cr. to MF/NF Confluences Little Naches R. (RM 10. | 0.0095 | 0.9522 | Pool-Riffle |
| 38 | Little Rattle | Naches Trib | Little Rattlesnake | Mouth to steelhead access limit (RM 0 to 7.6 ) | 0.0393 | 3.9347 | Plane-Bed/Step-Pool |
| 38 | Matthew Cr. | Naches Trib | Mathew Cr. | Mouth to steep, low-flow section (RM 0 to 3.5 ) | 0.0666 | 6.6614 | Step-Pool |
| 38 | Little Naches | Naches Trib | MF Little Naches R | Mouth to impassible falls (RM 0 to 2.5). | 0.0364 | 3.6404 | Step-Pool/Forced Step-Pool |
| 38 | American R | American | Miner Cr. | Mouth to impassibly steep section (RM 0 to 0.5) | 0.0541 | 5.4067 | Step-Pool |
| 38 | American R | American | Morse Cr. | Mouth to impassibly steep section (RM 0 to 0.3 ) | 0.0232 | 2.3195 | Plane-Bed |
| 38 | Naches R | Naches | Naches R.-1 | Mouth to Cowiche Cr. (RM 0 to 2.7) | 0.0040 | 0.4050 | Pool-Riffle |



Appendix A3-Slope Gradient Analysis for WRIA 39

| 39 | Badger Cr | Upper Yakima Tribs | Badger Cr.-1 | Mouth to Impassible irrigation diversion (RM 0 to 0.7 ) | 0.0072 | 0.7194 | Pool-Riffle |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 39 | Badger Cr | Upper Yakima Tribs | Badger Cr.-2 | Passage barrier to Highline Canal (RM 0.7 to 10.2) | 0.0085 | 0.8484 | Pool-Riffle |
| 39 | Badger Cr | Upper Yakima Tribs | Badger Cr.-3 | Highline Canal to steep, intermittant section (RM 10.2 to 10 . | 0.0123 | 1.2332 | Pool-Riffle |
| 39 | BearCr.(TWay | Teanaway | Bear Cr.(TWay)-1 | Mouth to small falls (spring chinook access limit)(RM 0 to 0.5 ) | 0.0264 | 2.6432 | Plane-Bed |
| 39 | BearCr.(TWay | Teanaway | Bear Cr.(TWay)-2 | Mouth to steelhead access limit (RM 0.5 to 2.0 ) | 0.0370 | 3.7014 | Plane-Bed |
| 39 | Beverly Cr. | Teanaway | Beverly Cr. | Mouth to steelhead access limit (RM 0 to 1.0) | 0.0888 | 8.8780 | Cascade |
| 39 | Big Cr | Upper Yakima Tribs | Big Cr. 1 | Mouth to impassible diversion dam (RM 0 to 2.1). | 0.0143 | 1.4317 | Pool-Riffle |
| 39 | Big Cr | Upper Yakima Tribs | Big Cr.-2 | Diversion dam to estimated spring chinook access limit (RM 2.1 to 4.8) | 0.0167 | 1.6742 | Plane-Bed |
| 39 | $\mathrm{Big}_{\mathrm{Cr}}$ | Upper Yakima Tribs | BigCr. 3 | Spring chinook access limit to Greek Cr. (RM 4.8 to 7.5 ) | 0.0261 | 2.6075 | Plane-Bed |
| 39 | Bie Cr | Upper Yakima Tribs | Big Cr. 4 | Greek Cr. to steelhead access limit (RM 7.5 to 10 ) | 0.0357 | 3.5654 | Plane-Bed |
| 39 | Box Canyon Cr | Upper Yakima Tribs | Box Canyon Cr. | Mouth to forks (RM 0 to 1.4). | 0.0207 | 2.0683 | Plane-Bed |
| 39 | Wilson Cr | Upper Yakima Tribs | Bull Ditch | Bull Ditch from Yakima take-off to Wilson Cr. Outlet | 0.0028 | 0.2794 | Pool-Riffle |
| 39 | Cabin Cr | Upper Yakima Tribs | Cabin Cr. | Mouth to falls (RM0 to 3.3). | 0.024 | 2.4350 | Plane-Bed |
| 39 | Caribou Cr | Upper Yakima Tribs | Caribou Cr.-1 | Mouth to Highline Canal (RM 0 to 10.0) | 0.0119 | 1.1892 | Pool-Riffle |
| 39 | Caribou Cr | Upper Yakima Tribs | Caribou Cr.-2 | Highline Canal to steep, intermittant section (RM 10 to 17.9) | 0.0381 | 3.8105 | Plane-Bed |
| 39 | Cherry Cr | Upper Yakima Tribs | Cherry Cr.-1 | Mouth to Badger Cr. (RM 0 to 0.3) | 0.0003 | 0.0301 | Pool-Riffle |
| 39 | Cherry Cr | Upper Yakima Tribs | Cherry Cr.-2 | Badger Cr. to impassible irrigation diversion (RM0.3 to 1.3) | 0.0046 | 0.4572 | Pool-Riffle |
| 39 | Cherry Cr | Upper Yakima Tribs | Cherry Cr.-3 | Passage barrier to Cooke Cr. (RM 1.3 to 1.6 ) | 0.0039 | 0.3893 | Pool-Riffle |
| 39 | Cherry Cr | Upper Yakima Tribs | Cherry Cr.-4 | Cooke Cr. to Caribou/Park confluence (RM 1.6 to 2.7) | 0.0038 | 0.3764 | Pool-Riffle |
| 39 | ClarkFlatAccl | Upper Yakima Tribs | ClarkFlatHatchery | Taneum Cr. to Swauk Cr. (RM 166.1 to 169.9). | 0.0061 | 0.6092 | Pool-Riffle |
| 39 | Cle Elum R | Upper Yakima Tribs | Cle Elum R.-1 | Mouth to dam (RM0 to 8.2). | 0.0072 | 0.7209 | Pool-Riffle |
| 39 | Cle Elum R | Upper Yakima Tribs | Cle Elum R.-2A (Cle Elum Dam) | Cle Elum Dam |  |  |  |
| 39 | Cle Elum R | Upper Yakima Tribs | Cle Elum R.-2B (Lake Cle Elum) | Dam to Cle Elum R. confluence (RM 8.2 to 15.9). |  |  |  |
| 39 | Cle Elum R | Upper Yakima Tribs | Cle Elum R.-3 | Lake inlet to Cooper R. (RM 15.9 to 19.2). | 0.0060 | 0.6016 | Pool-Riffle |
| 39 | Cle Elum R | Upper Yakima Tribs | Cle Elum R.-4 | Cooper R. to Waptus R. (RM 19.2 to 21.5). | 0.0135 | 1.3513 | Pool-Riffle |
| 39 | Cle Elum R | Upper Yakima Tribs | Cle Elum R.-5 | Waptus R. to headwaters (RM 21.5 to 34.2). | 0.0138 | 1.3767 | Pool-Riffle |
| 39 | Coleman Cr | Upper Yakima Tribs | Coleman Cr.-1 | Mouth to impassible irrigation diversion (RM 0 to 0.5) | 0.0054 | 0.5385 | Pool-Riffle |
| 39 | Coleman Cr | Upper Yakima Tribs | Coleman Cr.-2 | Passage barrier to Highline Canal (RM 0.5 to 10.3) | 0.0113 | 1.1317 | Pool-Riffle |
| 39 | Coleman Cr | Upper Yakima Tribs | Coleman Cr.-3 | Highline Canal to Coleman Falls (RM 10.3 to 16.7) | 0.0294 | 2.9353 | Plane-Bed |
| 39 | Cooke Cr | Upper Yakima Tribs | Cooke Cr.-1 | Mouth to Highline Canal (RM 0 to 10.4) | 0.0128 | 1.2808 | Pool-Riffle |
| 39 | Cooke Cr | Upper Yakima Tribs | Cooke Cr.-2 | Highline Canal to steep, intermittant section (RM 10.4 to 19. | 0.0374 | 3.7372 | Plane-Bed |
| 39 | Cooper R | Upper Yakima Tribs | Cooper R. | Mouth to impassible falls (RM 0 to 3.2). | 0.0216 | 2.1613 | Plane-Bed |
| 39 | Currier Cr. | Upper Yakima Tribs | Currier Cr . | Mouth to steep, intermittant scetion (RM 0 to 8.1) | 0.0135 | 1.3497 | Pool-Riffle |
| 39 | Dickey Cr. | Teanaway | Dickey Cr . | Mouth to steelhead access limit (RM 0 to 1.0) | 0.0315 | 3.1520 | Plane-Bed |
| 39 | East Branch Wilson Cr | Upper Yakima Tribs | East Branch Wilson Cr.-1 | Mouth to impassible irrigation barrier (RM 0 to 1.1) | 0.0072 | 0.7156 | Pool-Riffle |
| 39 | East Branch Wilson Cr | Upper Yakima Tribs | East Branch Wilson Cr.-2 | Passage barrier to point of diversion from Wilson Cr. (RM 1.1 | 0.0111 | 1.1066 | Pool-Riffle |
| 39 | EastonAcclima | Upper Yakima | EastonHatchery | Easton Hatchery without enhancement |  |  |  |
| 39 | Gold Cr | Upper Yakima Tribs | GoldCr. | Mouth to steep, intermittant section (RM 0 to 4.3). | 0.0134 | 1.3450 | Plane-Bed |
| 39 | Greek Cr . | Upper Yakima Tribs | Greek Cr. | Mouth to steelhead access limit (RM 0 to 1.5) | 0.0845 | 8.4504 | Cascade |
| 39 | Indian Cr. (T | Teanaway | Indian Cr. (Teanaway) | Mouth to steelhead access limit (RM 02.8 ) | 0.0292 | 2.9195 | Plane-Bed |
| 39 | Iron Cr. | Upper Yakima Tribs | Iron Cr . | Mouth sto steelhead access limit (RM 0 to 2.0) | 0.0559 | 5.5907 | Plane-Bed/Step-Pool |
| 39 | Jack Cr. | Teanaway | Jack Cr. | Mouth to ateelhead access limit (RM 0 to 2.1) | 0.0193 | 1.9336 | Plane-Bed |
| 39 | Johnson Cr. | Teanaway | Johnson Cr. | Mouth to steelhead access limit (RM 0 to 1.0) | 0.0419 | 4.1871 | Step-Pool/Forced Step-Pool |
| 39 | Jungle Cr. | Teanaway | Jungle Cr. | Mouth to ateelhead access limit (RM 0 to 1.0) | 0.024 | 2.4116 | Plane-Bed |
| 39 | Kachess R | Upper Yakima Tribs | Kachess R.-1 | Mouth to Kachess Dam (RM 0 to 1). | 0.0080 | 0.7993 | Pool-Riffle |
| 39 | Kachess R | Upper Yakima Tribs | Kachess R.-2A (Kachess Dam) | Kachess Dam |  |  |  |
| 39 | Kachess R | Upper Yakima Tribs | Kachess R.-2B (Kachess Lake first reach) | Dam to Box Canyon Cr. (RM 1 to 7.9). | 0.0000 | 0.0001 | Pool-Riffle |
| 39 | Kachess R | Upper Yakima Tribs | Kachess R.-3 (Kachess Lake second reach) | Box Canyon to Kachess R. (RM 7.9 to 11.1) | 0.0002 | 0.0159 | Pool-Riffle |
| 39 | Kachess R | Upper Yakima Tribs | Kachess R.-4 | Lake confluence to falls (RM 11.1 to 11.6). | 0.0196 | 1.9621 | Plane-Bed |
| 39 | Lickr. | Teanaway | Lick Cr. | Mouth to steelhead access limit (RM 0 to 1.5) | 0.0108 | 1.0795 | Pool-Riffle |
| 39 | Little Cr | Upper Yakima Tribs | Little Cr.-1 | Mouth to passge/flow problems associated with subdivision (RM 0 to 0.9) | 0.0118 | 1.1773 | Pool-Riffle |
| 39 | Little Cr | Upper Yakima Tribs | Little Cr.-2 | Subdivion to steep gradient (access limit spring chinook)(RM 0.9 to 3.6 | 0.0212 | 2.1200 | Plane-Bed |
| 39 | Little Cr | Upper Yakima Tribs | Little Cr.-3 | Steep gradient section to estimated steelhead access limit (RM 3.6 to 5 | 0.0374 | 3.7430 | Plane-Bed |
| 39 | Little Naneum | Upper Yakima Tribs | Little Naneum Cr. | Mouth to diversion point out of Lower Naneum Cr. (RM 0 to 4.5) |  |  |  |
| 39 | Lower Naneum | Upper Yakima Tribs | Lower Naneum Cr.-1 | Mouth to Little Naneum Cr. inflow (RM 0 to 0.7) | 0.0022 | 0.2245 | Pool-Riffle |
| 39 | Lower Naneum | Upper Yakima Tribs | Lower Naneum Cr.-2 | Little Naneum Cr. to Coleman Cr. (RM 0.7 to 1.1) | 0.0064 | 0.6370 | Pool-Riffle |
| 39 | Lower Naneum | Upper Yakima Tribs | Lower Naneum Cr.-3 | Coleman Cr. to impassible irrigation diversion (RM 1.1 to 1.9) | 044 | 0.4422 | Pool-Riffle |
| 39 | Lower Naneum | Upper Yakima Tribs | Lower Naneum Cr.-4 | Passage barrier to diversion into Little Naneum Cr. (RM 1.9 to 4.5 ) | 0.0079 | 0.7870 | Pool-Riffle |
| 39 | Lower Naneum | Upper Yakima Tribs | Lower Naneum Cr.-5 | Little Naneum diversion to Highline Canal (RM 4.5 to 11.1) | 0.0140 | 1.3975 | Pool-Riffle |
| 39 | Lower Naneum | Upper Yakima Tribs | Lower Naneum Cr.-6 | Highline Canal to diversion point from Wilson Cr. (RM 11.1 to 13.2) | 0.0182 | 1.8196 | Plane-Bed |



| 39 | Taneum Cr | Upper Yakima Tribs | Taneum Cr.-4 | KRD inflow to Knudsen Diversion Dam (RM 2.6 to 3.5 ) | 0.0129 | 1.2892 |  | Pool-Riffle/Plane-Bed |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 39 | Taneum Cr | Upper Yakima Tribs | Taneum Cr.-4A | Knudsen Diversion Dam (RM 3.5) |  |  |  |  |
| 39 | Taneum Cr | Upper Yakima Tribs | Taneum Cr.-5 | Knudsen Diversion Dam to Taneum C.G. and beginning of confine | 0.0146 | 1.4579 |  | Pool-Riffle/Step-Pool |
| 39 | Taneum Cr | Upper Yakima Tribs | Taneum Cr.-6 | Taneum C.G. to forks (RM 8.2 to 12.7) | 0.0176 | 1.7619 |  | Pool-Riffle/Step-Pool |
| 39 | Teanaway R | Teanaway | Teanaway R.-1 | Mouth to NF Teanaway R. (RM 0 to 10.6). | 0.0065 | 0.6483 |  | Pool-Riffle |
| 39 | Teanaway R | Teanaway | Teanaway R.-2 | NF to MF/WF (RM 10.6 to 11.7). | 0.0070 | 0.7031 |  | Pool-Riffle |
| 39 | Teanaway Accli | Teanaway | TeanawayHatchery | Acclimation Site in this reach: NF T'way, Jack Cr. to Staffor |  |  |  |  |
| 39 | Tucker Cr. | Upper Yakima Tribs | Tucker Cr.-1 | Mouth to impassible KRD siphon (RM 0 to 0.9 ) | 0.0149 | 1.4924 |  | Pool-Riffle |
| 39 | Tucker Cr. | Upper Yakima Tribs | Tucker Cr.-2 | KRD siphon to access limit (RM 0.9 to 2.4) | 0.0310 | 3.1031 |  | Plane-Bed |
| 39 | Umtanum Cr | Upper Yakima Tribs | Umtanum Cr . | Mouth to headwaters (RM 0 to 8). | 0.0246 | 2.4575 |  | Plane-Bed |
| 39 | Upper Naneum | Upper Yakima Tribs | Upper Naneum Cr. | Confluence with Wilson Cr. to steep, intermittant section (RM 0 to 8.5) | 0.0224 | 2.2435 |  | Plane-Bed |
| 39 | Waptus R | Upper Yakima Tribs | Waptus R. | Mouth to impassibly steep section (RM 0 to 3.5 ). | 0.0161 | 1.6128 |  | Plane-Bed |
| 39 | Wenas Cr | Middle Yakima Trib | Wenas Cr.-1 | Mouth to dewatered section above Cottonwood Cr. (RM 0 to 1.4) | 0.0052 | 0.5154 |  | Pool-Riffle |
| 39 | Wenas Cr | Middle Yakima Trib | Wenas Cr.-2 | Mouth to confluence of NF and SF Wenas (RM 1.4 to 22.1) | 0.0101 | 1.0052 |  | Pool-Riffle |
| 39 | Teanaway R WF | Teanaway | WF Teanaway R.-1 | Mouth to spring chinook access limit (RM 0 to 3). | 0.0105 | 1.0486 |  | Pool-Riffle |
| 39 | Teanaway R WF | Teanaway | WF Teanaway R.-2 | Spring chinook access limit to impassibly steep gradient (RM 3 to 7.3 ) | 0.0150 | 1.4992 |  | Pool-Riffle |
| 39 | Whisky Cr. | Upper Yakima Tribs | Whisky Cr.-1 | Mouth to impassible irrigation diversion (RM0 to 2.1) | 0.0103 | 1.0348 |  | Pool-Riffle |
| 39 | Whisky Cr. | Upper Yakima Tribs | Whisky Cr.-2 | Passage barrier to diversion point from Wilson Cr. (RM 2.1 to | 0.0206 | 2.0583 |  | Plane-Bed |
| 39 | Williams Cr. | Upper Yakima Tribs | Williams Cr. | Mouth to steelhead access limit at Cougar Gulch (RM 0 to 2.8 ) | 0.0305 | 3.0544 |  | Plane-Bed/Step-Pool |
| 39 | Wilson Cr | Upper Yakima Tribs | Wilson Cr.-1 | Mouth to Cherry Cr. (RM 0 to 1.1). | 0.0044 | 0.4360 |  | Pool-Riffle |
| 39 | Wilson Cr | Upper Yakima Tribs | Wilson Cr.-10 | Diversion point of Whisky Cr. to diversion point of Lower Naneum Cr. (R | 0.0179 | 1.7866 |  | Plane-Bed |
| 39 | Wilson Cr | Upper Yakima Tribs | Wilson Cr.-11 | Diversion point of Lower Naneum Cr. to confluence with Upper Naneum Cr. | 0.0184 | 1.8420 |  | Plane-Bed |
| 39 | Wilson Cr | Upper Yakima Tribs | Wilson Cr.-12 | Upper Naneum Cr. confluence to steep, intermittant section (RM 19.5 to | -0.1626 | -16.2554 |  | Plane-Bed |
| 39 | Wilson Cr | Upper Yakima Tribs | Wilson Cr.-2 | Cherry Cr. to Lower Naneum Cr. (RM 1.1 to 1.7 ) | 0.0007 | 0.0744 |  | Pool-Riffle |
| 39 | Wilson Cr | Upper Yakima Tribs | Wilson Cr.-3 | Lower Naneum to East Branch Wilson Cr. (RM 1.7 to 5.9) |  |  |  |  |
| 39 | Wilson Cr | Upper Yakima Tribs | Wilson Cr.-4 | Mouth of East Branch Wilson Cr.to Bull Ditch inlet (RM 5.9 to 7.8 ) |  |  |  |  |
| 39 | Wilson Cr | Upper Yakima Tribs | Wilson Cr.-4A | Bull Ditch inlet to mouth of Mercer Cr. (RM 7.8 to 8.5 ) | 0.0032 | 0.3214 |  | Pool-Riffle |
| 39 | Wilson Cr | Upper Yakima Tribs | Wilson Cr.-5 | Mercer Cr. mouth to impassible irrigation diversion (RM 8.5 to 9.0 ) | 0.0090 | 0.8960 |  | Pool-Riffle |
| 39 | Wilson Cr | Upper Yakima Tribs | Wilson Cr.-6 | Passage barrier to diversion point of East Branch Wilson Cr. (RM 9.0 to | 0.0110 | 1.0976 |  | Pool-Riffle |
| 39 | Wilson Cr | Upper Yakima Tribs | Wilson Cr.-7 | Diversion point of East Branch Wilson Cr. to Highline Canal (RM 11.3 to | 0.0147 | 1.4712 |  | Pool-Riffle |
| 39 | Wilson Cr | Upper Yakima Tribs | Wilson Cr.-8 | Highline Canal to diversion point of Mercer Cr. (RM 15.9 to 16.4) | 0.0181 | 1.8132 |  | Plane-Bed |
| 39 | Wilson Cr | Upper Yakima Tribs | Wilson Cr.-9 | Diversion point of Mercer Cr. to diversion point of Whisky Cr. (RM 16.4 | 0.0154 | 1.5411 |  | Pool-Riffle |
| 39 | YakimaR | Upper Yakima | Yakima R. 10 | Umtanum Cr. to Wilson Cr. (RM 139.8 to 147). | 0.0019 | 0.1918 |  | Pool-Riffle |
| 39 | YakimaR | Upper Yakima | Yakima R.-11 | Wilson Cr. to Bull Ditch outtake (RM 147 to 153.5). | 0.0026 | 0.2638 |  | Pool-Riffle |
| 39 | YakimaR | Upper Yakima | Yakima R.-11A | Bull Ditch outtake to Reecer Cr. (RM 153.5 to 153.7). | 0.0025 | 0.2488 |  | Pool-Riffle |
| 39 | YakimaR | Upper Yakima | Yakima R.-11B | Reecer Cr. to Manastash Cr. (RM 153.7 to 154.5 ) | 0.0005 | 0.0541 |  | Pool-Riffle |
| 39 | YakimaR | Upper Yakima | Yakima R.-12 | Manastash Cr. to Taneum Cr. (RM 154.5 to 166.1). | 0.0026 | 0.2628 |  | Pool-Riffle |
| 39 | YakimaR | Upper Yakima | Yakima R.-13 | Taneum Cr. to Swauk Cr. (RM 166.1 to 169.9). | 0.0028 | 0.2753 |  | Pool-Riffle |
| 39 | YakimaR | Upper Yakima | Yakima R.-14 | Swauk Cr. to Teanaway R. (RM 169.9 to 176.1). | 0.0022 | 0.2233 |  | Pool-Riffle |
| 39 | YakimaR | Upper Yakima | Yakima R.-15 | Teanaway R. to Cle Elum R. (RM 176.1 to 185.6). | 0.0025 | 0.2458 |  | Plane-Bed/Pool-Riffle |
| 39 | YakimaR | Upper Yakima | Yakima R.-16 | Cle Elum R. to Little Cr. (RM 185.6 to 194.6). | 0.0028 | 0.2834 |  | Plane-Bed/Pool-Riffle |
| 39 | YakimaR | Upper Yakima | Yakima R.-17 | Little Cr. to Big Cr. (RM 194.6 to 195.8). | 0.0016 | 0.1633 |  | Plane-Bed/Pool-Riffle |
| 39 | YakimaR | Upper Yakima | Yakima R.-17A | Little Cr. to Tucker Cr. (RM 194.6 to 199.9) | 0.0022 | 0.2226 |  | Plane-Bed/Pool-Riffle |
| 39 | Yakima R. | Upper Yakima | Yakima R.-18 | Tucker Cr. to Easton Dam/Lake Easton (RM 199.9 to 202.5). | 0.0052 | 0.5193 |  | Plane-Bed/Pool-Riffle |
| 39 | YakimaR | Upper Yakima | Yakima R.-19A (Easton Dam) | Easton Dam (RM 202.5) |  |  |  |  |
| 39 | YakimaR | Upper Yakima | Yakima R.-19B (Lake Easton) | Lake Easton -- Easton Dam to Kachess R. (RM 202.5 to 203.4). |  |  |  |  |
| 39 | YakimaR | Upper Yakima | Yakima R.-20 | Kachess R. (upstream end of Lake Easton) to Cabin Cr. (RM 203 | 0.0024 | 0.2423 |  | Pool-Riffle |
| 39 | YakimaR | Upper Yakima | Yakima R.-21 | Cabin Cr. to Keechelus Dam (RM 205 to 214.5). | 0.0063 | 0.6346 |  | Pool-Riffle |
| 39 | YakimaR | Upper Yakima | Yakima R.-22A (Keechelus Dam) | Keechelus Dam |  |  |  |  |
| 39 | YakimaR | Upper Yakima | Yakima R.-22B (Keechelus Lake) | Dam to Gold Cr. (RM 214.5 to 220). |  |  |  |  |
| 39 | YakimaR | Middle Yakima | Yakima R.-7 | Naches R. to Wenas Cr. (RM 116.3 to 122.4) | 0.0019 | 0.1910 |  | Pool-Riffle |
| 39 | YakimaR | Middle Yakima | Yakima R.-8 | Wenas Cr. to Roza Dam (RM 122.4 to 127.9). | 0.0027 | 0.2681 |  | Pool-Riffle |
| 39 | YakimaR | Upper Yakima | Yakima R.-9A (Roza Dam) | Roza Dam (RM 127.9). |  |  |  |  |
| 39 | YakimaR | Upper Yakima | Yakima R.-9B | Roza Dam to Umtanum Cr. (RM 127.9 to 139.8). | 0.0019 | 0.1938 |  | Pool-Riffle |

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## APPENDIX B

2002 Annual Report
Wild and Hatchery Smolt Survival of Roza Spring Chinook Releases

# IntSTATS 

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# 2002 Annual Report Wild and Hatchery Smolt Survival of Roza Spring Chinook Releases 

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Submitted April 1, 2003

## 1. Introduction

In outmigration years 1999, 2000, 2001, and 2002, outmigrating spring Chinook smolt were trapped at Roza, anesthetized, and PIT-tagged if not previously tagged in OCT-SNT raceways. The fish were identified as to whether they were wild in origin (wild: not adipose-fin clipped), were hatchery fish that were previously PIT-tagged (tagged: adipose-fin clipped and PIT-tag found), or were hatchery fish not previously tagged (untagged: adipose-fin clipped but PIT-tag not found). Fish that were not previously tagged (wild and untagged) were then tagged, and all tagged fish, including previously tagged fish, were measured for fork-length and released.

The main purpose of this trial was to determine whether there was a difference in wild and hatchery release-to-smolt survival indices. The survival index was estimated from release to McNary passage using the same estimation procedures that were used to estimate OCT and SNT survival to McNary (refer to Doug Neeley’s 2002 Annual Report: OCT-SNT Survival).

Data from releases were pooled into weekly groupings. The numbers of fish released within a given week were pooled, and the numbers of expanded McNary detections ${ }^{1}$ from those weeks of release were also pooled. Within each weekly groupings, the total of the pooled expanded McNary detections was divided by the pooled release number as an index of release-to-McNary survival. The estimations were performed separately for wild, tagged hatchery, and untagged hatchery fish. The week was selected such that the beginning date of the week shared a common starting Julian date over years; for example, one week began with Julian date 28, the next began with Julian date $35(35=28+7)$, the next began with Julian date $42(42=35+7)$, etc. The same Julian dates were used for each year. The weekly survival index estimates are given in Appendix A.

[^1]
## 2. Wild versus Hatchery Survival Indices

First, survival indices for previously tagged and untagged hatchery fish were compared to determine whether their estimates could be pooled to provide more precise estimates of survival indices for hatchery fish. The weighted logistic analysis of variation was performed in which the data-summary weeks for which there were hatchery releases were treated as blocks. The weighting variable was the number of fish released during the weekly blocks. In the analysis, tagged and untagged treatment means, adjusted for block effects, were compared. In none of the four years did the tagged and untagged mean survival indices differ significantly $(\mathrm{P}=0.552, \mathrm{P}=1.000, \mathrm{P}=0.248$, and $\mathrm{P}=0.362$ respectively for out-migration years 1999-2002; see Appendix B). The tagged and untagged estimates were therefore pooled within blocks for common estimates of hatchery survival indices.

Wild and pooled hatchery survival indices for each Julian week are presented in Figures 1.a., through 1.d. respectively for outmigration years 1999 through $2002^{2}$ (brood years 1997 through 2000, respectively). Formal weighted logistic analyses of variation and means for 1999 through 2002 survival indices are respectively presented in Tables 1.a through 1.d. using as blocks Julian weeks having both wild and hatchery estimates. There was a significant difference only in outmigration- year 2000 [ $\mathrm{P}=0.001$ in year 2000, Table 1.b.1)]. In that year the survival index of the wild was greater than that of the hatchery. In outmigration years 1999 and 2002 the survival index of wild also exceeded that of hatchery, but the differences were not significant [respectively, $\mathrm{P}=$ 0.252 , Table 1.a.1); $\mathrm{P}=0.212$, Table 1.d.1)].

In outmigration-year 2001, the wild survival index was less then that of the hatchery, but not significantly so $[\mathrm{P}=0.619$, Table 1.c.1)]. In last year's annual report, a significant difference was attributed to outmigration-year 2001. This conclusion was based on an analysis error in which Prosser detections of hatchery fish were inadvertently used for McNary detections, and the 2001 outmigration-year survival-indices presented in that report are incorrect. It should be noted that outmigration-year 2001 had protracted, record-low flows. The outmigration-2001 analyses, means, and data presented in this report have been corrected.

As can be seen in Figure 1.c., the wild fish survival index was much higher for releases prior to the outmigration of hatchery fish than sur vival index of wild fish released concurrently with hatchery fish. The week beginning Julian date 77 when hatchery fish were first being trapped at Roza corresponded to a large drop in wild fish survival. Wild fish survival prior to Julian date 77 consistently exceeded that of both wild and hatchery fish from Julian date 77 onward; if wild fish passed McNary prior to

[^2]April 1 (Julian Date 91), the date McNary detectors came on line, the pre-Julian date 77 wild survival estimate may actually have been relatively higher than indicated in Figure 1.c.

Fish lengths are plotted in Figure 2.a., 2.b., 2.c., and 2.d., respectively for 1999, 2000, 2001, and 2002 outmigrants. In 1999, 2001, and 2002 wild fish tagged at Roza were significantly smaller than hatchery fish [respectively $\mathrm{P}=0.007$, Table 2.a.1); $\mathrm{P}<0.001$, Table 2.c.1); $\mathrm{P}=<0.001$, Table 2.d.1)]. In the outmigration-year 2000, the mean lengths of hatchery and wild fish were nearly identical [Table 2.b.2); $\mathrm{P}=0.754$, Table 2.b.1)].

## 3. Informal Descriptive Findings

The initiation of trapping and release at Roza varied dramatically over years, starting late (early April) in water-year 1999, very early in water-year 2000 and 2002 (actually commencing in the previous calendar year), and not as early in 2001 (late January). Survival indices using wild fish released on or after Julian date 98 can be used as an indicator for survival comparisons among years; the week commencing Julian date 98 being the latest starting week of trapping for the four years of evaluation. For survival of fish released on or after Julian date 98, the wild indicator survival index was highest for outmigration-year 1999 ( $64.5 \%$ survival); 1999 also had the second highest flows on record. The wild indicator was lowest for outmigration- year 2001 ( $8.9 \%$ survival); 2001 had the lowest protracted flow on record. The wild indicator was moderate in years 2000 and 2002 ( $36.5 \%$ and $26.0 \%$, respectively). Comparable hatchery indicators for Julian date 98 onward followed the same general trend; highest in 1999 (50.8\%), lowest in 2001 ( $12.5 \%$ ); in between in 2000 and 2002 ( $34.2 \%$ and $16.4 \%$, respectively). The low hatchery survival-index in 2002 may be associated with the effect of BKD observed in the hatchery in that year ${ }^{3}$.

[^3]Figure 1. Outmigration-Year 1999-2002 (Brood-Year 1997-2000) Survival Indices to McNary Dam of Wild and Hatchery Spring Chinook Smolt Released at Roza Dam. (NOTE: Numbers above bars in figures are number of fish released at Roza.)


Figure 1. Outmigration-Year 1999-2002 (Brood-Year 1997-2000) Survival Indices to McNary Dam of Wild and Hatchery Spring Chinook Smolt Released at Roza Dam (continued)



Table 1. Outmigration-Year 1999-2002 (Brood-Year 1997-2000) Logistic Analysis of Variation between Wild and Hatchery Survival Indices adjusted for common Julian Weeks for which both estimates are available and Associated Means

Table 1.a.1) 1999 Weighted* Survival Index Logistic Analysis of Variation

|  |  | Degrees of <br> Deviance <br> (Dev) | Mean <br> Feveedom <br> (DF) | Type 1 <br> (Dev.DF) | Tyance <br> F- <br> Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Source | 53.40 | 7 | 7.63 |  |  |
| P |  |  |  |  |  |

Table 1.a.2) 1999 Wild and Hatchery Mean Release-to-McNary Survival Indices

|  | Hatchery | Wild |
| :---: | :---: | :---: |
| Mean over common Julian Weeks (adjusted**) | 0.549 | 0.611 |
| Mean over common Julian Weeks (unadjusted**) | 0.508 | 0.645 |

Table 1.b.1) 2000 Weighted* Survival Index Logistic Analysis of Variation

|  | Degrees of <br> Freedom <br> Source |  |  |  | Mean <br> Deviance <br> (Dev.DF) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| (Dev) | F- <br> Ratio | Type 1 <br> P |  |  |  |
| Julian Week and Wild versus Hatchery | 255.89 | 15 | 17.06 |  |  |
| Julian Week | 168.71 | 14 | 12.05 |  |  |
| Wild versus Hatchery (adjusted for Julian Week) | 87.18 | 1 | 87.18 | 17.43 | 0.0013 |
| Error | 60.02 | 12 | 5.00 |  |  |

Table 1.b.2) 2000 Wild and Hatchery Mean Release-to-McNary Survival Indices

|  | Hatchery | Wild |
| :---: | :---: | :---: |
| Mean over common Julian Weeks (adjusted ${ }^{\star \star}$ ) | 0.236 | 0.487 |
| Mean over common Julian Weeks (unadjusted ${ }^{\star \star}$ ) | 0.318 | 0.463 |

*Variable is survial index, weight is number of released fish
** Respectively adjusted and unadjusted for common Julian Week effects

Table 1.c.1) 2001 Weighted* Survival Index Logistic Analysis of Variation

|  | $\begin{array}{c}\text { Degrees of } \\ \text { Deviance } \\ \text { (Dev) }\end{array}$ |  |  |  | $\begin{array}{c}\text { Mean } \\ \text { Sourcedom } \\ \text { (DF) }\end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | \(\left.\begin{array}{c}Deviance <br>

(Dev.DF)\end{array} \quad $$
\begin{array}{c}\text { F- } \\
\text { Ratio }\end{array}
$$ $$
\begin{array}{c}\text { Type 1 } \\
\text { P }\end{array}
$$\right]\)

Table 1.c.2) 2001 Wild and Hatchery Mean Release-to-McNary Survival Indices

|  | Hatchery | Wild |
| :---: | :---: | :---: |
| Mean over common Julian Weeks (adjusted**) | 0.135 | 0.149 |
| Mean over common Julian Weeks (unadjusted**) | 0.174 | 0.132 |

[^4]Table 1. Outmigration-Year 1999-2002 (Brood-Year 1997-2000) Logistic Analysis of Variation between Wild and Hatchery Survival Indices adjusted for common Julian Weeks for which both estimates are available and Associated Means (continued)

Table 1.d.1) 2002 Weighted* Survival Index Logistic Analysis of Variation

|  | $\begin{array}{c}\text { Degrees of } \\ \text { Sreedom } \\ \text { Source }\end{array}$ |  |  |  | $\begin{array}{c}\text { Mean } \\ \text { Deviance } \\ \text { (Dev.DF) }\end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | \(\left.\begin{array}{cccc}F- <br>

Ratio\end{array} $$
\begin{array}{c}\text { Type 1 } \\
\text { P }\end{array}
$$\right]\)

Table 1.d.2) 2002 Wild and Hatchery Mean Release-to-McNary Survival Indices

|  | Hatchery | Wild |
| :---: | :---: | :---: |
| Mean over common Julian Weeks (adjusted ${ }^{\star \star}$ ) | 0.280 | 0.339 |
| Mean over common Julian Weeks (unadjusted ${ }^{\star \star}$ ) | 0.276 | 0.347 |

[^5]Figure 2. Outmigration-Year 1999-2002 (Brood-Year 1997-2000) Mean Lengths of Wild and Hatchery Spring Chinook Smolt at Roza Dam


Fork Lengths
b. Outmigration-Year 2000


N Wild $7 /$ Hatchery

Figure 2. Outmigration-Year 1999-2002 (Brood-Year 1997-2000) Mean Lengths of Wild and Hatchery Spring Chinook Smolt at Roza Dam (continued)


Fork Lengths
d. Outmigration-Year 2002


Nwild $7 /$ Hatchery

Table 2. Outmigration-Year 1999-2002 (Brood-Year 1997-2000) Analysis of Variance and Means of Rosa-Released Wild and Hatchery Fish Lengths (mm) for Weeks of Common Passage

Table 2.a.1) 1999 Wild versus Hatchery: Fish Length Least Squares Analysis of Variance

|  | Degrees of <br> Deviance <br> (Deveedom |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Source | Mean <br> (DF) | Deviance <br> (Dev.DF) | F- <br> Ratio | Type 1 <br> P |  |
| Julian Week and Wild versus Hatchery | 451.45 | 7 | 64.49 |  |  |
| Julian Week | 347.76 | 6 | 57.96 |  |  |
| Wild versus Hatchery (adjusted for Julian Week) | 103.69 | 1 | 103.69 | 16.26 | 0.0069 |
| Error | 38.26 | 6 | 6.38 |  |  |

Table 2.a.2) 1999 Mean Fish Lengths at Release

|  | Hatchery | Wild |
| :---: | :---: | :---: |
| Mean over common Julian Weeks | 126.4 | 121.0 |

Table 2.b.1) 2000 Wild versus Hatchery: Fish Length Least Squares Analysis of Variance

|  | $\begin{array}{c}\text { Degrees of } \\ \text { Sreedom } \\ \text { Source }\end{array}$ |  |  |  | $\begin{array}{c}\text { Mean } \\ \text { Deviance } \\ \text { (Dev.DF) }\end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | \(\left.\begin{array}{c}F- <br>

Fationce <br>
Rev)\end{array} $$
\begin{array}{c}\text { Type 1 } \\
\text { P }\end{array}
$$\right]\)

Table 2.b.2) 2000 Mean Fish Lengths at Release

|  | Hatchery | Wild |
| :---: | :---: | :---: |
| Mean over common Julian Weeks | 110.7 | 110.4 |

Table 2.c.1) 2001 Wild versus Hatchery: Fish Length Least Squares Analysis of Variance

|  | Degrees of <br> Source |  |  |  | Mean <br> Deviance <br> (Dev) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| (DF) <br> (Deviance | F- <br> (Dev.DF) | Type 1 <br> Ratio | P |  |  |
| Julian Week and Wild versus Hatchery | 517.90 | 5 | 103.58 |  |  |
| Julian Week | 230.60 | 4 | 57.65 |  |  |
| Wild versus Hatchery (adjusted for Julian Week) | 287.30 | 1 | 287.30 | 85.61 | 0.0008 |
| Error | 13.42 | 4 | 3.36 |  |  |

Table 2.c.2) 2001 Mean Fish Lengths at Release

|  | Hatchery | Wild |
| :---: | :---: | :---: |
| Mean over common Julian Weeks | 128.8 | 118.0 |

Table 2. Outmigration-Year 1999-2002 (Brood-Year 1997-2000) Analysis of Variance and Means of Rosa-Released Wild and Hatchery Fish Lengths (mm) for Weeks of Common Passage (continued)

Table 2.d.1) 2002 Wild versus Hatchery: Fish Length Least Squares Analysis of Variance

|  | $\begin{array}{c}\text { Degrees of } \\ \text { Freedom } \\ \text { Source }\end{array}$ |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | \(\left.\begin{array}{c}Mean <br>

Deviance <br>
(Dev.DF)\end{array} \quad $$
\begin{array}{c}\text { F- } \\
\text { Ratio }\end{array}
$$ $$
\begin{array}{c}\text { Type 1 } \\
\text { P }\end{array}
$$\right]\)

Table 2.d.2) 2002 Mean Fish Lengths at Release

|  | Hatchery | Wild |
| :---: | :---: | :---: |
| Mean over common Julian Weeks | 124.2 | 105.6 |

Appendix A. 1999-2002 Outmigration-Year (1997-2000 Brood-Year) Mean Roza-to-McNary Smolt Survival Indices and Mean Fork Lengths at Roza Release

Table A.1.a. 1999 Outmigrant Roza-to-McNary Survival Indices and Fork-Lengths at Roza

| Week <br> Beginning Julian Date | Wild |  |  | Hatchery |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Released | S.I. | Length | Previously Untagged |  | Previously Tagged |  | Combined |  | Length |
|  |  |  |  | Released | S.I. | Released | S.I. | Release | S.I. |  |
| 98 | 29 | 0.2812 | 111.5 | 77 | 0.5637 | 153 | 0.3727 | 230 | 0.4366 | 120.7 |
| 105 | 13 | 0.7420 | 117.2 | 21 | 0.2767 | 45 | 0.3740 | 66 | 0.3430 | 118.1 |
| 112 | 54 | 0.6032 | 120.7 | 42 | 0.7465 | 36 | 0.8097 | 78 | 0.7757 | 128.4 |
| 119 | 57 | 0.6316 | 122.0 | 107 | 0.4228 | 205 | 0.5926 | 312 | 0.5344 | 128.9 |
| 126 | 20 | 0.6525 | 118.5 | 21 | 0.4930 | 44 | 0.3484 | 65 | 0.3951 | 127.4 |
| 133 | 48 | 0.6068 | 126.4 | 48 | 0.6543 | 96 | 0.5109 | 144 | 0.5587 | 129.7 |
| 140 | 91 | 0.7995 | 130.6 | 46 | 0.6734 | 92 | 0.4045 | 138 | 0.4941 | 131.8 |

Table A.1.b. 2000 Outmigrant Roza-to-McNary Survival Indices and Fork-Lengths at Roza

| Week Beginning Julian Date | Wild |  |  | Hatchery |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Released | S.I. | Length | Previously Untagged |  | Previously Tagged |  | Combined |  | Length |
|  |  |  |  | Released | S.I. | Released | S.I. | Releas | S.I. |  |
| 344 | 62 | 0.2092 | 104.0 |  |  |  |  |  |  |  |
| 351 | 0 | 0.0000 | 0.0 |  |  |  |  |  |  |  |
| 358 | 55 | 0.3700 | 102.0 |  |  |  |  |  |  |  |
| 365 | 1369 | 0.3009 | 104.4 |  |  |  |  |  |  |  |
| 7 | 971 | 0.3091 | 105.3 |  |  |  |  |  |  |  |
| 14 | 515 | 0.2867 | 106.0 |  |  |  |  |  |  |  |
| 21 | 181 | 0.3395 | 107.3 | 0 | 0.0000 | 2 | 0.0000 | 2 | 0.0000 | 106.0 |
| 28 | 467 | 0.3977 | 107.8 | 1 | 0.0000 | 17 | 0.5437 | 18 | 0.5135 | 98.3 |
| 35 | 326 | 0.4315 | 106.8 | 0 | 0.0000 | 6 | 0.5540 | 6 | 0.5540 | 109.6 |
| 42 | 330 | 0.5339 | 105.8 | 0 | 0.0000 | 3 | 0.0000 | 3 | 0.0000 | 106.8 |
| 49 | 146 | 0.4081 | 105.7 | 0 | 0.0000 | 1 | 0.0000 | 1 | 0.0000 | 109.0 |
| 56 | 819 | 0.5159 | 107.5 | 19 | 0.2320 | 49 | 0.3464 | 68 | 0.3145 | 107.5 |
| 63 | 230 | 0.2947 | 108.1 | 115 | 0.3405 | 39 | 0.2183 | 154 | 0.3095 | 108.5 |
| 70 | 77 | 0.6486 | 105.9 | 78 | 0.3261 | 5 | 0.0000 | 83 | 0.3065 | 110.7 |
| 77 | 272 | 0.4870 | 113.0 | 558 | 0.3231 | 182 | 0.3529 | 740 | 0.3304 | 112.6 |
| 84 | 77 | 0.5426 | 106.8 | 150 | 0.2654 | 64 | 0.6232 | 214 | 0.3724 | 108.5 |
| 91 | 171 | 0.6118 | 112.0 | 351 | 0.2639 | 328 | 0.2342 | 679 | 0.2495 | 112.8 |
| 98 | 0 | 0.0000 | 0.0 | 0 | 0.0000 | 0 | 0.0000 | 0 | 0.0000 | 0.0 |
| 105 | 36 | 0.3693 | 120.68249 | 300 | 0.3871 | 180 | 0.3933 | 480 | 0.3894 | 121.45399 |
| 112 | 23 | 0.3834 | 127.175 | 201 | 0.4786 | 71 | 0.2806 | 272 | 0.4269 | 126.88444 |
| 119 | 41 | 0.3507 | 131.13756 | 177 | 0.1935 | 102 | 0.1516 | 279 | 0.1782 | 135.59597 |

Table A.1.c. 2001 Outmigrant Roza-to-McNary Survival Indices and Fork-Lengths at Roza


## Appendix A. 1999-2002 Outmigration-Year (1997-2000 Brood-Year) Mean Roza-to-McNary Smolt Survival Indices and Mean Fork Lengths at Roza Release (continued)

Table A.1.d. 2002 Outmigrant Roza-to-McNary Survival Indices and Fork-Lengths at Roza

| Week Beginning Julian Date | Wild |  |  | Hatchery |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Released | S.I. | Length | Previously Untagged |  | Previously Tagged |  | Combined |  | Length |
|  |  |  |  | Released | S.I. | Released | S.I. | Releas | S.I. |  |
| 351 | 500 | 0.1794 | 105.0 |  |  |  |  |  |  |  |
| 358 | 501 | 0.1361 | 102.9 |  |  |  |  |  |  |  |
| 365 | 230 | 0.3191 | 101.1 |  |  |  |  |  |  |  |
| 7 | 576 | 0.2487 | 99.8 |  |  |  |  |  |  |  |
| 14 | 1210 | 0.1682 | 96.2 |  |  |  |  |  |  |  |
| 21 | 490 | 0.1586 | 96.6 |  |  |  |  |  |  |  |
| 28 | 202 | 0.3216 | 96.9 |  |  |  |  |  |  |  |
| 35 | 187 | 0.1688 | 96.8 |  |  |  |  |  |  |  |
| 42 | 548 | 0.2199 | 97.9 |  |  |  |  |  |  |  |
| 49 | 762 | 0.2803 | 100.9 |  |  |  |  |  |  |  |
| 56 | 190 | 0.2080 | 93.8 |  |  |  |  |  |  |  |
| 63 | 484 | 0.2950 | 98.0 |  |  |  |  |  |  |  |
| 70 | 724 | 0.3220 | 103.6 |  |  |  |  |  |  |  |
| 77 | 558 | 0.3869 | 102.5 | 309 | 0.3961 | 71 | 0.4561 | 380 | 0.4073 | 123.8 |
| 84 | 500 | 0.3143 | 104.3 | 251 | 0.3201 | 24 | 0.0812 | 275 | 0.2992 | 123.3 |
| 91 | 777 | 0.3327 | 104.8 | 379 | 0.2636 | 71 | 0.1782 | 450 | 0.2501 | 122.8 |
| 98 | 47 | 0.4212 | 106.7 | 100 | 0.1940 | 8 | 0.3229 | 108 | 0.2036 | 121.2 |
| 105 | 0 |  | 0.0 | 0 |  | 0 |  | 0 |  | 0.0 |
| 112 | 232 | 0.3560 | 109.6 | 233 | 0.1772 | 57 | 0.0323 | 290 | 0.1487 | 129.7 |

## Appendix B. Weighted Logistic Analysis of Variation between Tagged and Untagged Hatchery Survival Indices adjusted for common Blocks (Julian Weeks for which both estimates are available) and Associated Means

B.1.a. 1999 Tagged versus Untagged: Weighted* Survival Index Logistic Analysis of Variation

| Source | $\begin{gathered} \text { Deviance } \\ \text { (Dev) } \\ \hline \end{gathered}$ | Degrees of Freedom (DF) | Mean Deviance (Dev.DF) | FRatio | Type 1 Error P |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Julian Week and Tagged versus Untagged | 43.35 | 7 | 6.19 | 0.40 | 0.5525 |
| Julian Week | 41.51 | 6 | 6.92 |  |  |
| Tagged versus Untagged (adjusted for Julian Week) | 1.84 | 1 | 1.84 |  |  |
| Error | 27.90 | 6 | 4.65 |  |  |

B.1.b. 1999 Untagged and Tagged Hatchery Mean Release-to-McNary Survival Indices

|  | Untagged | Tagged |
| :---: | :---: | :---: |
| Mean over common Julian Weeks (adjusted**) | 0.553 | 0.487 |
| Mean over common Julian Weeks (unadjusted ${ }^{* \star}$ ) | 0.548 | 0.486 |

*Variable is survial index, weight is number of released fish
** Respectively adjusted and unadjusted for common Julian Week effects
B.2.a. 2000 Tagged versus Untagged: Weighted* Survival Index Logistic Analysis of Variation

|  | $\begin{array}{c}\text { Degrees of } \\ \text { Sreedom } \\ \text { Source }\end{array}$ |  |  |  | $\begin{array}{c}\text { Mean } \\ \text { Deviance } \\ \text { (Dev.DF) }\end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | \(\left.\begin{array}{c}F- <br>

Fationce <br>
(Dev)\end{array} $$
\begin{array}{c}\text { Type 1 } \\
\text { P }\end{array}
$$\right]\)
B.2.b. 2000 Untagged and Tagged Hatchery Mean Release-to-McNary Survival Indices

|  | Untagged | Tagged |
| :---: | :---: | :---: |
| Mean over common Julian Weeks (adjusted**) | 0.317 | 0.297 |
| Mean over common Julian Weeks (unadjusted**) | 0.322 | 0.310 |

Appendix B. Weighted Logistic Analysis of Variation between Tagged and Untagged Hatchery Survival Indices adjusted for common Blocks (Julian Weeks for which both estimates are available) and Associated Means (continued)
B.3.a. 2001 Tagged versus Untagged: Weighted* Survival Index Logistic Analysis of Variation

|  | $\begin{array}{c}\text { Degrees of } \\ \text { Deviance } \\ \text { Sreedom } \\ \text { SF) }\end{array}$ |  |  |  | $\begin{array}{c}\text { Mean } \\ \text { Deviance } \\ \text { (Dev.DF) }\end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | \(\left.\begin{array}{c}F- <br>

Ratio\end{array} \quad $$
\begin{array}{c}\text { Type 1 } \\
\text { P }\end{array}
$$\right]\)
B.3.b. 2001 Untagged and Tagged Hatchery Mean Release-to-McNary Survival Indices

|  | Untagged | Tagged |
| :---: | :---: | :---: |
| Mean over common Julian Weeks (adjusted**) | 0.180 | 0.140 |
| Mean over common Julian Weeks (unadjusted**) | 0.177 | 0.162 |

*Variable is survial index, weight is number of released fish
** Respectively adjusted and unadjusted for common Julian Week effects
B.4.a. 2002 Tagged versus Untagged: Weighted* Survival Index Logistic Analysis of Variation

|  | $\begin{array}{c}\text { Degrees of } \\ \text { Source }\end{array}$ |  |  |  | $\begin{array}{c}\text { Mean } \\ \text { Deviance } \\ \text { (Dev) }\end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| (DF) |  |  |  |  |  | \(\left.\begin{array}{c}Deviance <br>

(Dev.DF)\end{array} \quad $$
\begin{array}{c}\text { F- } \\
\text { Ratio }\end{array}
$$ $$
\begin{array}{c}\text { Type 1 } \\
\text { P }\end{array}
$$\right]\)
B.4.b. 2002 Untagged and Tagged Hatchery Mean Release-to-McNary Survival Indices

|  | Untagged | Tagged |
| :---: | :---: | :---: |
| Mean over common Julian Weeks (adjusted*) | 0.276 | 0.210 |
| Mean over common Julian Weeks (unadjusted*) | 0.286 | 0.223 |

[^6]
## APPENDIX C

2002 Annual Report OCT-SNT Survival

# IntSTATS 

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# 2002 Annual Report <br> OCT-SNT Survival 

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Submitted April 1, 2003

## 1. Summary

Smolt-Smolt Survival: The release-to-McNary-Dam survivals of brood-year 2000 (2002-outmigrant) PIT-tagged smolt reared under the semi-natural treatment (SNT) were uniformly less than those reared under the optimal conventional treatment (OCT). There was also a significant difference between the two treatments' effects on pre-release survival as measured by the released-to-tagged-number ratio. There were no significant differences in release-to-McNary survival between the SNT and OCT fish in any of the previous brood years (1997 through 1999). The 2000 brood-year survivals were highly and negatively correlated with mean BKD severity measures. When adjusted for these BKD measure means, there were no significant SNT-OCT differences in either the release-to-McNary survival indices or the released-to-tagged-number ratios. It is likely that SNT rearing conditions increased the severity of BKD and increased the impact of the disease on survival in PIT-tagged fish.

Smolt-to-Adult Survival: For the 1997 brood based on combined Age-3, Age-4, and Age-5 PIT-tagged returns (return years 2000, 2001, and 2002, respectively), there were no significant differences between the OCT and SNT effects on the survival from juvenile-tagging ${ }^{1}$-to-adult passage at Roza Dam on the Upper Yakima River. For the 1998 brood there were no significant differences between the SNT and OCT effects on juvenile-release ${ }^{1}$-to-Roza-return sur vival at the Clark Flats and Easton release sites based on combined Age-3 and Age-4 returns (return- years 2001 and 2002, respectively); however, SNT's survival was significantly less than the OCT's at the Jack Creek Site. The Jack Creek site was not available to the 1997 brood.

The 1998-brood results should be regarded as tentative because age- 5 adults are not included. In last year's report, when only age-4 returns were analyzed for the brood-year 1997 analysis, there was some statistical evidence that SNT fish had higher survival than

[^7]OCT fish; however, as indicated above, when all age groups were included, the analysis indicated no significant difference between the treatment effects.

## 2. Brood-2000 OCT-SNT Release-to-McNary-Dam Smolt-to-Smolt Survival

Table 1.a. presents the estimated SNT and OCT release-to-McNary survival-index proportions ${ }^{2}$ for each acclimation site (Clark Flats, Jack Creek, and Easton) as well as treatment main-effect estimates and site main-effect estimates for brood- year 2000 (2002 outmigrant) PIT-tagged smolt. Table 1.a. also presents the ratio of the number of detected volitionally released fish to the number of fish PIT-tagged (release-to-taggednumber ratio) expressed as a proportion. Table 1.b.1) presents the weighted logistic analysis of variation ${ }^{3}$ of the release-to-McNary survival indices as a measure of postrelease smolt-to-smolt survival, and Table 1.b.2) presents a comparable analysis for the released-to-tagged-number ratio as a measure of pre-release survival. There was a significant difference between the release-to-McNary survival indices of SNT- and OCTreared fish $[\mathrm{P}=0.045$ from Table 1.b.1)] and a significant difference between the SNT and OCT released/tagged-number ratios $[\mathrm{P}=0.001$ from Table 1.b.2)].

The underlying hypothesis to be tested was that the survival index of the SNT-reared fish exceeded that of the OCT-reared fish relative to the hypothesis that the treatments' survival indices did not differ. However, what is clear is that the SNT measures of survival are significantly less than those of the OCT. There were no significant treatment-effect differences in smolt-to-smolt survival in previous brood years (brood years 1997 through $1999^{4}$ ). Figure 1, which plots SNT and OCT release-to-McNary survival indices for all sites within each brood year, graphically contrasts the OCT and SNT differences for brood-year 2000 and those of previous brood years.

One possibility for the treatment difference in brood-year 2000 is that there were greater disease symptoms associated with rearing conditions under the SNT than under the OCT. Several diseases were monitored, and one found to be present was Bacterial Kidney Disease (BKD). The survival indices were reanalyzed adjusting the survival indices for BKD averages of severity measures taken from sampled sacrificed juveniles from each pond ${ }^{5}$. This analysis was a logistic analysis of covariation using the perraceway BKD severity mean as a concomitant variable or covariate. The covariateadjusted analysis revealed no significant difference between the SNT and OCT BKD-

[^8]adjusted smolt-to-smolt treatment survival indices [ $\mathrm{P}=0.616$ for BKD-adjusted analysis from Table 1.c.1) compared to $\mathrm{P}=0.045$ for the BKD -unadjusted analysis from Table 1.b.1)]. The same was true for the released-to-tagged-number ratio $[P=0.654$ for BKDadjusted analysis from Table 1.c.2) compared to $\mathrm{P}=0.001$ for BKD - unadjusted analysis from Table 1.b.2)].

Individual raceway BKD-severity measure means are given in Table 1.d. along with the release-to-McNary survival indices and released-to-tagged-number ratios. The raceway pairs presented in the tables are blocks. The SNT and OCT raceways within the pairs are adjacent to each other, but, more importantly, the fish from the OCT and SNT raceways within the same block share the same sets of parental crosses (different blocks have fish from different sets of parental crosses). It can be seen from Table 1.d. that 1) the SNT survival index was lower than that of the OCT within all nine raceway pairs, 2) the SNT released-to-tagged-number ratio was lower than that of the OCT within seven of the nine raceway pairs ${ }^{6}$, and 3) the BKD measure mean of the SNT exceeded that of the OCT within eight of the nine raceway pairs. This disease difference could not be because of parentalsource differences in BKD levels because the pairs shared the same parental crosses ${ }^{7}$. The disease differences can most likely be attributed to a rearing difference between SNT and OCT treatments. It is also likely that these disease differences contributed to the differences in the survival indices. Table 1.e. gives the relative distribution of severity measures for the sampled fish within each acclimation pond.

Table 1.a. 2000-Brood (2002-Outmigrant) SNT and OCT Release-to-McNaryDam Survival Indices and Tagged-to-Release-Number Ratio Estimates within Sites and Treatment and Site Main Effect Estimates

| Site | Treatment | Release-McNary Survival Index |  |  |  | Released-to-Tagged Numbers Ratio |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Means <br> Release <br> Number | in Sites Survival Index | Site M <br> Release <br> Number | Effect Survival Index | Means <br> Tagged <br> Numbe | in Sites <br> Ratio | Site Main Effect Tagged Number Ratio |  |
| Clark Flats | $\begin{aligned} & \hline \text { OCT } \\ & \text { SNT } \end{aligned}$ | $\begin{aligned} & 6340 \\ & 5858 \end{aligned}$ | $\begin{aligned} & 0.4093 \\ & 0.2930 \end{aligned}$ | 12198 | 0.3535 | $\begin{aligned} & 6677 \\ & 6675 \end{aligned}$ | $\begin{aligned} & \hline 0.9495 \\ & 0.8776 \end{aligned}$ | 13352 | 0.9136 |
| Jack Creek | $\begin{aligned} & \mathrm{OCT} \\ & \mathrm{SNT} . \end{aligned}$ | $\begin{array}{r} 6480 \\ 6466 \end{array}$ | $\begin{aligned} & 0.3654 \\ & 0.2971 \end{aligned}$ | 12946 | 0.3313 | $\begin{array}{r} 6675 \\ 6678 \end{array}$ | $\begin{aligned} & 0.9708 \\ & 0.9683 \end{aligned}$ | 13353 | 0.9695 |
| Easton | $\begin{aligned} & \mathrm{OCT} \\ & \mathrm{SNT} \end{aligned}$ | $\begin{array}{r} 6512 \\ 5924 \\ \hline \end{array}$ | $\begin{aligned} & 0.3167 \\ & 0.1837 \end{aligned}$ | 12436 | 0.2533 | $\begin{array}{r} 6677 \\ 6675 \\ \hline \end{array}$ | $\begin{aligned} & 0.9753 \\ & 0.8875 \\ & \hline \end{aligned}$ | 13352 | 0.9314 |
| Treatment Main Effect | OCT SNT | 19332 18248 | 0.3634 0.2590 | 37580 | 0.3127 | $\begin{aligned} & 20029 \\ & 20028 \end{aligned}$ | $\begin{aligned} & \hline 0.9652 \\ & 0.9111 \end{aligned}$ | 40057 | 0.9382 |

[^9]Table 1.b.1) Weighted Logistic Analysis of Variation of 2000-Brood (2002Outmigrant) Release-to-McNary-Dam Survival Indices (weight is number released)

| Source | Deviance (Dev) | Degrees of Freedom (DF) | Mean Deviance (Dev/DF) | F-Ratio* | $\begin{gathered} \text { Type } 1 \\ \text { P } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Site | 336.74 | 2 | 168.37 | 0.77 | 0.5025 |
| Block (Raceway Pair) | 1306.02 | 6 | 217.67 | 2.52 | 0.1431 |
| Treatment (OCT versus SNT) | 549.22 | 1 | 549.22 | 6.35 | 0.0453 |
| Site x Treatment Interaction | 64.59 | 2 | 32.30 | 0.37 | 0.7034 |
| Error | 519.09 | 6 | 86.52 |  |  |

* Site tested against block; block, treatment, and interaction tested against error

Table 1.b.2) Weighted Logistic Analysis of Variation of 2000-Brood (2002Outmigrant) Released-to-Tagged-Number Ratio (weight is number tagged)

| Source | Deviance <br> $($ Dev $)$ | Degrees of <br> Freedom <br> (DF) | Mean <br> Deviance <br> (Dev/DF) | F-Ratio* | Type 1 <br> P |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Site | 415.33 | 2 | 207.67 | 0.87 | 0.4652 |
| Block (Raceway Pair) | 1429.1 | 6 | 238.18 | 13.80 | 0.0028 |
| Treatment | 551.62 | 1 | 551.62 | 31.95 | 0.0013 |
| Site $\times$ Treatment Interaction | 138.04 | 2 | 69.02 | 4.00 | 0.0788 |
| Error | 103.59 | 6 | 17.27 |  |  |

* Site tested against block; block, treatment, and interaction tested against error

Figure 1. Outmigrant SNT and OCT Treatment Release-to-McNary-Dam Survival Indices within Sites for Brood Years 1997 through 2000 (1999 through 2002 Outmigrants)


CF-Clark Flats, ES-Easton, JC-Jack Creek
SNT $\square$ OCT

Table 1.c.1) Weighted Logistic Analysis of Covariation of 2000-Brood (2002Outmigrant) Release-to-McNary-Dam Survival Indices adjusted for Mean BKD Severity Measure as Covariate (weight is number released) ${ }^{8}$

| Source | Deviance (Dev) | Degrees of Freedom (DF) | Mean Deviance (Dev/DF) | F-Ratio* | $\begin{gathered} \text { Type } 1 \\ \text { P } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Site | 10.83 | 2 | 5.42 | 0.06 | 0.9440 |
| Block (Raceway Pair) | 558.66 | 6 | 93.11 | 1.65 | 0.2995 |
| Treatment (OCT versus SNT) | 16.11 | 1 | 16.11 | 0.29 | 0.6160 |
| Site $\times$ Treatment | 80.86 | 2 | 40.43 | 0.72 | 0.5326 |
| Error | 282.12 | 5 | 56.42 |  |  |

* Site tested against block; block, treatment, and interaction tested against error

Table 1.c.2) Weighted Logistic Analysis of Covariation of 2000-Brood (2002Outmigrant) Released-to-Tagged-Number Ratio adjusted for Mean BKD Severity Measure as Covariate (weight is number tagged) ${ }^{9}$

|  | Deviance <br> $($ Dev $)$ | Degrees of <br> Freedom <br> $($ DF $)$ | Mean <br> Deviance <br> (Dev/DF) | F-Ratio | Type 1 <br> Source |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Site | 792 | 2 | 396.00 | 6.35 | 0.0330 |
| Block (Raceway Pair) | 374.27 | 6 | 62.38 | 3.59 | 0.0911 |
| Treatment | 3.93 | 1 | 3.93 | 0.23 | 0.6545 |
| Site $\times$ Treatment | 30.42 | 2 | 15.21 | 0.88 | 0.4722 |
| Error | 86.9 | 5 | 17.38 |  |  |

* Site tested against block; block, treatment, and interaction tested against error

[^10]Table 1.d. Number of Spring Chinook Tagged and Released, Release-to-McNary Survival Index and Released-to-Tagged-Number Ratio, and Mean BKD Severity Measure ${ }^{10}$ by Raceway for 2000 Brood (CF--Clark Flats, JC--Jack Creek, ES--Easton)

| Raceway Pair | Acclimation Pond Raceway* | Treatment** | Number <br> Tagged | Overall <br> Number <br> Released | Overall Survival Index | Released Tagged Ratio | Average BKD Index |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | CF-01 | SNT | 2225 | 1618 | 0.0456 | 0.7272 | 2.1695 |
| 1 | CF-02 | OCT | 2226 | 1953 | 0.2802 | 0.8774 | 1.3333 |
| 2 | CF-03 | SNT | 2225 | 2094 | 0.3124 | 0.9411 | 1.5000 |
| 2 | CF-04 | OCT | 2225 | 2186 | 0.4400 | 0.9825 | 1.1833 |
| 3 | CF-05 | SNT | 2225 | 2146 | 0.4606 | 0.9645 | 1.2034 |
| 3 | CF-06 | OCT | 2226 | 2201 | 0.4935 | 0.9888 | 1.0169 |
| 7 | JC-01 | SNT | 2225 | 2179 | 0.3046 | 0.9793 | 1.9333 |
| 7 | JC-02 | OCT | 2225 | 2137 | 0.3084 | 0.9604 | 1.3000 |
| 8 | JC-03 | SNT | 2226 | 2185 | 0.3619 | 0.9816 | 1.4500 |
| 8 | JC-04 | OCT | 2225 | 2182 | 0.4346 | 0.9807 | 1.5000 |
| 9 | JC-05 | SNT | 2227 | 2102 | 0.2220 | 0.9439 | 1.5500 |
| 9 | JC-06 | OCT | 2225 | 2161 | 0.3518 | 0.9712 | 1.2459 |
| 4 | ES-01 | SNT | 2225 | 2161 | 0.3229 | 0.9712 | 1.9667 |
| 4 | ES-02 | OCT | 2226 | 2185 | 0.3556 | 0.9816 | 1.1667 |
| 5 | ES-03 | SNT | 2225 | 2026 | 0.1406 | 0.9106 | 2.4667 |
| 5 | ES-04 | OCT | 2225 | 2160 | 0.2796 | 0.9708 | 1.3500 |
| 6 | ES-05 | SNT | 2225 | 1737 | 0.0607 | 0.7807 | 3.1167 |
| 6 | ES-06 | OCT | 2226 | 2167 | 0.3143 | 0.9735 | 1.3000 |

[^11]Table 1.e. Relative Distribution over BKD Severity Measure (Rank) of Sampled Fish within Raceway as well as Number of Fish Sampled and Sample Mean per Raceway for 2000-Brood (2002-Outmigrant) OCT-SNT Spring Chinook. [Data provided by Ray Brunson (United States Fish and Wildlife Service, Olympia, Washington.]


| Site $>$Treatment $\gg$Acclimation Pond $\gg$ |  | Jack Creek (JC) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | SNT |  |  | OCT |  |  |
|  |  | JC-1 | JC-3 | JC-5 | JC-2 | JC-4 | JC-6 |
| Risk* | Rank |  |  |  |  |  |  |
| ND | 0 | 0.00000 | 0.16667 | 0.10000 | 0.13333 | 0.08333 | 0.09836 |
| VL | 1 | 0.41667 | 0.36667 | 0.45000 | 0.43333 | 0.45000 | 0.55738 |
|  | 2 | 0.45000 | 0.43333 | 0.36667 | 0.43333 | 0.40000 | 0.34426 |
|  | 3 | 0.06667 | 0.00000 | 0.03333 | 0.00000 | 0.03333 | 0.00000 |
| L | 4 | 0.01667 | 0.00000 | 0.01667 | 0.00000 | 0.01667 | 0.00000 |
|  | 5 | 0.00000 | 0.00000 | 0.01667 | 0.00000 | 0.01667 | 0.00000 |
| M | 6 | 0.00000 | 0.01667 | 0.00000 | 0.00000 | 0.00000 | 0.00000 |
| H | 7 | 0.05000 | 0.01667 | 0.01667 | 0.00000 | 0.00000 | 0.00000 |
| VH | 8 | 0.00000 | 0.00000 | 0.00000 | 0.00000 | 0.00000 | 0.00000 |
|  | 9 | 0.00000 | 0.00000 | 0.00000 | 0.00000 | 0.00000 | 0.00000 |
| Total Sampled > |  | 60 | 60 | 60 | 60 | 60 | 61 |
| Mean Severity Rank** |  | 1.933 | 1.450 | 1.550 | 1.300 | 1.500 | 1.246 |



[^12]The differences in the survival indices tended to persist throughout the volitional release period. For the Clark Flats, Jack Creek, Easton sites respectively, Figures 2.a through 2.c present individual raceway survival indices for six periods of volitional release:

1) before March $25^{11}$,
2) March 25 through April 3,
3) April 4 through April 13,
4) April 14 though April 23,
5) April 24 through May 3, and
6) after May 3 .

The general trend observable from these figures is that the SNT release-to-McNary survival index tends to be less than for the OCT for most raceway pairs over the volitional release period. (Data used to produce Figure 2 are given in Appendix B.)

Figure 2. Brood-Year 2000 (2002-Outmigrant) SNT and OCT Treatment Release-to-McNary-Dam Survival Indices for Different Periods of Volitional Release


[^13]Figure 2. Brood-Year 2000 (2002-Outmigrant) SNT and OCT Treatment
Release-to-McNary-Dam Survival Indices for Different Periods of Volitional Release

c. Easton


## 3. Brood-Year 1997 and 1998 OCT-SNT Smolt Release-to-Adult-Return Dam Survival

Release-to-Adult survival estimates based on PIT-tagged fish for the 1997 brood have been revised to include age- 5 returns in addition to the age- 3 and age -4 returns which were used to provide the estimates given in the 2001 annual report. Estimates presented here for the 1998 brood are based on age- 3 and age- 4 returns and should be regarded as incomplete because age- 5 returns are not yet available.

## 3.a. Brood-Year 1997 Release-to-Adult Survival (age-3 through age-5 returns)

Brood-year 1997 release-to-adult survival indices were computed on a raceway basis by dividing the number of PIT-tagged adults detected at Roza by the number of juveniles that were originally tagged. In last year's annual report, there was an indication of a higher survival associated with the SNT treatment. The logistic analysis of variation's Fratio associated with the OCT versus SNT treatment comparison had an associated estimated Type 1 error probability of $\mathrm{P}=0.141$ when based on only age -4 returns. This was the equivalent of $\mathrm{P}=0.070$ for a one-sided test for the SNT survival exceeding the OCT survival. However, the survival estimate based on the combined age-3 (year 2000) returns, age-4 (year 2001) returns, and age-5 (year 2002) returns revealed no significant difference between the OCT and SNT treatments. Table 2.a.1) presents the OCT and SNT weighted ${ }^{12}$ means within each site as well as the treatment and site main effect means. Table 2.a.2) presents the associated logistic analysis of variation. Since the variation among blocks (among raceway pairs) was not significantly greater than that of error [Error (1) in Table 2.A.2)] at the $20 \%$ significance level $(\mathrm{P}=0.205)$, the block and error sources were pooled to give the source of error [Error (2) in Table 2.a.2)] that is used for the statistical tests. None of the sources were significant when tested against Error (2) ( $\mathrm{P}=0.518$ for Site, $\mathrm{P}=0.339$ for the Oct versus SNT Treatment comparison, and $\mathrm{P}=0.400$ for the Site x Treatment interaction). Individual raceway survival information is given in Table 2.a.3).

## Table 2.a.1) 1997-Brood OCT and SNT Juvenile-Tagging ${ }^{13}$-to-Adult-Return Survival to Roza Dam on the Upper Yakima

| Treatment |  | Site |  |  |
| :---: | ---: | :---: | :---: | :---: |
|  | Clark Flats | Jack Creek | Mean |  |
| SNT | Survival Index | 0.0171 | 0.0143 | 0.0160 |
|  | Number Tagged | 11879 | 7891 | 19770 |
|  | Survival Index | 0.0141 | 0.0145 | 0.0142 |
|  | Number Tagged | 11867 | 7933 | 19800 |
| Mean | Survival Index | 0.0156 | 0.0144 | 0.0151 |
|  | Number Tagged | 23746 | 15824 | 39570 |

[^14]Table 2.a.2) Logistic Analysis of Variation of 1997-Brood-Year Juvenile-Tagging-to-Adult-Return Survival to Roza Dam on the Upper Yakima

| Source | Deviance (Dev) | Degrees of Freedom (DF) | Mean <br> Deviance <br> (Dev/DF) | F-Ratio* | Type 1 <br> Error <br> P |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Site | 0.88 | 1 | 0.88 | 0.32 | 0.6122 |
| Block | 8.3 | 3 | 2.77 | 2.86 | 0.2054 |
| Treatment (Trt) | 2.01 | 1 | 2.01 | 2.08 | 0.2450 |
| Site $\times$ Trt | 1.53 | 1 | 1.53 | 1.58 | 0.2974 |
| Error (1) | 2.9 | 3 | 0.97 |  | 0.4073 |
| Block and Error (1) Pooled serving as base for statitsical tests below |  |  |  |  |  |
|  | Dev | DF | Dev/DF | F-Ratio** | P |
| Site | 0.88 | 1 | 0.88 | 0.47 | 0.5180 |
| Trt | 2.01 | 1 | 2.01 | 1.08 | 0.3394 |
| Site $\times$ Trt | 1.53 | 1 | 1.53 | 0.82 | 0.4002 |
| Error(2) | 11.2 | 6 | 1.87 |  |  |

* Site initially tested against Block source; Block, Treatment and Interaction tested against Error(1)
** Site, Treatment and Interaction tested against Error(2)
Table 2.a.3) Individual raceway survival information for the 1998-brood SNT and OCT releases.

| Site | $\begin{gathered} \text { Raceway } \\ \text { Pair } \\ \hline \end{gathered}$ | Treatment | Returns |  |  |  | Number <br> Released |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Age 3 | Age 4 | Age 5 | Total to Date |  |
| Easton | 1 | SNT | 11 | 58 | 1 | 70 | 3958 |
| Easton | 1 | OCT | 9 | 44 | 6 | 59 | 3969 |
| Easton | 2 | SNT | 2 | 40 | 1 | 43 | 3933 |
| Easton | 2 | OCT | 0 | 55 | 1 | 56 | 3964 |
| Clark Flats | 3 | SNT | 5 | 53 | 0 | 58 | 3936 |
| Clark Flats | 3 | OCT | 4 | 41 | 1 | 46 | 3929 |
| Clark Flats | 4 | SNT | 5 | 70 | 0 | 75 | 3968 |
| Clark Flats | 4 | OCT | 10 | 47 | 2 | 59 | 3972 |
| Clark Flats | 5 | SNT | 9 | 59 | 2 | 70 | 3975 |
| Clark Flats | 5 | OCT | 10 | 50 | 2 | 62 | 3966 |
|  |  |  | Proportion |  |  |  |  |
| Site | Raceway | Treatment | Age 3 | Age 4 | Age 5 | Total |  |
| Easton | 1 | SNT | 0.002779 | 0.014654 | 0.000253 | 0.017686 |  |
| Easton | 1 | OCT | 0.002268 | 0.011086 | 0.001512 | 0.014865 |  |
| Easton | 2 | SNT | 0.000509 | 0.010170 | 0.000254 | 0.010933 |  |
| Easton | 2 | OCT | 0.000000 | 0.013875 | 0.000252 | 0.014127 |  |
| Clark Flats | 3 | SNT | 0.001270 | 0.013465 | 0.000000 | 0.014736 |  |
| Clark Flats | 3 | OCT | 0.001018 | 0.010435 | 0.000255 | 0.011708 |  |
| Clark Flats | 4 | SNT | 0.001260 | 0.017641 | 0.000000 | 0.018901 |  |
| Clark Flats | 4 | OCT | 0.002518 | 0.011833 | 0.000504 | 0.014854 |  |
| Clark Flats | 5 | SNT | 0.002264 | 0.014843 | 0.000503 | 0.017610 |  |
| Clark Flats | 5 | OCT | 0.002521 | 0.012607 | 0.000504 | 0.015633 |  |

## 3.b. Brood-Year 1998 Release-to-Adult Survival (age-3 through age-4 returns)

Brood-year 1998 release-to-adult survival indices were computed on a raceway basis by dividing the number of PIT-tagged adults detected at Roza by the number of juveniles detected leaving the raceway. The survival index based on the combined age-3 (year 2001) returns and age-4 (year 2002) returns revealed evidence of an interaction of the OCT and SNT comparisons with sites. Table 2.b.1) presents the OCT and SNT weighted ${ }^{14}$ means. Table 2.b.2) presents the associated logistic analysis. Referring to Table 2.b.2) a), as with the 1997 brood, the Block and Error (1) source were pooled to produce Error (2) because the F-test of Block against Error (1) was not significant at the $20 \%$ level ( $\mathrm{P}=0.451$ ). The OCT versus SNT Treatment x Site Interaction was nearly significant at the $5 \%$ level ( $\mathrm{P}=0.051$ ); therefore comparisons were made between the OCT and SNT treatments within the individual sites. Referring to Table 2.b.2) b), the treatments did not significantly differ within Clark Flats and Easton $(\mathrm{P}=0.431$ and $\mathrm{P}=$ 0.207 , respectively) but did significantly differ within Jack Creek $(\mathrm{P}=0.029)$ with the survival of the SNT fish being less than that of the OCT fish.

The significant reduction in the SNT survival relative to OCT is driven primarily by the results for raceway pair 8 which are boldfaced in Table 2.b.3). The number of fish released from the OCT raceway is less than half of that in the SNT (1109 for OCT compared to 2279 for SNT) because of a major loss of the raceway-pair-8 OCT fish when transferred from Cle Elum to Jack Creek. The returns, however, are slightly higher for the OCT ( 14 for the OCT and 13 for the SNT). The result for raceway pair 8 is that the adult-return survival for the OCT is more than twice that of the SNT ( 0.0126 compared to 0.0057). The OCT adult survival from Jack Creek's other raceway pair (raceway pair 7) is also greater than that of SNT, but the magnitude of the difference is much less than that of raceway pair 8 (survival were 0.0118 for OCT and 0.0085 for SNT). Final determination of the effects of the treatments will require the inclusion of brood-year 1998's age-5 returns in 2003. Also presented in the Table 2.b.3) are the estimated smolt-to-smolt survival indices from release to McNary (2000 outmigrants). As can be seen in Table 2.b.3), the OCT and SNT juvenile survival indices to McNary for raceways 7 and 8 indicate little difference between the Jack Creek OCT and SNT estimates, although the direction of the juvenile survival is the same as that for the release-to-adult survival proportion (OCT's survival index > SNT's).

[^15]Table 2.b.1) 1998-Brood OCT and SNT Juvenile-Release ${ }^{15}$-to-Adult-Return Survival to Roza Dam on the Upper Yakima

| Treatment |  | Site |  |  |  |
| :---: | ---: | :---: | :---: | :---: | :---: |
|  |  | Clark Flats | Jack Creek | Easton | Mean |
| SNT | Survival Index | 0.0120 | 0.0071 | 0.0095 | 0.0099 |
|  | Release Number | 7253 | 4619 | 7151 | 19023 |
| OCT | Survival Index | 0.0107 | 0.0121 | 0.0076 | 0.0098 |
|  | Retelease Number | 7287 | 3562 | 7192 | 18041 |
| Mean | Survival Index | 0.0113 | 0.0093 | 0.0086 | 0.0098 |
|  | Release Number | 14540 | 8181 | 14343 | 37064 |

Table 2.b.2) Logistic Analysis of 1998-Brood Juvenile-Release-to-Adult-Return Survival to Roza Dam on the Upper Yakima
a) Analysis of Variation

| Source | $\begin{gathered} \text { Deviance } \\ \text { (Dev) } \\ \hline \end{gathered}$ | Degrees of Freedom (DF) | Mean Deviance (Dev/DF) | F-Ratio* | $\begin{gathered} \hline \text { Type } 1 \\ \text { Error } \\ \mathbf{P} \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Site | 5.96 | 2 | 2.98 | 3.52 | 0.1110 |
| -memembiock | 4.23 | 5 | 0.85 | 1.12 | 0.4513 |
| Treatment (SNT vs, OCT) | 0.06 | 1 | 0.06 | 0.08 | 0.7892 |
| Site x Treatment | 6.47 | 2 | 3.24 | 4.29 | 0.0822 |
| Error (1) | 3.77 | 5 | 0.75 |  | 0.5830 |
| Block and Error (1) Pooled serving as base for statitsical tests below |  |  |  |  |  |
|  | Dev | DF | Dev/DF | F-Ratio** | P |
| Site | 5.96 | 2 | 2.98 | 3.73 | 0.0618 |
| Treatment | 0.06 | 2 | 0.03 | 0.08 | 0.9283 |
| Site x Treatment | 6.47 | 2 | 3.24 | 4.04 | 0.0517 |
| Error(2) | 8 | 10 | 0.80 |  | 0.6288 |

* Site initially tested against Block source; Block, Treatment and Interaction tested against Error(1)
** Site, Treatment and Interaction tested against Error(2)

[^16]Table 2.b.2) Logistic Analysis of 1998-Brood Juvenile-Release-to-Adult-Return Survival to Roza Dam on the Upper Yakima (continued)
b) Mean Comparisons

| Treatment |  | Site |  |  |  |
| :---: | ---: | :---: | :---: | :---: | :---: |
|  |  | Clark Flats | Jack Creek | Easton | Mean |
| SNT | Survival Index | 0.0120 | 0.0071 | 0.0095 | 0.0099 |
|  | Logit Transform | -4.4112 | -4.9343 | -4.6459 |  |
|  | SE(Logit) | 0.0965 | 0.1562 | 0.1090 |  |
|  | Release Number | 7253 | 4619 | 7151 | 19023 |
| Survival Index | 0.0107 | 0.0121 | 0.0076 | 0.0098 |  |
| OCT | Logit Transform | -4.5264 | -4.4047 | -4.8657 |  |
|  | SE(Logit) | 0.1018 | 0.1372 | 0.1211 |  |
|  | Release Number | 7287 | 3562 | 7192 | 18041 |
| Survival Index | 0.0113 | 0.0093 | 0.0086 | 0.0098 |  |
|  | Rean | 14540 | 8181 | 14343 | 37064 |
| SNT versus OCT |  |  |  |  |  |
|  | Difference | 0.001291 | -0.004927 | 0.001862 |  |
|  | Logit Difference | 0.1152 | -0.5295 | 0.2198 |  |
|  | SE(Logit Difference) | 0.1403 | 0.2079 | 0.1629 |  |
|  | t-ratio | 0.82 | -2.55 | 1.35 |  |
|  | Type 1 P | 0.4307 | 0.0290 | 0.2070 |  |
|  |  |  |  |  |  |

Table 2.b.3) Individual raceway survival information for the 1998-Brood SNT and OCT releases.

| Site | Raceway Pair | Treatment | Adult Returns |  |  | Number <br> Released |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Age 3 | Age 4 | Total to Date |  |
| Easton | 1 | SNT | 3 | 11 | 14 | 2343 |
| Easton | 1 | OCT | 5 | 13 | 18 | 2404 |
| Easton | 2 | SNT | 7 | 19 | 26 | 2392 |
| Easton | 2 | OCT | 1 | 18 | 19 | 2349 |
| Easton | 3 | SNT | 1 | 27 | 28 | 2416 |
| Easton | 3 | OCT | 0 | 18 | 18 | 2439 |
| Clark Flats | 4 | SNT | 3 | 27 | 30 | 2426 |
| Clark Flats | 4 | OCT | 0 | 23 | 23 | 2453 |
| Clark Flats | 5 | SNT | 5 | 23 | 28 | 2429 |
| Clark Flats | 5 | OCT | 5 | 23 | 28 | 2423 |
| Clark Flats | 6 | SNT | 7 | 22 | 29 | 2398 |
| Clark Flats | 6 | OCT |  | 26 | 27 | 2411 |
| Jack Creek | 7 | SNT | 5 | 15 | 20 | 2340 |
| Jack Creek | 7 | OCT | 2 | 27 | 29 | 2453 |
| Jack Creek | 8 | SNT | 3 | 10 | 13 | 2279 |
| Jack Creek | 8 | OCT | 6 | 8 | 14 | 1109 |
|  |  |  |  | ult Proportio |  | Juvenile |
| Site | Raceway Pair | Treatment | Age 3 | Age 4 | Total to Date | Survival to McNary |
| Easton | 1 | SNT | 0.001280 | 0.004695 | 0.005975 | 0.3969 |
| Easton | 1 | OCT | 0.002080 | 0.005408 | 0.007488 | 0.3663 |
| Easton | 2 | SNT | 0.002926 | 0.007943 | 0.010870 | 0.3921 |
| Easton | 2 | OCT | 0.000426 | 0.007663 | 0.008089 | 0.4190 |
| Easton | 3 | SNT | 0.000414 | 0.011175 | 0.011589 | 0.3740 |
| Easton | 3 | OCT | 0.000000 | 0.007380 | 0.007380 | 0.3595 |
| Clark Flats | 4 | SNT | 0.001237 | 0.011129 | 0.012366 | 0.2832 |
| Clark Flats | 4 | OCT | 0.000000 | 0.009376 | 0.009376 | 0.3063 |
| Clark Flats | 5 | SNT | 0.002058 | 0.009469 | 0.011527 | 0.3367 |
| Clark Flats | 5 | OCT | 0.002064 | 0.009492 | 0.011556 | 0.3355 |
| Clark Flats | 6 | SNT | 0.002919 | 0.009174 | 0.012093 | 0.3209 |
| Clark Flats | 6 | OCT | 0.000415 | 0.010784 | 0.011199 | 0.2956 |
| Jack Creek | 7 | SNT | 0.002137 | 0.006410 | 0.008547 | 0.3600 |
| Jack Creek | 7 | OCT | 0.000815 | 0.011007 | 0.011822 | 0.3879 |
| Jack Creek | 8 | SNT | 0.001316 | 0.004388 | 0.005704 | 0.3390 |
| Jack Creek | 8 | OCT | 0.005410 | 0.007214 | 0.012624 | 0.3442 |

## 3.c. Consistency between release-to-adult survival and juvenile release-toMcNary survival

For the 1997-brood adult returns, Clark Flats release survival exceeded that of Easton. For the 1998-brood adult returns, Clark Flats (CF) survival exceeded that of Jack Creek ${ }^{16}$ (JC), and Jack Creek survival exceeded that of Easton (ES). The same relations held for smolt-to-smolt survival indices (release-to-McNary passage survival). Referring to Figure 3 for all brood years to date (1997, 1998, 1999, 2000), the relative release-toMcNary smolt survival-index estimates over the sites have been the same:

For the 1997-brood release-to-McNary smolt survival index, neither the treatment nor the treatment x site interactions were significant [respectively $\mathrm{P}=0.811$ and $\mathrm{P}=0.873$, refer to Table 2.d.1)]. This was also the case for the tagged-to-adult-return analysis [respectively $\mathrm{P}=0.339$ and $\mathrm{P}=0.400$, refer back to Table 2.a.2)]. However, for the 1998 brood, there was not a consistency in the degrees of significance between the smoltsurvival and the adult-survival analyses. Neither the treatment nor the treatment x site interactions were significant for the smolt survival index analysis [respectively $\mathrm{P}=0.713$ and $\mathrm{P}=0.673$, refer to Table 2.d.2)], but recall that, for the adult analysis, the site x treatment interaction was nearly significant at the 5\% level [Site x Treatment interaction $\mathrm{P}=0.052$, refer back to Table 2.b.2)a)]. Even so, the direction of OCT and SNT smolt-to-smolt survival-index differences was the same as that of the smolt-to-adult survival for the 1998 brood: the SNT survival index being less than that of the OCT for Jack Creek but not for Clark Flats or Easton (refer to Figure 4); however, the difference for Jack Creek is far more dramatic for smolt-to-adult survival than for smolt-smolt survival.

Figure 3. Smolt Survival to McNary Dam for Brood Years 1997 through 2000


[^17]Figure 4. Relative 1998-Brood SNT and OCT Smolt Survivals (release-toMcNary Dam) and Adult Survivals (Release to Roza Dam Return-Age 3 and 4)


Table 2.d.1) Weighted Logistic Analysis of Variation of OCT and SNT Smolt-toSmolt Survival Indices over Sites for 1997-Brood Outmigrants (Weight $=$ Number of PIT-Tagged Fish Released)

| Source | Deviance (Dev) | Degrees of Freedom (DF) | Mean Deviance (Dev/DF) | F-Ratio | Type 1 Error P |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Site ${ }^{1}$ | 30.86 | I | 30.86 | 1.89 | 0.2628 |
| Block $^{2}$ | 48.96 | 3 | 16.32 | 1.29 | 0.4187 |
| Treatment (Trt) ${ }^{2}$ | 0.90 | 1 | 0.90 | 0.07 | 0.8067 |
| Site x Trt Interaction ${ }^{2}$ | 0.40 | 1 | 0.40 | 0.03 | 0.8700 |
| Error(1) | 37.84 | 3 | 12.61 |  |  |
| Site ${ }^{3}$ | 30.86 | 1 | 30.86 | 2.13 | 0.1944 |
| Trt ${ }^{3}$ | 0.90 | 1 | 0.90 | 0.06 | 0.8114 |
| Site $\times$ Trt Interaction ${ }^{3}$ | 0.40 | 1 | 0.40 | 0.03 | 0.8734 |
| Error(2) ${ }^{4}$ | 86.80 | 6 | 14.47 |  |  |

${ }^{1}$ Site is initially tested against Block
${ }^{2}$ Block, Treatment, Ineraction initially tested against Error(1)
${ }^{3}$ Block, Treatment, Ineraction finally tested against Error(2)
${ }^{4}$ Error (2) is pooling of Error(1) and Block

Table 2.d.2) Weighted Logistic Analysis of Variation of OCT and SNT Smolt-toSmolt Survival Indices over Sites for 1998-Brood Outmigrants (Weight $=$ Number of PIT-Tagged FishReleased $)$

| Source | Deviance (Dev) | Degrees o Freedom (DF) | Mean Deviance (Dev/DF) | F-Ratio | Type 1 Error P |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Site ${ }^{1}$ | 137.31 | 2 | 68.66 | 7.18 | 0.0339 |
| Block ${ }^{2}$ | 47.82 | 5 | 9.56 | 9.56 | 0.2405 |
| Treatment (Trt) ${ }^{2}$ | 1.00 | 1 | 1.00 | 0.23 | 0.6539 |
| Site $\times$ Trt Interaction ${ }^{2}$ | 5.76 | 2 | 2.88 | 0.65 | 0.5595 |
| Error(1) | 22.03 | 5 | 4.41 |  |  |
| Site ${ }^{3}$ | 137.31 | 2 | 68.66 | 9.83 | 0.0044 |
| Trt ${ }^{3}$ | 1.00 | 1 | 1.00 | 0.14 | 0.7131 |
| Site $\times$ Trt Interaction ${ }^{3}$ | 5.76 | 2 | 2.88 | 0.41 | 0.6729 |
| Error(2) ${ }^{4}$ | 69.85 | 10 | 6.99 |  |  |

${ }^{1}$ Site is initially tested against Block
${ }^{2}$ Block, Treatment, Ineraction initially tested against Error(1)
${ }^{3}$ Block, Treatment, Ineraction finally tested against Error(2)
${ }^{4}$ Error (2) is pooling of $\operatorname{Error}(1)$ and Block

## Appendix A. Estimated Survival Index and Logistic Analysis

Weighted logistic analyses of variation of release-to-McNary survival index estimates were undertaken using release number as the weighting variable instead of a traditional least-squares-based analysis of variance ${ }^{17}$. Least squares analysis assumes that the variance of the estimates is constant over releases. In the case of survival proportions, this is not expected to be true. The assumption behind the logistic analysis of variation used is that the variance in survival is proportional to what would be expected in the case of a binomially distributed proportion:

$$
\text { Variance proportional to } \frac{S^{*}(1-S)}{n}
$$

wherein S is the expected proportion surviving for the release and n is the number of fish released. The variance of the survival estimate would change as the survival changed over releases, making the traditional analysis of variance inappropriate. Further, the number released varied over releases; this variation in $n$ in the above equation would also contribute to the variance of the survival proportion estimate changing over releases. For this reason, the release number was used as a weighting variable.

In the logistic analysis of variation, the comparison is effectively made among the estimated logit transforms of the survival index, the logit transform being

$$
y=\operatorname{logit}(s)=\text { natural } \log \left(\frac{s}{1-s}\right)
$$

$s$ being the estimated proportion surviving. The reverse transform, survival as a function of the logit, is

$$
s=\frac{1}{1+\exp (-y)}
$$

wherein $\exp (-y)$ is the exponential constant raised to the power given within the parentheses.

In running the analysis, site main effect, treatment main effect, and site x treatment measures of variation were computed as follows

Site Main Effect = Difference in
Regression on Site and Treatment Indicators and
Regression on Treatment Indicators

[^18]
## Treatment Main Effect = Difference in <br> Regression on Block (including Site) and Treatment Indicators and Regression on Block Indicators

Site x Treatment Indicator $=$ Difference between Regression on Block, and Treatment, and Interaction Indicators and Regression on Block, and Treatment Indicators

The release-to-McNary smolt-to-smolt survival index in this study is estimated as follows:

> Release - to - McNary Survival Index
> $=$


Number of PIT - Tagged Fish Released
wherein a stratum is a group of contiguous McNary detection dates among which the daily detection efficiencies ${ }^{18}$ are sufficiently homogeneous to permit the use of a pooled estimate of the detection efficiency for that stratum. The pooled estimate is a pooling of the daily detection efficiency estimates over all dates within the stratum.

Within a stratum, the detection efficiency is estimated as follows:

McNary detection efficiency
$=$
number of joint detections at McNary and downstream dam
estimated total number of detections at downstream dam
The method of pooling is given in Appendix B. The downstream-dam count actually represents a pooling of counts from John Day and Bonneville dams ${ }^{19}$. A major reason for

[^19]referring to the survival measure as a survival index instead of survival is that there are known biases associated estimate, which are discussed in Appendix B.

The release-to-McNary smolt-to-smolt survival-index estimates for the different volitional release periods as well as the estimates for the whole volitional release are given in Table A. 1 for each acclimation pond; this is the data-base summary used to generate Figure 2 in the text as well a data-base used for the analysis summaries for release-to-McNary smolt survival indices in the text. The number of detections for each McNary detection-efficiency stratum, the associated estimated detection efficiencies, and the number of expanded McNary detections are given in Table A.2; the information in this table was that used to create the data summaries given in Table A.1.

The Brood-Year-2000 estimators for Release-to-McNary survival indices are somewhat different than those for previous brood years. For the 2000-brood (2002outmigrant) McNary detection efficiencies were based on detections of all Spring Chinook released into the Upper Yakima; in previous years the efficiencies were based on only OCT-SNT releases. Further, for the 2000 brood, McNary detections of OCTSNT fish were restricted to only those fish previously detected when exiting the acclimation ponds; for previous broods, all OCT-SNT detections at McNary were used, whether or not previously detected exiting raceways. Efforts will be made next year to standardize the estimation procedures over all brood years (1997 through 2001). Alternative survival estimation procedures will also be investigated.

Weighted logistic analyses of variation were also used to analyze the released-to-tagged-number ratios and to analyze the release-to-Roza adult return survival estimates; the weighting variable for the former being the number of fish tagged, and the weighting variable for the latter being the number of volitionally released fish ${ }^{20}$. The tagged number and tagged-to-released-number ratio in text Table 1.d) served as the database used for the analyses of the released-to-tagged-number ratio. The data in text Tables 2.a.3) and 2.b.3) were used respectively for the 1997-brood and 1998-brood smolt-toadult survival analyses.

Whenever standard errors are presented, the binomially based standard errors from computer output were expanded by the error mean deviance. The underlying assumption is not that the variance of the survival proportion is that for the binomial; rather, the assumption is that variance is proportional to that for the binomial, the mean deviance serving as an estimate of the proportionality constant.

[^20]Table A.1. Estimates of Release-to-McNary Survival Indices within Release-Day Periods ${ }^{21}$ and over the Whole Release Period

| Site <br> Pond | Treatment | Ending Exit Date 3/24/02 |  | Ending Exit Date 4/3/02 |  | Ending Exit Date 4/13/02 |  | Ending Exit Date 4/23/02 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Number <br> Released | Survival Index | Number <br> Released | Survival Index | Number <br> Released | Survival Index | Number <br> Released | Survival Index |
| CF-01 | SNT | 151 | 0.05072 | 438 | 0.08163 | 400 | 0.01025 | 345 | 0.05097 |
| CF-02 | OCT | 110 | 0.22380 | 298 | 0.311 | 460 | 0.28992 | 675 | 0.38709 |
| CF-03 | SNT | 203 | 0.36923 | 983 | 0.31901 | 198 | 0.14313 | 542 | 0.38474 |
| CF-04 | OCT | 198 | 0.33230 | 573 | 0.42616 | 382 | 0.47036 | 878 | 0.50649 |
| CF-05 | SNT | 135 | 0.49492 | 932 | 0.40021 | 113 | 0.49991 | 706 | 0.57202 |
| CF-06 | OCT | 120 | 0.40213 | 707 | 0.5012 | 208 | 0.4685 | 914 | 0.54506 |
| CF-Total |  | 917 | 0.31417 | 3931 | 0.35959 | 1761 | 0.2836 | 4060 | 0.45176 |
| JC-01 | SNT | 325 | 0.32059 | 616 | 0.26605 | 449 | 0.29584 | 673 | 0.35453 |
| JC-02 | OCT | 668 | 0.34291 | 124 | 0.26099 | 225 | 0.35853 | 499 | 0.35276 |
| JC-03 | SNT | 479 | 0.44313 | 514 | 0.32033 | 492 | 0.36269 | 671 | 0.34086 |
| JC-04 | OCT | 525 | 0.41620 | 208 | 0.39241 | 320 | 0.43631 | 894 | 0.45612 |
| JC-05 | SNT | 338 | 0.22071 | 354 | 0.1921 | 332 | 0.23627 | 793 | 0.26268 |
| JC-06 | OCT | 805 | 0.36533 | 116 | 0.33281 | 217 | 0.29117 | 650 | 0.4006 |
| JC_Total |  | 3140 | 0.36074 | 1932 | 0.28423 | 2035 | 0.3308 | 4180 | 0.36359 |
| ES-01 | SNT | 1469 | 0.31048 | 142 | 0.2923 | 68 | 0.31491 | 396 | 0.33975 |
| ES-02 | OCT | 1249 | 0.34783 | 42 | 0.24894 | 43 | 0.52654 | 540 | 0.42618 |
| ES-03 | SNT | 767 | 0.22171 | 273 | 0.13526 | 180 | 0.07432 | 599 | 0.09668 |
| ES-04 | OCT | 1643 | 0.28811 | 43 | 0.15946 | 21 | 0.21123 | 236 | 0.32748 |
| ES-05 | SNT | 666 | 0.08191 | 352 | 0.06921 | 209 | 0.05115 | 367 | 0.03715 |
| ES-06 | OCT | 1203 | 0.30710 | 61 | 0.20266 | 55 | 0.83463 | 614 | 0.28665 |
| ES-Total |  | 6997 | 0.27983 | 913 | 0.14509 | 576 | 0.20567 | 2752 | 0.25055 |


| Site <br> Pond | Treatment | Ending Exit Date 5/3/02 |  | Ending Exit Date 5/13/02 |  | Ending Exit Date 5/23/02 |  | Ending Exit Date 6/2/02 |  | Ending Exit Date <br> Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Number Released | Survival Index | Number Released | Survival Index | Number Released | Survival Index | Number Released | Survival Index | Number Released | Survival Index |
| CF-01 | SNT | 212 | 0.03039 | 56 | 0.04045 | 11 | 0.00000 | 5 | 0.00000 | 1618 | 0.04561 |
| CF-02 | OCT | 253 | 0.09186 | 88 | 0.11018 | 60 | 0.03775 | 9 | 0.00000 | 1953 | 0.28016 |
| CF-03 | SNT | 127 | 0.15757 | 31 | 0.28396 | 10 | 0.00000 | 0 | 0.00000 | 2094 | 0.31243 |
| CF-04 | OCT | 56 | 0.12672 | 42 | 0.06150 | 32 | 0.27958 | 25 | 0.35333 | 2186 | 0.43999 |
| CF-05 | SNT | 144 | 0.30253 | 36 | 0.35437 | 55 | 0.49930 | 25 | 0.17667 | 2146 | 0.46055 |
| CF-06 | OCT | 107 | 0.32265 | 83 | 0.33623 | 42 | 0.41918 | 20 | 0.39547 | 2201 | 0.49350 |
| CF-Total |  | 899 | 0.15003 | 336 | 0.19051 | 210 | 0.26799 | 84 | 0.25190 | 12198 | 0.35346 |
| JC-01 | SNT | 31 | 0.08333 | 30 | 0.07550 | 55 | 0.35399 | 0 | 0.00000 | 2179 | 0.30464 |
| JC-02 | OCT | 404 | 0.24546 | 192 | 0.21760 | 25 | 0.00000 | 0 | 0.00000 | 2137 | 0.30841 |
| JC-03 | SNT | 4 | 0.00000 | 19 | 0.35166 | 6 | 0.00000 | 0 | 0.00000 | 2185 | 0.36190 |
| JC-04 | OCT | 95 | 0.43308 | 86 | 0.26144 | 54 | 0.68714 | 0 | 0.00000 | 2182 | 0.43457 |
| JC-05 | SNT | 96 | 0.08830 | 139 | 0.20672 | 50 | 0.00000 | 0 | 0.00000 | 2102 | 0.22196 |
| JC-06 | OCT | 127 | 0.31195 | 233 | 0.27587 | 13 | 0.00000 | 0 | 0.00000 | 2161 | 0.35176 |
| JC Total |  | 757 | 0.25229 | 699 | 0.23780 | 203 | 0.27870 | 0 | 0.00000 | 12946 | 0.33127 |
| ES-01 | SNT | 81 | 0.49146 | 3 | 1.47222 | 2 | 0.00000 | 0 | 0.00000 | 2161 | 0.32290 |
| ES-02 | OCT | 225 | 0.25159 | 46 | 0.34420 | 40 | 0.16987 | 0 | 0.00000 | 2185 | 0.35557 |
| ES-03 | SNT | 163 | 0.04011 | 28 | 0.00000 | 9 | 0.00000 | 6 | 0.00000 | 2026 | 0.14058 |
| ES-04 | OCT | 133 | 0.25346 | 61 | 0.07240 | 22 | 0.17976 | 1 | 0.00000 | 2160 | 0.27964 |
| ES-05 | SNT | 96 | 0.00000 | 32 | 0.07078 | 14 | 0.00000 | 1 | 0.00000 | 1737 | 0.06074 |
| ES-06 | OCT | 164 | 0.30739 | 44 | 0.26204 | 26 | 0.59673 | 0 | 0.00000 | 2167 | 0.31434 |
| ES-Total |  | 862 | 0.21703 | 214 | 0.17972 | 113 | 0.23243 | 8 | 0.00000 | 12436 | 0.25331 |

[^21]
# Table. A.2. Numbers Used to Estimate Release-to-McNary Survival Indices for 2002 OCT-SNT releases into the Upper Yakima 

## a. Site--Clark Flats, Raceway 1, SNT Treatment <br> Total Detections

| Total Detections Ending Release Date |  | Detection <br> Efficiency <br> Stratum 1 | Detection <br> Efficiency <br> Stratum 2 | Detection <br> Efficiency <br> Stratum 3 | Detection <br> Efficiency Stratum 4 | Detection <br> Efficiency <br> Stratum 5 | Detection <br> Efficiency <br> Stratum 6 | Detection <br> Efficiency <br> Stratum 7 | Detection <br> Efficiency <br> Stratum 8 | Detection <br> Efficiency <br> Stratum 9 | Detection <br> Efficiency <br> Stratum 10 | Detection <br> Efficiency <br> Stratum 11 | Detection <br> Efficiency <br> Stratum 12 | Detection <br> Efficiency <br> Stratum 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 | 5/14/2002 | 5/24/2002 | 5/27/2002 | 5/30/2002 | 7/31/2002 |
| 3/24/02 | 83 | 0 | 0 | 0 | 0 | 1 | 2 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| 4/3/02 | 93 | 2 | 1 | 0 | 1 | 1 | 5 | 3 | 2 | 1 | 1 | 0 | 0 | 0 |
| 4/13/02 | 103 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 |
| 4/23/02 | 113 | 0 | 1 | 0 | 0 | 1 | 5 | 1 | 0 | 0 | 1 | 0 | 0 | 0 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 1 | 0 | 0 | 0 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


| Expanded Detections Ending Release Date |  | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Expanded |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Detection | Release | Survival |
| Calendar | Julian | 0.3325609 | 0.388115 | 0.4170322 | 0.4589023 | 0.5131199 | 0.579215 | 0.5426723 | 0.4958233 | 0.4431187 | 0.3871335 | 0.4415138 | 0.2528658 | 0.2264151 | Total | Number |  |
| 3/24/02 | 83 | 0.00 | 0.00 | 0.00 | 0.00 | 1.95 | 3.45 | 0.00 | 0.00 | 2.26 | 0.00 | 0.00 | 0.00 | 0.00 | 7.6585429 | 151 | 0.0507 |
| 4/3/02 | 93 | 6.01 | 2.58 | 0.00 | 2.18 | 1.95 | 8.63 | 5.53 | 4.03 | 2.26 | 2.58 | 0.00 | 0.00 | 0.00 | 35.752554 | 438 | 0.0816 |
| 4/13/02 | 103 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 1.84 | 0.00 | 2.26 | 0.00 | 0.00 | 0.00 | 0.00 | 4.0994644 | 400 | 0.0102 |
| 4/23/02 | 113 | 0.00 | 2.58 | 0.00 | 0.00 | 1.95 | 8.63 | 1.84 | 0.00 | 0.00 | 2.58 | 0.00 | 0.00 | 0.00 | 17.583612 | 345 | 0.0510 |
| 5/3/02 | 123 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 1.84 | 2.02 | 0.00 | 2.58 | 0.00 | 0.00 | 0.00 | 6.4426686 | 212 | 0.0304 |
| 5/13/02 | 133 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.26 | 0.00 | 0.00 | 2.2649351 | 56 | 0.0404 |
| 5/23/02 | 143 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 | 11 | 0.0000 |
| 6/2/02 | 153 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 | 5 | 0.0000 |
| 6/12/02 | 163 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 | 0 |  |
|  | Total | 6.01 | 5.15 | 0.00 | 2.18 | 5.85 | 20.72 | 11.06 | 6.05 | 6.77 | 7.75 | 2.26 | 0.00 | 0.00 | 73.801777 | 1618 | 0.0456 |

## b. Site--Clark Flats, Raceway 2, OCT Treatment

| Total Detections Ending Release Date |  | Detection | Detection | Defection | Defection | Deffection | Detection | Detection Efficiency | Deffection | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Stratum 1 | Stratum 2 | Stratum 3 | Stratum 4 | Stratum 5 | Stratum 6 | Stratum 7 | Stratum 8 | Stratum 9 | Stratum 10 | Stratum 11 | Stratum 12 | Stratum 13 |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 | 5/14/2002 | 5/24/2002 | 5/27/2002 | 5/30/2002 | 7/31/2002 |
| 3/24/02 | 83 | 3 | 3 | 1 | 0 | 0 | 2 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 4/3/02 | 93 | 7 | 8 | 8 | 5 | 2 | 6 | 1 | 0 | 1 | 1 | 0 | 0 | 0 |
| 4/13/02 | 103 | 2 | 2 | 3 | 10 | 10 | 22 | 6 | 10 | 2 | 0 | 0 | 0 | 0 |
| 4/23/02 | 113 | 1 | 0 | 5 | 9 | 11 | 66 | 15 | 17 | 5 | 7 | 0 | 0 | 0 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 1 | 5 | 0 | 0 | 0 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 0 | 0 |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Expanded Total Detections

| Expanded Detections Ending Release Date |  | Detection | Detection | Detection | Detection | Detection | Detfection | Detection Efficiency | Detection Efficiency | Detection | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | ExpandedDetectionTotal | Release Number | Survival Index |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Calendar | Julian | 0.3325609 | 0.388115 | 0.4170322 | 0.4589023 | 0.5131199 | 0.579215 | 0.5426723 | 0.4958233 | 0.4431187 | 0.3871335 | 0.4415138 | 0.2528658 | 0.2264151 |  |  |  |
| 3/24/02 | 83 | 9.02 | 7.73 | 2.40 | 0.00 | 0.00 | 3.45 | 0.00 | 2.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 24.618266 | 110 | 0.2238 |
| 4/3/02 | 93 | 21.05 | 20.61 | 19.18 | 10.90 | 3.90 | 10.36 | 1.84 | 0.00 | 2.26 | 2.58 | 0.00 | 0.00 | 0.00 | 92.679089 | 298 | 0.3110 |
| 4/13/02 | 103 | 6.01 | 5.15 | 7.19 | 21.79 | 19.49 | 37.98 | 11.06 | 20.17 | 4.51 | 0.00 | 0.00 | 0.00 | 0.00 | 133.36127 | 460 | 0.2899 |
| 4/23/02 | 113 | 3.01 | 0.00 | 11.99 | 19.61 | 21.44 | 113.95 | 27.64 | 34.29 | 11.28 | 18.08 | 0.00 | 0.00 | 0.00 | 261.28595 | 675 | 0.3871 |
| 5/3/02 | 123 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 8.07 | 2.26 | 12.92 | 0.00 | 0.00 | 0.00 | 23.239563 | 253 | 0.0919 |
| 5/13/02 | 133 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 5.17 | 4.53 | 0.00 | 0.00 | 9.6960469 | 88 | 0.1102 |
| 5/23/02 | 143 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.26 | 0.00 | 0.00 | 2.2649351 | 60 | 0.0377 |
| 6/2/02 | 153 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 | 9 | 0.0000 |
| 6/12/02 | 163 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 | 0 |  |
|  | Total | 39.09 | 33.50 | 40.76 | 52.30 | 44.82 | 165.74 | 40.54 | 64.54 | 20.31 | 38.75 | 6.79 | 0.00 | 0.00 | 547.14512 | 1953 | 0.2802 |

# Table. A.2. Numbers Used to Estimate Release-to-McNary Survival Indices for 2002 OCT-SNT releases into the Upper Yakima (continued) 

## c. Site--Clark Flats, Raceway 3, SNT Treatment <br> Total Detections

| Total Detections Ending Release Date |  | Detection Efficiency | Detection | Detection Efficiency | Detection | Detection Efficiency | Detection | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Stratum 1 | Stratum 2 | Stratum 3 | Stratum 4 | Stratum 5 | Stratum 6 | Stratum 7 | Stratum 8 | Stratum 9 | Stratum 10 | Stratum 11 | Stratum 12 | Stratum 13 |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 | 5/14/2002 | 5/24/2002 | 5/27/2002 | 5/30/2002 | 7/31/2002 |
| 3/24/02 | 83 | 6 | 5 | 3 | 2 | 6 | 7 | 1 | 1 | 1 | 1 | 0 | 0 | 0 |
| 4/3/02 | 93 | 7 | 11 | 19 | 15 | 12 | 55 | 12 | 12 | 6 | 3 | 0 | 0 | 0 |
| 4/13/02 | 103 | 0 | 0 | 1 | 0 | 3 | 6 | 2 | 3 | 0 | 0 | 0 | 0 | 0 |
| 4/23/02 | 113 | 0 | 0 | 3 | 3 | 4 | 37 | 11 | 27 | 10 | 10 | 0 | 0 | 0 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 6 | 0 | 0 | 0 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 0 |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |



## d. Site--Clark Flats, Raceway 4, OCT Treatment

| Total Detections |  | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ending Release Date |  | Stratum 1 | Stratum 2 | Stratum 3 | Stratum 4 | Stratum 5 | Stratum 6 | Stratum 7 | Stratum 8 | Stratum 9 | Stratum 10 | Stratum 11 | Stratum 12 | Stratum 13 |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 | 5/14/2002 | 5/24/2002 | 5/27/2002 | 5/30/2002 | 7/31/2002 |
| 3/24/02 | 83 | 4 | 5 | 3 | 2 | 2 | 6 | 1 | 4 | 0 | 2 | 0 | 0 | 0 |
| 4/3/02 | 93 | 8 | 8 | 11 | 10 | 11 | 43 | 11 | 7 | 6 | 3 | 0 | 0 | 0 |
| 4/13/02 | 103 | 2 | 0 | 3 | 8 | 8 | 42 | 16 | 10 | 5 | 0 | 0 | 0 | 0 |
| 4/23/02 | 113 | 0 | 0 | 5 | 7 | 12 | 92 | 37 | 43 | 22 | 11 | 1 | 0 | 0 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 1 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Expanded Total Detections

| Expanded Detections <br> Ending Release Date |  | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Expanded |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Detection | Release | Survival |
| Calendar | Julian | 0.3325609 | 0.388115 | 0.4170322 | 0.4589023 | 0.5131199 | 0.579215 | 0.5426723 | 0.4958233 | 0.4431187 | 0.3871335 | 0.4415138 | 0.2528658 | 0.2264151 | Total | Number | Index |
| 3/24/02 | 83 | 12.03 | 12.88 | 7.19 | 4.36 | 3.90 | 10.36 | 1.84 | 8.07 | 0.00 | 5.17 | 0.00 | 0.00 | 0.00 | 65.80 | 198 | 0.3323 |
| 4/3/02 | 93 | 24.06 | 20.61 | 26.38 | 21.79 | 21.44 | 74.24 | 20.27 | 14.12 | 13.54 | 7.75 | 0.00 | 0.00 | 0.00 | 244.19 | 573 | 0.4262 |
| 4/13/02 | 103 | 6.01 | 0.00 | 7.19 | 17.43 | 15.59 | 72.51 | 29.48 | 20.17 | 11.28 | 0.00 | 0.00 | 0.00 | 0.00 | 179.68 | 382 | 0.4704 |
| 4/23/02 | 113 | 0.00 | 0.00 | 11.99 | 15.25 | 23.39 | 158.84 | 68.18 | 86.72 | 49.65 | 28.41 | 2.26 | 0.00 | 0.00 | 444.70 | 878 | 0.5065 |
| 5/3/02 | 123 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 4.51 | 2.58 | 0.00 | 0.00 | 0.00 | 7.10 | 56 | 0.1267 |
| 5/13/02 | 133 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.58 | 0.00 | 0.00 | 0.00 | 2.58 | 42 | 0.0615 |
| 5/23/02 | 143 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 4.53 | 0.00 | 4.42 | 8.95 | 32 | 0.2796 |
| 6/2/02 | 153 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 8.83 | 8.83 | 25 | 0.3533 |
| 6/12/02 | 163 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
|  | Total | 42.10 | 33.50 | 52.75 | 58.84 | 64.31 | 315.94 | 119.78 | 129.08 | 78.99 | 46.50 | 6.79 | 0.00 | 13.25 | 961.82 | 2186 | 0.4400 |

# Table. A.2. Numbers Used to Estimate Release-to-McNary Survival Indices for 2002 OCT-SNT releases into the Upper Yakima (continued) 

e. Site--Clark Flats, Raceway 5, SNT Treatment

| Total Detections Ending Release Date |  | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Stratum 1 | Stratum 2 | Stratum 3 | Stratum 4 | Stratum 5 | Stratum 6 | Stratum 7 | Stratum 8 | Stratum 9 | Stratum 10 | Stratum 11 | Stratum 12 | Stratum 13 |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 | 5/14/2002 | 5/24/2002 | 5/27/2002 | 5/30/2002 | 7/31/2002 |
| 3/24/02 | 83 | 9 | 4 | 3 | 2 | 1 | 7 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| 4/3/02 | 93 | 18 | 17 | 16 | 17 | 15 | 55 | 10 | 11 | 4 | 10 | 0 | 0 | 0 |
| 4/13/02 | 103 | 0 | 3 | 2 | 2 | 2 | 9 | 6 | 1 | 2 | 1 | 0 | 0 | 0 |
| 4/23/02 | 113 | 0 | 4 | 5 | 8 | 11 | 78 | 21 | 38 | 21 | 16 | 0 | 1 | 0 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 12 | 0 | 1 | 0 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 2 | 0 |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 4 | 0 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Expanded Total Detections

| Expanded Detections Ending Release Date |  | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Expanded Detection Total | Release Number | Survival Index |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Calendar | Julian | 0.3325609 | 0.388115 | 0.4170322 | 0.4589023 | 0.5131199 | 0.579215 | 0.5426723 | 0.4958233 | 0.4431187 | 0.3871335 | 0.4415138 | 0.2528658 | 0.2264151 |  |  |  |
| 3/24/02 | 83 | 27.06 | 10.31 | 7.19 | 4.36 | 1.95 | 12.09 | 1.84 | 2.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 66.814619 | 135 | 0.4949 |
| 4/3/02 | 93 | 54.13 | 43.80 | 38.37 | 37.04 | 29.23 | 94.96 | 18.43 | 22.19 | 9.03 | 25.83 | 0.00 | 0.00 | 0.00 | 372.99765 | 932 | 0.4002 |
| 4/13/02 | 103 | 0.00 | 7.73 | 4.80 | 4.36 | 3.90 | 15.54 | 11.06 | 2.02 | 4.51 | 2.58 | 0.00 | 0.00 | 0.00 | 56.489479 | 113 | 0.4999 |
| 4/23/02 | 113 | 0.00 | 10.31 | 11.99 | 17.43 | 21.44 | 134.67 | 38.70 | 76.64 | 47.39 | 41.33 | 0.00 | 3.95 | 0.00 | 403.84415 | 706 | 0.5720 |
| 5/3/02 | 123 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 1.84 | 0.00 | 6.77 | 31.00 | 0.00 | 3.95 | 0.00 | 43.564655 | 144 | 0.3025 |
| 5/13/02 | 133 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.58 | 2.26 | 7.91 | 0.00 | 12.757357 | 36 | 0.3544 |
| 5/23/02 | 143 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.58 | 9.06 | 15.82 | 0.00 | 27.461495 | 55 | 0.4993 |
| 6/2/02 | 153 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 4.42 | 4.4166667 | 25 | 0.1767 |
| 6/12/02 | 163 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 | 0 |  |
|  | Total | 81.19 | 72.14 | 62.35 | 63.19 | 56.52 | 257.24 | 71.87 | 102.86 | 67.70 | 105.91 | 11.32 | 31.64 | 4.42 | 988.34607 | 2146 | 0.4606 |

## f. Site--Clark Flats, Raceway 6, OCT Treatment

Total Detections

| Total Detections Ending Release Date |  | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection <br> Efficiency <br> Stratum 8 <br> 5/8/2002 | Detection <br> Efficiency <br> Stratum 9 <br> 5/14/2002 | Detection <br> Efficiency <br> Stratum 10 <br> 5/24/2002 | Detection <br> Efficiency <br> Stratum 11 <br> 5/27/2002 | Detection Efficiency Stratum 12 5/30/2002 | Detection Efficiency |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Stratum 1 | Stratum 2 | Stratum 3 | Stratum 4 | Stratum 5 | Stratum 6 | Stratum 7 |  |  |  |  |  | Stratum 13 |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 |  |  |  |  |  | 7/31/2002 |
| 3/24/02 | 83 | 7 | 2 | 1 | 2 | 1 | 4 | 1 | 1 | 0 | 1 | 0 | 0 | 0 |
| 4/3/02 | 93 | 25 | 24 | 16 | 7 | 15 | 39 | 10 | 15 | 3 | 3 | 0 | 1 | 0 |
| 4/13/02 | 103 | 1 | 1 | 3 | 3 | 2 | 23 | 6 | 7 | 3 | 1 | 0 | 0 | 0 |
| 4/23/02 | 113 | 1 | 2 | 13 | 12 | 20 | 110 | 34 | 41 | 15 | 8 | 0 | 1 | 0 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 6 | 5 | 0 | 0 | 0 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 5 | 0 | 2 |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 2 | 0 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Expanded Total Detections

| Expanded Detections Ending Release Date |  | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Expanded |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Detection Total | Release Number | Survival Index |
| Calendar | Julian | 0.3325609 | 0.388115 | 0.4170322 | 0.4589023 | 0.5131199 | 0.579215 | 0.5426723 | 0.4958233 | 0.4431187 | 0.3871335 | 0.4415138 | 0.2528658 | 0.2264151 |  |  |  |
| 3/24/02 | 83 | 21.05 | 5.15 | 2.40 | 4.36 | 1.95 | 6.91 | 1.84 | 2.02 | 0.00 | 2.58 | 0.00 | 0.00 | 0.00 | 48.26 | 120 | 0.4021 |
| 4/3/02 | 93 | 75.17 | 61.84 | 38.37 | 15.25 | 29.23 | 67.33 | 18.43 | 30.25 | 6.77 | 7.75 | 0.00 | 3.95 | 0.00 | 354.35 | 707 | 0.5012 |
| 4/13/02 | 103 | 3.01 | 2.58 | 7.19 | 6.54 | 3.90 | 39.71 | 11.06 | 14.12 | 6.77 | 2.58 | 0.00 | 0.00 | 0.00 | 97.45 | 208 | 0.4685 |
| 4/23/02 | 113 | 3.01 | 5.15 | 31.17 | 26.15 | 38.98 | 189.91 | 62.65 | 82.69 | 33.85 | 20.66 | 0.00 | 3.95 | 0.00 | 498.19 | 914 | 0.5451 |
| 5/3/02 | 123 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 8.07 | 13.54 | 12.92 | 0.00 | 0.00 | 0.00 | 34.52 | 107 | 0.3226 |
| 5/13/02 | 133 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 7.75 | 11.32 | 0.00 | 8.83 | 27.91 | 83 | 0.3362 |
| 5/23/02 | 143 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 5.17 | 4.53 | 7.91 | 0.00 | 17.61 | 42 | 0.4192 |
| 6/2/02 | 153 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 7.91 | 0.00 | 7.91 | 20 | 0.3955 |
| 6/12/02 | 163 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
|  | Total | 102.24 | 74.72 | 79.13 | 52.30 | 74.06 | 303.86 | 93.98 | 137.15 | 60.93 | 59.41 | 15.85 | 23.73 | 8.83 | 1086.19 | 2201 | 0.4935 |

# Table. A.2. Numbers Used to Estimate Release-to-McNary Survival Indices for 2002 OCT-SNT releases into the Upper Yakima (continued) 

## g. Site--Jack Creek, Raceway 1, SNT Treatment <br> Total Detections

| Total Detections Ending Release Date |  | Detection Efficiency | Detection | Detection Efficiency | Detection | Detection Efficiency | Detection | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Stratum 1 | Stratum 2 | Stratum 3 | Stratum 4 | Stratum 5 | Stratum 6 | Stratum 7 | Stratum 8 | Stratum 9 | Stratum 10 | Stratum 11 | Stratum 12 | Stratum 13 |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 | 5/14/2002 | 5/24/2002 | 5/27/2002 | 5/30/2002 | 7/31/2002 |
| 3/24/02 | 83 | 4 | 7 | 6 | 1 | 5 | 17 | 1 | 2 | 1 | 4 | 0 | 0 | 0 |
| 4/3/02 | 93 | 1 | 2 | 6 | 6 | 5 | 28 | 8 | 10 | 5 | 6 | 2 | 1 | 0 |
| 4/13/02 | 103 | 1 | 1 | 1 | 3 | 4 | 27 | 6 | 11 | 7 | 4 | 2 | 0 | 0 |
| 4/23/02 | 113 | 0 | 0 | 1 | 2 | 4 | 21 | 16 | 25 | 13 | 28 | 1 | 1 | 0 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 3 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


| Expanded Detections Ending Release Date |  | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Expanded |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Detection | Release | Survival |
| Calendar | Julian | 0.3325609 | 0.388115 | 0.4170322 | 0.4589023 | 0.5131199 | 0.579215 | 0.5426723 | 0.4958233 | 0.4431187 | 0.3871335 | 0.4415138 | 0.2528658 | 0.2264151 | Total | Number |  |
| 3/24/02 | 83 | 12.03 | 18.04 | 14.39 | 2.18 | 9.74 | 29.35 | 1.84 | 4.03 | 2.26 | 10.33 | 0.00 | 0.00 | 0.00 | 104.19 | 325 | 0.3206 |
| 4/3/02 | 93 | 3.01 | 5.15 | 14.39 | 13.07 | 9.74 | 48.34 | 14.74 | 20.17 | 11.28 | 15.50 | 4.53 | 3.95 | 0.00 | 163.88 | 616 | 0.2660 |
| 4/13/02 | 103 | 3.01 | 2.58 | 2.40 | 6.54 | 7.80 | 46.61 | 11.06 | 22.19 | 15.80 | 10.33 | 4.53 | 0.00 | 0.00 | 132.83 | 449 | 0.2958 |
| 4/23/02 | 113 | 0.00 | 0.00 | 2.40 | 4.36 | 7.80 | 36.26 | 29.48 | 50.42 | 29.34 | 72.33 | 2.26 | 3.95 | 0.00 | 238.60 | 673 | 0.3545 |
| 5/3/02 | 123 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.58 | 0.00 | 0.00 | 0.00 | 2.58 | 31 | 0.0833 |
| 5/13/02 | 133 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.26 | 0.00 | 0.00 | 2.26 | 30 | 0.0755 |
| 5/23/02 | 143 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.26 | 3.95 | 13.25 | 19.47 | 55 | 0.3540 |
| 6/2/02 | 153 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
| 6/12/02 | 163 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
|  | Total | 18.04 | 25.77 | 33.57 | 26.15 | 35.08 | 160.56 | 57.12 | 96.81 | 58.68 | 111.07 | 15.85 | 11.86 | 13.25 | 663.82 | 2179 | 0.3046 |

## h. Site--Jack Creek, Raceway 2, OCT Treatment

| Total Detections Ending Release Date |  | Detection Efficiency <br> Stratum 1 | Detection <br> Efficiency <br> Stratum 2 | Detection <br> Efficiency <br> Stratum 3 | Detection Efficiency <br> Stratum 4 | Detection Efficiency <br> Stratum 5 | Detection <br> Efficiency <br> Stratum 6 | Detection Efficiency <br> Stratum 7 | Detection Efficiency <br> Stratum 8 | Detection <br> Efficiency <br> Stratum 9 | Detection Efficiency <br> Stratum 10 | Detection Efficiency <br> Stratum 11 | Detection Efficiency <br> Stratum 12 | Detection Efficiency <br> Stratum 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Stratum 1 | Stratum 2 | Stratum 3 | Stratum 4 | Stratum 5 | Stratum 6 | Stratum 7 | Stratum 8 | Stratum 9 | Stratum 10 | Stratum 11 | Stratum 12 | Stratum 13 |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 | 5/14/2002 | 5/24/2002 | 5/27/2002 | 5/30/2002 | 7/31/2002 |
| 3/24/02 | 83 | 9 | 10 | 9 | 5 | 9 | 32 | 12 | 6 | 8 | 4 | 0 | 1 | 1 |
| 4/3/02 | 93 | 1 | 1 | 1 | 1 | 2 | 3 | 2 | 1 | 1 | 2 | 0 | 0 | 0 |
| 4/13/02 | 103 | 1 | 0 | 1 | 1 | 1 | 10 | 8 | 4 | 4 | 7 | 0 | 1 | 0 |
| 4/23/02 | 113 | 0 | 0 | 0 | 3 | 1 | 10 | 8 | 17 | 16 | 20 | 4 | 0 | 1 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 25 | 4 | 2 | 2 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 3 | 4 | 2 |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Expanded Total Detections

| Expanded Detections Ending Release Date |  | Detection | Detection | Detection | Detection | Detection | Detfection | Detection Efficiency | Detection Efficiency | Detection | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | $\begin{gathered} \hline \text { Expanded } \\ \text { Detection } \\ \text { Total } \end{gathered}$ | Release Number | Survival Index |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Calendar | Julian | 0.3325609 | 0.388115 | 0.4170322 | 0.4589023 | 0.5131199 | 0.579215 | 0.5426723 | 0.4958233 | 0.4431187 | 0.3871335 | 0.4415138 | 0.2528658 | 0.2264151 |  |  |  |
| 3/24/02 | 83 | 27.06 | 25.77 | 21.58 | 10.90 | 17.54 | 55.25 | 22.11 | 12.10 | 18.05 | 10.33 | 0.00 | 3.95 | 4.42 | 229.06 | 668 | 0.3429 |
| 4/3/02 | 93 | 3.01 | 2.58 | 2.40 | 2.18 | 3.90 | 5.18 | 3.69 | 2.02 | 2.26 | 5.17 | 0.00 | 0.00 | 0.00 | 32.36 | 124 | 0.2610 |
| 4/13/02 | 103 | 3.01 | 0.00 | 2.40 | 2.18 | 1.95 | 17.26 | 14.74 | 8.07 | 9.03 | 18.08 | 0.00 | 3.95 | 0.00 | 80.67 | 225 | 0.3585 |
| 4/23/02 | 113 | 0.00 | 0.00 | 0.00 | 6.54 | 1.95 | 17.26 | 14.74 | 34.29 | 36.11 | 51.66 | 9.06 | 0.00 | 4.42 | 176.03 | 499 | 0.3528 |
| 5/3/02 | 123 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.02 | 6.77 | 64.58 | 9.06 | 7.91 | 8.83 | 99.17 | 404 | 0.2455 |
| 5/13/02 | 133 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 10.33 | 6.79 | 15.82 | 8.83 | 41.78 | 192 | 0.2176 |
| 5/23/02 | 143 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 25 | 0.0000 |
| 6/2/02 | 153 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
| 6/12/02 | 163 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
|  | Total | 33.08 | 28.34 | 26.38 | 21.79 | 25.34 | 94.96 | 55.28 | 58.49 | 72.22 | 160.15 | 24.91 | 31.64 | 26.50 | 659.07 | 2137 | 0.3084 |

# Table. A.2. Numbers Used to Estimate Release-to-McNary Survival Indices for 2002 OCT-SNT releases into the Upper Yakima (continued) 

## i. Site--Jack Creek, Raceway 3, SNT Treatment <br> Total Detections

| Total Detections Ending Release Date |  | Detection <br> Efficiency <br> Stratum 1 | Detection <br> Efficiency <br> Stratum 2 | Detection <br> Efficiency <br> Stratum 3 | Detection <br> Efficiency <br> Stratum 4 | Detection <br> Efficiency <br> Stratum 5 | Detection <br> Efficiency <br> Stratum 6 | Detection <br> Efficiency <br> Stratum 7 | Detection <br> Efficiency <br> Stratum 8 | Detection <br> Efficiency <br> Stratum 9 | Detection <br> Efficiency <br> Stratum 10 | Detection <br> Efficiency <br> Stratum 11 | Detection Efficiency Stratum 12 | Detection <br> Efficiency <br> Stratum 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 | 5/14/2002 | 5/24/2002 | 5/27/2002 | 5/30/2002 | 7/31/2002 |
| 3/24/02 | 83 | 14 | 7 | 8 | 7 | 5 | 26 | 9 | 12 | 3 | 6 | 0 | 0 | 0 |
| 4/3/02 | 93 | 7 | 4 | 4 | 10 | 4 | 20 | 9 | 13 | 3 | 3 | 1 | 0 | 0 |
| 4/13/02 | 103 | 1 | 1 | 3 | 3 | 7 | 26 | 13 | 14 | 9 | 10 | 1 | 0 | 0 |
| 4/23/02 | 113 | 0 | 0 | 2 | 1 | 4 | 26 | 10 | 21 | 16 | 24 | 1 | 2 | 0 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | , |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


| Expanded Detections Ending Release Date |  | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Expanded |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Detection | Release | Survival |
| Calendar | Julian | 0.3325609 | 0.388115 | 0.4170322 | 0.4589023 | 0.5131199 | 0.579215 | 0.5426723 | 0.4958233 | 0.4431187 | 0.3871335 | 0.4415138 | 0.2528658 | 0.2264151 | Total | Number | Index |
| 3/24/02 | 83 | 42.10 | 18.04 | 19.18 | 15.25 | 9.74 | 44.89 | 16.58 | 24.20 | 6.77 | 15.50 | 0.00 | 0.00 | 0.00 | 212.26 | 479 | 0.4431 |
| 4/3/02 | 93 | 21.05 | 10.31 | 9.59 | 21.79 | 7.80 | 34.53 | 16.58 | 26.22 | 6.77 | 7.75 | 2.26 | 0.00 | 0.00 | 164.65 | 514 | 0.3203 |
| 4/13/02 | 103 | 3.01 | 2.58 | 7.19 | 6.54 | 13.64 | 44.89 | 23.96 | 28.24 | 20.31 | 25.83 | 2.26 | 0.00 | 0.00 | 178.44 | 492 | 0.3627 |
| 4/23/02 | 113 | 0.00 | 0.00 | 4.80 | 2.18 | 7.80 | 44.89 | 18.43 | 42.35 | 36.11 | 61.99 | 2.26 | 7.91 | 0.00 | 228.72 | 671 | 0.3409 |
| 5/3/02 | 123 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 4 | 0.0000 |
| 5/13/02 | 133 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.26 | 0.00 | 4.42 | 6.68 | 19 | 0.3517 |
| 5/23/02 | 143 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 6 | 0.0000 |
| 6/2/02 | 153 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
| 6/12/02 | 163 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
|  | Total | 66.15 | 30.92 | 40.76 | 45.76 | 38.98 | 169.19 | 75.55 | 121.01 | 69.96 | 111.07 | 9.06 | 7.91 | 4.42 | 790.75 | 2185 | 0.3619 |

j. Site--Jack Creek, Raceway 4, OCT Treatment

| Total D | ctions | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection <br> Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ending Re | ase Date | Stratum 1 | Stratum 2 | Stratum 3 | Stratum 4 | Stratum 5 | Stratum 6 | Stratum 7 | Stratum 8 | Stratum 9 | Stratum 10 | Stratum 11 | Stratum 12 | Stratum 13 |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 | 5/14/2002 | 5/24/2002 | 5/27/2002 | 5/30/2002 | 7/31/2002 |
| 3/24/02 | 83 | 17 | 14 | 11 | 6 | 8 | 20 | 4 | 6 | 3 | 6 | 0 | 0 | 0 |
| 4/3/02 | 93 | 2 | 2 | 8 | 2 | 2 | 13 | 0 | 3 | 3 | 3 | 0 | 0 | 0 |
| 4/13/02 | 103 | 1 | 2 | 2 | 7 | 7 | 27 | 5 | 8 | 8 | 3 | 0 | 0 | 0 |
| 4/23/02 | 113 | 0 | 1 | 4 | 5 | 10 | 49 | 32 | 41 | 30 | 25 | 1 | 0 | 1 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 4 | 5 | 3 | 1 | 1 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 1 | 2 |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 2 | 5 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Expanded Total Detections

| Expanded Detections Ending Release Date |  | Detection | Detection | Detection | Detection | Detection | Detfection | Detection Efficiency | Detection Efficiency | Detection | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | $\begin{gathered} \hline \text { Expanded } \\ \text { Detection } \\ \text { Total } \end{gathered}$ | Release Number | Survival Index |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Calendar | Julian | 0.3325609 | 0.388115 | 0.4170322 | 0.4589023 | 0.5131199 | 0.579215 | 0.5426723 | 0.4958233 | 0.4431187 | 0.3871335 | 0.4415138 | 0.2528658 | 0.2264151 |  |  |  |
| 3/24/02 | 83 | 51.12 | 36.07 | 26.38 | 13.07 | 15.59 | 34.53 | 7.37 | 12.10 | 6.77 | 15.50 | 0.00 | 0.00 | 0.00 | 218.50 | 525 | 0.4162 |
| 4/3/02 | 93 | 6.01 | 5.15 | 19.18 | 4.36 | 3.90 | 22.44 | 0.00 | 6.05 | 6.77 | 7.75 | 0.00 | 0.00 | 0.00 | 81.62 | 208 | 0.3924 |
| 4/13/02 | 103 | 3.01 | 5.15 | 4.80 | 15.25 | 13.64 | 46.61 | 9.21 | 16.13 | 18.05 | 7.75 | 0.00 | 0.00 | 0.00 | 139.62 | 320 | 0.4363 |
| 4/23/02 | 113 | 0.00 | 2.58 | 9.59 | 10.90 | 19.49 | 84.60 | 58.97 | 82.69 | 67.70 | 64.58 | 2.26 | 0.00 | 4.42 | 407.77 | 894 | 0.4561 |
| 5/3/02 | 123 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 4.03 | 9.03 | 12.92 | 6.79 | 3.95 | 4.42 | 41.14 | 95 | 0.4331 |
| 5/13/02 | 133 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 5.17 | 4.53 | 3.95 | 8.83 | 22.48 | 86 | 0.2614 |
| 5/23/02 | 143 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.58 | 4.53 | 7.91 | 22.08 | 37.11 | 54 | 0.6871 |
| 6/2/02 | 153 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
| 6/12/02 | 163 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
|  | Total | 60.14 | 48.95 | 59.95 | 43.58 | 52.62 | 188.19 | 75.55 | 121.01 | 108.32 | 116.24 | 18.12 | 15.82 | 39.75 | 948.24 | 2182 | 0.4346 |

Table. A.2. Numbers Used to Estimate Release-to-McNary Survival Indices for 2002 OCT-SNT releases into the Upper Yakima (continued)
k. Site--Jack Creek, Raceway 5, SNT Treatment

| Total D | ctions | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection Efficiency | Detection Efficiency | Detection Efficiency |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ending Re | se Date | Stratum 1 | Stratum 2 | Stratum 3 | Stratum 4 | Stratum 5 | Stratum 6 | Stratum 7 | Stratum 8 | Stratum 9 | Stratum 10 | Stratum 11 | Stratum 12 | Stratum 13 |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 | 5/14/2002 | 5/24/2002 | 5/27/2002 | 5/30/2002 | 7/31/2002 |
| 3/24/02 | 83 | 3 | 1 | 4 | 1 | 0 | 15 | 6 | 1 | 2 | 3 | 0 | 0 | 0 |
| 4/3/02 | 93 | 1 | 1 | 2 | 1 | 4 | 8 | 4 | 8 | 0 | 4 | 0 | 0 | 0 |
| 4/13/02 | 103 | 0 | 0 | 2 | 1 | 1 | 11 | 4 | 7 | 6 | 6 | 0 | 0 | 0 |
| 4/23/02 | 113 | 0 | 0 | 1 | 1 | 3 | 25 | 13 | 17 | 20 | 19 | 1 | 0 | 0 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 1 | 0 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 1 | 4 |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


| Expanded Total Detections |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Expanded Detections Ending Release Date |  | DetectionEfficiency0.3325609 | DetectionEfficiency0.388115 | Detection | DetectionEfficiency0.4589023 | DetectionEfficiency0.5131199 | Detection | DetectionEfficiency0.5426723 | DetectionEfficiency0.4958233 | $\begin{aligned} & \hline \text { Detection } \\ & \text { Efficiency } \\ & 0.4431187 \end{aligned}$ | DetectionEfficiency0.3871335 | DetectionEfficiency0.4415138 | Detection Detection <br> Efficiency Efficiency <br> 0.2528658 0.2264151 |  | Expanded |  |  |
|  |  | Efficiency |  | Efficiency |  |  | $\begin{aligned} & \text { Detection } \\ & \text { Total } \end{aligned}$ |  |  |  |  |  |  |  | Release Number | Survival Index |
| Calendar | Julian |  |  | 0.4170322 |  |  |  |  |  |  |  |  |  |  | 0.579215 |  |
| 3/24/02 | 83 |  | 9.02 | 2.58 | 9.59 | 2.18 | 0.00 | 25.90 | 11.06 | 2.02 | 4.51 | 7.75 | 0.00 | 0.00 |  | 0.00 | 74.60 | 338 | 0.2207 |
| 4/3/02 | 93 | 3.01 | 2.58 | 4.80 | 2.18 | 7.80 | 13.81 | 7.37 | 16.13 | 0.00 | 10.33 | 0.00 | 0.00 | 0.00 | 68.00 | 354 | 0.1921 |
| 4/13/02 | 103 | 0.00 | 0.00 | 4.80 | 2.18 | 1.95 | 18.99 | 7.37 | 14.12 | 13.54 | 15.50 | 0.00 | 0.00 | 0.00 | 78.44 | 332 | 0.2363 |
| 4/23/02 | 113 | 0.00 | 0.00 | 2.40 | 2.18 | 5.85 | 43.16 | 23.96 | 34.29 | 45.13 | 49.08 | 2.26 | 0.00 | 0.00 | 208.31 | 793 | 0.2627 |
| 5/3/02 | 123 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.26 | 0.00 | 2.26 | 3.95 | 0.00 | 8.48 | 96 | 0.0883 |
| 5/13/02 | 133 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.58 | 4.53 | 3.95 | 17.67 | 28.73 | 139 | 0.2067 |
| 5/23/02 | 143 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 50 | 0.0000 |
| 6/2/02 | 153 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
| 6/12/02 | 163 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
|  | Total | 12.03 | 5.15 | 21.58 | 8.72 | 15.59 | 101.86 | 49.75 | 66.56 | 65.45 | 85.24 | 9.06 | 7.91 | 17.67 | 466.56 | 2102 | 0.2220 |

## 1. Site--Jack Creek, Raceway 6, OCT Treatment

| Total Detections |  | Defection | Detection | Defection | Defection | Deffection | Detection | Detection Efficiency | Deffection | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ending Release Date |  | Stratum 1 | Stratum 2 | Stratum 3 | Stratum 4 | Stratum 5 | Stratum 6 | Stratum 7 | Stratum 8 | Stratum 9 | Stratum 10 | Stratum 11 | Stratum 12 | Stratum 13 |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 | 5/14/2002 | 5/24/2002 | 5/27/2002 | 5/30/2002 | 7/31/2002 |
| 3/24/02 | 83 | 11 | 10 | 12 | 7 | 13 | 47 | 16 | 13 | 6 | 6 | 0 | 0 | 0 |
| 4/3/02 | 93 | 2 | 2 | 1 | 0 | 1 | 6 | 1 | 3 | 1 | 1 | 0 | 0 | 0 |
| 4/13/02 | 103 | 1 | 1 | 1 | 0 | 1 | 11 | 2 | 7 | 4 | 2 | 1 | 0 | 0 |
| 4/23/02 | 113 | 0 | 0 | 1 | 3 | 3 | 30 | 21 | 22 | 21 | 18 | 2 | 2 | 1 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 12 | 1 | 0 | 0 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 5 | 1 | 7 |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Expanded Total Detections

| Expanded Detections Ending Release Date |  | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | $\begin{aligned} & \hline \text { Detection } \\ & \text { Efficiency } \end{aligned}$ | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Expanded <br> Detection <br> Total | Release Number | Survival Index |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Calendar | Julian | 0.3325609 | 0.388115 | 0.4170322 | 0.4589023 | 0.5131199 | 0.579215 | 0.5426723 | 0.4958233 | 0.4431187 | 0.3871335 | 0.4415138 | 0.2528658 | 0.2264151 |  |  |  |
| 3/24/02 | 83 | 33.08 | 25.77 | 28.77 | 15.25 | 25.34 | 81.14 | 29.48 | 26.22 | 13.54 | 15.50 | 0.00 | 0.00 | 0.00 | 294.09 | 805 | 0.3653 |
| 4/3/02 | 93 | 6.01 | 5.15 | 2.40 | 0.00 | 1.95 | 10.36 | 1.84 | 6.05 | 2.26 | 2.58 | 0.00 | 0.00 | 0.00 | 38.61 | 116 | 0.3328 |
| 4/13/02 | 103 | 3.01 | 2.58 | 2.40 | 0.00 | 1.95 | 18.99 | 3.69 | 14.12 | 9.03 | 5.17 | 2.26 | 0.00 | 0.00 | 63.18 | 217 | 0.2912 |
| 4/23/02 | 113 | 0.00 | 0.00 | 2.40 | 6.54 | 5.85 | 51.79 | 38.70 | 44.37 | 47.39 | 46.50 | 4.53 | 7.91 | 4.42 | 260.39 | 650 | 0.4006 |
| 5/3/02 | 123 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 1.84 | 0.00 | 4.51 | 31.00 | 2.26 | 0.00 | 0.00 | 39.62 | 127 | 0.3120 |
| 5/13/02 | 133 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 18.08 | 11.32 | 3.95 | 30.92 | 64.28 | 233 | 0.2759 |
| 5/23/02 | 143 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 13 | 0.0000 |
| 6/2/02 | 153 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
| 6/12/02 | 163 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
|  | Total | 42.10 | 33.50 | 35.97 | 21.79 | 35.08 | 162.29 | 75.55 | 90.76 | 76.73 | 118.82 | 20.38 | 11.86 | 35.33 | 760.16 | 2161 | 0.3518 |

# Table. A.2. Numbers Used to Estimate Release-to-McNary Survival Indices for 2002 OCT-SNT releases into the Upper Yakima (continued) 

## m. Site--Easton, Raceway 1, SNT Treatment <br> Total Detections

| Total Detections Ending Release Date |  | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Stratum 1 | Stratum 2 | Stratum 3 | Stratum 4 | Stratum 5 | Stratum 6 | Stratum 7 | Stratum 8 | Stratum 9 | Stratum 10 | Stratum 11 | Stratum 12 | Stratum 13 |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 | 5/14/2002 | 5/24/2002 | 5/27/2002 | 5/30/2002 | 7/31/2002 |
| 3/24/02 | 83 | 21 | 15 | 12 | 14 | 19 | 70 | 15 | 13 | 11 | 20 | 3 | 0 | 0 |
| 4/3/02 | 93 | 0 | 0 | 2 | 0 | 3 | 7 | 2 | 2 | 1 | 1 | 1 | 1 | 0 |
| 4/13/02 | 103 | 0 | 0 | 0 | 0 | 1 | 4 | 2 | 2 | 1 | 1 | 0 | 0 | 0 |
| 4/23/02 | 113 | 0 | 0 | 0 | 2 | 0 | 8 | 9 | 6 | 6 | 22 | 2 | 1 | 2 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 7 | 1 | 1 | 3 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


| Expanded Total Detections |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Expanded Detections Ending Release Date |  | DetectionEfficiency0.3325609 | Detection Efficiency 0.388115 | $\begin{aligned} & \hline \text { Detection } \\ & \text { Efficiency } \\ & 0.4170322 \end{aligned}$ | DetectionEfficiency0.4589023 | DetectionEfficiency0.5131199 | $\begin{aligned} & \hline \text { Detection } \\ & \text { Efficiency } \\ & 0.579215 \end{aligned}$ | DetectionEfficiency0.5426723 | DetectionEfficiency0.4958233 | $\begin{gathered} \hline \text { Detection } \\ \text { Efficiency } \\ 0.4431187 \end{gathered}$ | DetectionEfficiency0.3871335 | DetectionEfficiency0.4415138 | Detection Detection <br> Efficiency Efficiency <br> 0.2528658 0.2264151 |  | Expanded |  |  |
|  |  | Detection |  |  |  |  |  |  |  |  |  |  |  |  | Release | Survival |
| Calendar | Julian |  |  |  |  |  |  |  |  |  |  |  |  |  | Total | Number | Index |
| 3/24/02 | 83 |  | 63.15 | 38.65 | 28.77 | 30.51 | 37.03 | 120.85 | 27.64 | 26.22 | 24.82 | 51.66 | 6.79 | 0.00 | 0.00 | 456.10 | 1469 | 0.3105 |
| 4/3/02 | 93 | 0.00 | 0.00 | 4.80 | 0.00 | 5.85 | 12.09 | 3.69 | 4.03 | 2.26 | 2.58 | 2.26 | 3.95 | 0.00 | 41.51 | 142 | 0.2923 |
| 4/13/02 | 103 | 0.00 | 0.00 | 0.00 | 0.00 | 1.95 | 6.91 | 3.69 | 4.03 | 2.26 | 2.58 | 0.00 | 0.00 | 0.00 | 21.41 | 68 | 0.3149 |
| 4/23/02 | 113 | 0.00 | 0.00 | 0.00 | 4.36 | 0.00 | 13.81 | 16.58 | 12.10 | 13.54 | 56.83 | 4.53 | 3.95 | 8.83 | 134.54 | 396 | 0.3398 |
| 5/3/02 | 123 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.26 | 18.08 | 2.26 | 3.95 | 13.25 | 39.81 | 81 | 0.4915 |
| 5/13/02 | 133 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 4.42 | 4.42 | 3 | 1.4722 |
| 5/23/02 | 143 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2 | 0.0000 |
| 6/2/02 | 153 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
| 6/12/02 | 163 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
|  | Total | 63.15 | 38.65 | 33.57 | 34.87 | 44.82 | 153.66 | 51.60 | 46.39 | 45.13 | 131.74 | 15.85 | 11.86 | 26.50 | 697.79 | 2161 | 0.3229 |

## n. Site--Easton, Raceway 2, OCT Treatment

Total Detections

| Total Detections |  | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ending Release Date |  | Stratum 1 | Stratum 2 | Stratum 3 | Stratum 4 | Stratum 5 | Stratum 6 | Stratum 7 | Stratum 8 | Stratum 9 | Stratum 10 | Stratum 11 | Stratum 12 | Stratum 13 |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 | 5/14/2002 | 5/24/2002 | 5/27/2002 | 5/30/2002 | 7/31/2002 |
| 3/24/02 | 83 | 21 | 22 | 15 | 18 | 15 | 60 | 23 | 16 | 5 | 8 | 0 | 0 | 0 |
| 4/3/02 | 93 | 0 | 1 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 1 | 0 | 0 | 0 |
| 4/13/02 | 103 | 0 | 0 | 1 | 0 | 2 | 4 | 0 | 1 | 1 | 2 | 0 | 0 | 0 |
| 4/23/02 | 113 | 0 | 0 | 0 | 2 | 3 | 20 | 17 | 13 | 24 | 27 | 0 | 1 | 0 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 10 | 4 | 1 | 3 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 3 |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Expanded Total Detections

| Expanded Detections <br> Ending Release Date |  | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Detection | Expanded |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Efficiency | Detection | Release | Survival |
| Calendar | Julian | 0.3325609 | 0.388115 | 0.4170322 | 0.4589023 | 0.5131199 | 0.579215 | 0.5426723 | 0.4958233 | 0.4431187 | 0.3871335 | 0.4415138 | 0.2528658 | 0.2264151 | Total | Number | Index |
| 3/24/02 | 83 | 63.15 | 56.68 | 35.97 | 39.22 | 29.23 | 103.59 | 42.38 | 32.27 | 11.28 | 20.66 | 0.00 | 0.00 | 0.00 | 434.45 | 1249 | 0.3478 |
| 4/3/02 | 93 | 0.00 | 2.58 | 0.00 | 0.00 | 0.00 | 3.45 | 1.84 | 0.00 | 0.00 | 2.58 | 0.00 | 0.00 | 0.00 | 10.46 | 42 | 0.2489 |
| 4/13/02 | 103 | 0.00 | 0.00 | 2.40 | 0.00 | 3.90 | 6.91 | 0.00 | 2.02 | 2.26 | 5.17 | 0.00 | 0.00 | 0.00 | 22.64 | 43 | 0.5265 |
| 4/23/02 | 113 | 0.00 | 0.00 | 0.00 | 4.36 | 5.85 | 34.53 | 31.33 | 26.22 | 54.16 | 69.74 | 0.00 | 3.95 | 0.00 | 230.14 | 540 | 0.4262 |
| 5/3/02 | 123 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 4.51 | 25.83 | 9.06 | 3.95 | 13.25 | 56.61 | 225 | 0.2516 |
| 5/13/02 | 133 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.58 | 0.00 | 0.00 | 13.25 | 15.83 | 46 | 0.3442 |
| 5/23/02 | 143 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 6.79 | 0.00 | 0.00 | 6.79 | 40 | 0.1699 |
| 6/2/02 | 153 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
| 6/12/02 | 163 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
|  | Total | 63.15 | 59.26 | 38.37 | 43.58 | 38.98 | 148.48 | 75.55 | 60.51 | 72.22 | 126.57 | 15.85 | 7.91 | 26.50 | 776.92 | 2185 | 0.3556 |

# Table. A.2. Numbers Used to Estimate Release-to-McNary Survival Indices for 2002 OCT-SNT releases into the Upper Yakima (continued) 

## o. Site--Easton, Raceway 3, SNT Treatment <br> Total Detections

| Total Detections Ending Release Date |  | Detection <br> Efficiency <br> Stratum 1 | Detection <br> Efficiency <br> Stratum 2 | Detection <br> Efficiency <br> Stratum 3 | Detection Efficiency Stratum 4 | Detection <br> Efficiency <br> Stratum 5 | Detection <br> Efficiency <br> Stratum 6 | $\begin{aligned} & \text { Detection } \\ & \text { Efficiency } \\ & \text { Stratum } 7 \end{aligned}$ | Detection <br> Efficiency <br> Stratum 8 | Detection <br> Efficiency <br> Stratum 9 | Detection <br> Efficiency <br> Stratum 10 | Detection Efficiency Stratum 11 | Detection Efficiency Stratum 12 | Detection <br> Efficiency <br> Stratum 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 | 5/14/2002 | 5/24/2002 | 5/27/2002 | 5/30/2002 | 7/31/2002 |
| 3/24/02 | 83 | 9 | 10 | 7 | 6 | 3 | 22 | 9 | 6 | 2 | 4 | 0 | 0 | 0 |
| 4/3/02 | 93 | 0 | 0 | 1 | 2 | 1 | 5 | 1 | 4 | 2 | 2 | 0 | 0 | 0 |
| 4/13/02 | 103 | 0 | 0 | 0 | 0 | 0 | 3 | 2 | 0 | 2 | 0 | 0 | 0 | 0 |
| 4/23/02 | 113 | 0 | 0 | 0 | 0 | 0 | 2 | 6 | 2 | 6 | 10 | 0 | 0 | 0 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


| Expanded Total Detections |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Expanded Detections Ending Release Date |  | DetectionEfficiency0.3325609 | DetectionEfficiency0.388115 | Detection | DetectionEfficiency0.4589023 | DetectionEfficiency0.5131199 | Detection | DetectionEfficiency0.5426723 | DetectionEfficiency0.4958233 | $\begin{aligned} & \hline \text { Detection } \\ & \text { Efficiency } \\ & 0.4431187 \end{aligned}$ | DetectionEfficiency0.3871335 | DetectionEfficiency0.4415138 | Detection Detection <br> Efficiency Efficiency <br> 0.2528658 0.2264151 |  | Expanded |  |  |
|  |  | Efficiency |  | Efficiency |  |  | $\begin{aligned} & \text { Detection } \\ & \text { Total } \end{aligned}$ |  |  |  |  |  |  |  | Release Number | Survival Index |
| Calendar | Julian |  |  | 0.4170322 |  |  |  |  |  |  |  |  |  |  | 0.579215 |  |
| 3/24/02 | 83 |  | 27.06 | 25.77 | 16.79 | 13.07 | 5.85 | 37.98 | 16.58 | 12.10 | 4.51 | 10.33 | 0.00 | 0.00 |  | 0.00 | 170.05 | 767 | 0.2217 |
| 4/3/02 | 93 | 0.00 | 0.00 | 2.40 | 4.36 | 1.95 | 8.63 | 1.84 | 8.07 | 4.51 | 5.17 | 0.00 | 0.00 | 0.00 | 36.93 | 273 | 0.1353 |
| 4/13/02 | 103 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 5.18 | 3.69 | 0.00 | 4.51 | 0.00 | 0.00 | 0.00 | 0.00 | 13.38 | 180 | 0.0743 |
| 4/23/02 | 113 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 3.45 | 11.06 | 4.03 | 13.54 | 25.83 | 0.00 | 0.00 | 0.00 | 57.91 | 599 | 0.0967 |
| 5/3/02 | 123 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.58 | 0.00 | 3.95 | 0.00 | 6.54 | 163 | 0.0401 |
| 5/13/02 | 133 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 28 | 0.0000 |
| 5/23/02 | 143 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 9 | 0.0000 |
| 6/2/02 | 153 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 6 | 0.0000 |
| 6/12/02 | 163 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 1 |  |
|  | Total | 27.06 | 25.77 | 19.18 | 17.43 | 7.80 | 55.25 | 33.17 | 24.20 | 27.08 | 43.91 | 0.00 | 3.95 | 0.00 | 284.81 | 2026 | 0.1406 |

p. Site--Easton, Raceway 4, OCT Treatment

| Total Detections Ending Release Date |  | Detection <br> Efficiency <br> Stratum 1 | Detection <br> Efficiency <br> Stratum 2 | Detection <br> Efficiency <br> Stratum 3 | Detection <br> Efficiency <br> Stratum 4 | Detection <br> Efficiency <br> Stratum 5 | Detection <br> Efficiency <br> Stratum 6 | Detection <br> Efficiency <br> Stratum 7 | Detection <br> Efficiency <br> Stratum 8 | Detection Efficiency Stratum 9 | Detection <br> Efficiency <br> Stratum 10 | Detection <br> Efficiency <br> Stratum 11 | Detection <br> Efficiency <br> Stratum 12 | Detection Efficiency Stratum 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 | 5/14/2002 | 5/24/2002 | 5/27/2002 | 5/30/2002 | 7/31/2002 |
| 3/24/02 | 83 | 17 | 13 | 16 | 11 | 22 | 69 | 24 | 30 | 15 | 10 | 0 | 0 | 0 |
| 4/3/02 | 93 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 |
| 4/13/02 | 103 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| 4/23/02 | 113 | 0 | 0 | 0 | 0 | 0 | 5 | 1 | 12 | 4 | 13 | 0 | 0 | 0 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 3 | 0 | 2 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Expanded Total Detections

| Expanded Detections Ending Release Date |  | Detection Efficiency 0.3325609 | Detection Efficiency 0.388115 | $\begin{aligned} & \hline \text { Detection } \\ & \text { Efficiency } \\ & 0.4170322 \end{aligned}$ | $\begin{aligned} & \hline \text { Detection } \\ & \text { Efficiency } \\ & 0.4589023 \end{aligned}$ | $\begin{aligned} & \hline \text { Detection } \\ & \text { Efficiency } \\ & 0.5131199 \end{aligned}$ | $\begin{aligned} & \hline \text { Detection } \\ & \text { Efficiency } \\ & 0.579215 \end{aligned}$ | $\begin{aligned} & \hline \text { Detection } \\ & \text { Efficiency } \\ & 0.5426723 \end{aligned}$ | $\begin{aligned} & \hline \text { Detection } \\ & \text { Efficiency } \\ & 0.4958233 \end{aligned}$ | $\begin{gathered} \hline \text { Detection } \\ \text { Efficiency } \\ 0.4431187 \\ \hline \end{gathered}$ | DetectionEfficiency0.3871335 | $\begin{aligned} & \hline \text { Detection } \\ & \text { Efficiency } \\ & 0.4415138 \\ & \hline \end{aligned}$ | $\begin{aligned} & \hline \text { Detection } \\ & \text { Efficiency } \\ & 0.2528658 \end{aligned}$ | $\begin{array}{\|} \hline \text { Detection } \\ \text { Efficiency } \\ 0.2264151 \\ \hline \end{array}$ | Expanded <br> Detection Total | Release Number | Survival Index |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Calendar | Julian |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 3/24/02 | 83 | 51.12 | 33.50 | 38.37 | 23.97 | 42.87 | 119.13 | 44.23 | 60.51 | 33.85 | 25.83 | 0.00 | 0.00 | 0.00 | 473.36 | 1643 | 0.2881 |
| 4/3/02 | 93 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.02 | 2.26 | 2.58 | 0.00 | 0.00 | 0.00 | 6.86 | 43 | 0.1595 |
| 4/13/02 | 103 | 0.00 | 0.00 | 0.00 | 2.18 | 0.00 | 0.00 | 0.00 | 0.00 | 2.26 | 0.00 | 0.00 | 0.00 | 0.00 | 4.44 | 21 | 0.2112 |
| 4/23/02 | 113 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 8.63 | 1.84 | 24.20 | 9.03 | 33.58 | 0.00 | 0.00 | 0.00 | 77.28 | 236 | 0.3275 |
| 5/3/02 | 123 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 18.08 | 6.79 | 0.00 | 8.83 | 33.71 | 133 | 0.2535 |
| 5/13/02 | 133 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 4.42 | 4.42 | 61 | 0.0724 |
| 5/23/02 | 143 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 3.95 | 0.00 | 3.95 | 22 | 0.1798 |
| 6/2/02 | 153 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 1 | 0.0000 |
| 6/12/02 | 163 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
|  | Total | 51.12 | 33.50 | 38.37 | 26.15 | 42.87 | 127.76 | 46.07 | 86.72 | 47.39 | 80.08 | 6.79 | 3.95 | 13.25 | 604.02 | 2160 | 0.2796 |

Table. A.2. Numbers Used to Estimate Release-to-McNary Survival Indices for 2002 OCT-SNT releases into the Upper Yakima (continued)
q. Site--Easton, Raceway 5, SNT Treatment

Total Detections

| Total Detections Ending Release Date |  | Detection Efficiency Stratum 1 | Detection <br> Efficiency <br> Stratum 2 | Detection <br> Efficiency <br> Stratum 3 | Detection <br> Efficiency <br> Stratum 4 | Detection <br> Efficiency <br> Stratum 5 | Detection <br> Efficiency <br> Stratum 6 | Detection <br> Efficiency <br> Stratum 7 | Detection <br> Efficiency <br> Stratum 8 | Detection <br> Efficiency <br> Stratum 9 | Detection <br> Efficiency <br> Stratum 10 | Detection <br> Efficiency <br> Stratum 11 | Detection <br> Efficiency <br> Stratum 12 | Detection <br> Efficiency <br> Stratum 13 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 | 5/14/2002 | 5/24/2002 | 5/27/2002 | 5/30/2002 | 7/31/2002 |
| 3/24/02 | 83 | 1 | 4 | 1 | 1 | 5 | 3 | 3 | 1 | 4 | 2 | 0 | 0 | 0 |
| 4/3/02 | 93 | 0 | 2 | 1 | 2 | 0 | 1 | 1 | 2 | 0 | 1 | 1 | 0 | 0 |
| 4/13/02 | 103 | 0 | 0 | 1 | 0 | 0 | 2 | 0 | 0 | 1 | 1 | 0 | 0 | 0 |
| 4/23/02 | 113 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 2 | 1 | 2 | 0 | 0 | 0 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Expanded Total Detections

| Expanded Detections Ending Release Date |  | DetectionEfficiency0.3325609 | DetectionEfficiency0.388115 | DetectionEfficiency0.4170322 | DetectionEfficiency0.4589023 | DetectionEfficiency0.5131199 | Detection | DetectionEfficiency0.5426723 | DetectionEfficiency0.4958233 | DetectionEfficiency0.4431187 | DetectionEfficiency0.3871335 | DetectionEfficiency0.4415138 | Detection Detection <br> Efficiency Efficiency <br> 0.2528658 0.2264151 |  | Expanded |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Efficiency |  |  |  |  | $\begin{array}{\|c} \text { Detection } \\ \text { Total } \\ \hline \end{array}$ |  |  |  |  |  |  |  | Release Number | Survival Index |
| Calendar | Julian |  |  |  |  |  |  |  |  |  |  |  |  |  | 0.579215 |  |
| 3/24/02 | 83 |  | 3.01 | 10.31 | 2.40 | 2.18 | 9.74 | 5.18 | 5.53 | 2.02 | 9.03 | 5.17 | 0.00 | 0.00 |  | 0.00 | 54.55 | 666 | 0.0819 |
| 4/3/02 | 93 | 0.00 | 5.15 | 2.40 | 4.36 | 0.00 | 1.73 | 1.84 | 4.03 | 0.00 | 2.58 | 2.26 | 0.00 | 0.00 | 24.36 | 352 | 0.0692 |
| 4/13/02 | 103 | 0.00 | 0.00 | 2.40 | 0.00 | 0.00 | 3.45 | 0.00 | 0.00 | 2.26 | 2.58 | 0.00 | 0.00 | 0.00 | 10.69 | 209 | 0.0512 |
| 4/23/02 | 113 | 0.00 | 0.00 | 0.00 | 2.18 | 0.00 | 0.00 | 0.00 | 4.03 | 2.26 | 5.17 | 0.00 | 0.00 | 0.00 | 13.64 | 367 | 0.0372 |
| 5/3/02 | 123 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 96 | 0.0000 |
| 5/13/02 | 133 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.26 | 0.00 | 0.00 | 2.26 | 32 | 0.0708 |
| 5/23/02 | 143 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 14 | 0.0000 |
| 6/2/02 | 153 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 1 | 0.0000 |
| 6/12/02 | 163 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
|  | Total | 3.01 | 15.46 | 7.19 | 8.72 | 9.74 | 10.36 | 7.37 | 10.08 | 13.54 | 15.50 | 4.53 | 0.00 | 0.00 | 105.50 | 1737 | 0.0607 |

## r. Site--Easton, Raceway 6, OCT Treatment

| Total Detections Ending Release Date |  | Detection Efficiency | Detection Efficiency | Detection | Detection Efficiency | Detection Efficiency | Detection | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Stratum 1 | Stratum 2 | Stratum 3 | Stratum 4 | Stratum 5 | Stratum 6 | Stratum 7 | Stratum 8 | Stratum 9 | Stratum 10 | Stratum 11 | Stratum 12 | Stratum 13 |
| Calendar | Julian | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 | 5/14/2002 | 5/24/2002 | 5/27/2002 | 5/30/2002 | 7/31/2002 |
| 3/24/02 | 83 | 11 | 11 | 15 | 9 | 18 | 57 | 23 | 21 | 5 | 8 | 1 | 0 | 0 |
| 4/3/02 | 93 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 1 | 0 | 2 | 0 | 0 | 0 |
| 4/13/02 | 103 | 0 | 0 | 2 | 3 | 0 | 4 | 1 | 4 | 1 | 6 | 0 | 0 | 0 |
| 4/23/02 | 113 | 0 | 0 | 0 | 2 | 0 | 11 | 7 | 16 | 7 | 22 | 8 | 2 | 2 |
| 5/3/02 | 123 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 12 | 0 | 1 | 1 |
| 5/13/02 | 133 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 0 | 1 |
| 5/23/02 | 143 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 |
| 6/2/02 | 153 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6/12/02 | 163 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Expanded Total Detections

| Expanded Detections Ending Release Date |  | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | Detection Efficiency | $\begin{gathered} \hline \text { Expanded } \\ \text { Detection } \\ \text { Total } \\ \hline \end{gathered}$ | Release Number | Survival Index |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Calendar | Julian | 0.3325609 | 0.388115 | 0.4170322 | 0.4589023 | 0.5131199 | 0.579215 | 0.5426723 | 0.4958233 | 0.4431187 | 0.3871335 | 0.4415138 | 0.2528658 | 0.2264151 |  |  |  |
| 3/24/02 | 83 | 33.08 | 28.34 | 35.97 | 19.61 | 35.08 | 98.41 | 42.38 | 42.35 | 11.28 | 20.66 | 2.26 | 0.00 | 0.00 | 369.44 | 1203 | 0.3071 |
| 4/3/02 | 93 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 5.18 | 0.00 | 2.02 | 0.00 | 5.17 | 0.00 | 0.00 | 0.00 | 12.36 | 61 | 0.2027 |
| 4/13/02 | 103 | 0.00 | 0.00 | 4.80 | 6.54 | 0.00 | 6.91 | 1.84 | 8.07 | 2.26 | 15.50 | 0.00 | 0.00 | 0.00 | 45.90 | 55 | 0.8346 |
| 4/23/02 | 113 | 0.00 | 0.00 | 0.00 | 4.36 | 0.00 | 18.99 | 12.90 | 32.27 | 15.80 | 56.83 | 18.12 | 7.91 | 8.83 | 176.01 | 614 | 0.2867 |
| 5/3/02 | 123 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.02 | 9.03 | 31.00 | 0.00 | 3.95 | 4.42 | 50.41 | 164 | 0.3074 |
| 5/13/02 | 133 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.58 | 4.53 | 0.00 | 4.42 | 11.53 | 44 | 0.2620 |
| 5/23/02 | 143 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2.26 | 0.00 | 13.25 | 15.51 | 26 | 0.5967 |
| 6/2/02 | 153 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
| 6/12/02 | 163 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 |  |
|  | Total | 33.08 | 28.34 | 40.76 | 30.51 | 35.08 | 129.49 | 57.12 | 86.72 | 38.36 | 131.74 | 27.18 | 11.86 | 30.92 | 681.17 | 2167 | 0.3143 |

## Appendix B. Detection Efficiency Estimation

## B.1. Conceptual Computation

1. For each downstream dam, joint McNary and downstream detections were crosstabulated by McNary Dam's first date and downstream-dams' first date of detection [Table B.1.a)].
2. Within each downstream dam's detection date, the relative distribution of joint counts over McNary detection dates was estimated [Table B.1.b)].
3. The resulting relative distribution frequencies from Table B.1.b) were then multiplied by the total downstream dam's detections (whether or not previously detected at McNary) for the given downstream date to obtain estimates of the cross-tab number for the downstream dam's total detections [Table B.1.c)].
4. Once this was done for each downstream dam's detection date, the estimated total downstream detections that were allocated to a given McNary-detection date then were added over downstream-dam detection dates [Table B.1.c), far-right-hand column]. This gave the estimated total downstream-dam detections that passed McNary on the given McNary date.
5. The total joint downstream-dam McNary detections on a given McNary-detection date [Table B.1.a), far-right column] were then divided by the downstream-dam total from step 4 above [Table B.1.c), far-right column], giving an estimated McNary-detection efficiency associated with the McNary date [Table B.1.d), far-right-hand column].

Actually, before the last step, Table B.1.a)'s and Table B.1.b)'s numbers were pooled over John Day and Bonneville Dams ${ }^{22}$.

Daily detection efficiencies were then stratified into contiguous days of relatively homogeneous detection efficiencies, and the daily detection efficiencies were pooled over days within the strata. The strata's beginning and ending dates were chosen in a manner that minimized the variation among OCT-SNT daily detection efficiencies within strata, thereby maximizing the detection-rate variation among strata. This was done using stepwise logistic regression partitioning based on all possible partitionings. In the first step, the partitioning between all possible sets of two strata that minimized the variation among daily detection efficiencies within strata was selected. With that partitioning fixed, establishing two strata, the second partitioning was then selected in a similar manner among all possible sets of two strata within the strata that were already created in the first partitioning. Again, the partitioning that minimized variation among daily

[^22]detection efficiencies within the strata was selected. This second partitioning was then fixed and, along with the first fixed partitioning, established three strata. A third partitioning was similarly developed within the three established strata to form a fourth stratum. The process was continued as long as the difference between the step's created detection rates was significant at the $10 \%$ significance level ( $\mathrm{P} \leq 0.1$ ).

In the stratification process, there were two exceptions that would lead to the rejection of a given partitioning:

1. If either one of the resulting strata had less than twenty joint McNary detections, or
2. If the difference between the John Day Dam-based and Bonneville Dam-based detection-efficiency estimates were inconsistent in sign. For example, if the combined Bonneville-based McNary detection efficiency in one of the created strata was greater than that in an adjacent stratum, but the John Day-based McNary detection efficiency in the one was less than that in the adjacent, then the partitioning was not accepted.

On completion of the stepwise process, each partitioning was shifted at one-day increments between the two adjacent partitionings to see if the variation within strata could be further reduced. If so, the partitioning that resulted in the greatest reduction was selected.

There were rare cases for which downstream dam dates had total counts but had no associated joint downstream-dam and McNary Dam counts. Ignoring these dates would tend to over-estimate the detection efficiency. What was done to adjust for such an overestimation was to augment the joint counts in the following manner:

1. Take such a downstream dam date and use the McNary distributions from six contiguous downstream dates that immediately preceded this non-joint detection date and from six contiguous dates that followed this non-joint detection date.
2. Pool offset ${ }^{23} \mathrm{McNary}$ passage-time distributions from these twelve adjacent dates; and
3. Apply this pooled distribution (as a relative distribution) to the total downstream count for the non-joint-detection date. The resulting McNary-date-distributed counts were then allocated to the stratum to which the McNary date of detection belonged.
[^23]Table B.1. Conceptual method of estimating detection efficiencies
a) Joint McNary Dam (McN) and Downstream Dam (D.S.) Detections (n) by McN and D.S. Detection Dates

|  | D.S. Date (Julian) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\ldots$ | 98 | 99 | 100 | 101 | 102 | 103 | .... | Total |
| 90 | $\cdots$ | 0 | 0 | 0 | 0 | 0 | 0 | $\ldots$ | $\mathrm{n}(90 .$. |
| $\ldots$ | $\ldots$ | $\ldots$ | ... | $\ldots$ | ... | $\ldots$ | $\ldots$ | $\ldots$ | $\cdots$ |
| 94 | $\ldots$ | $\mathrm{n}(94,98)$ | $\mathrm{n}(94,99)$ | $\mathrm{n}(94,100)$ | $\mathrm{n}(94,101)$ | 0 | 0 | $\ldots$ | n (94,.) |
| 95 | $\ldots$ | 0 | $\mathrm{n}(95,99)$ | $\mathrm{n}(95,100)$ | $\mathrm{n}(95,101)$ | $\mathrm{n}(95,102)$ | 0 | $\ldots$ | $\mathrm{n}(95,$. |
| 96 | $\ldots$ | 0 | 0 | $\mathrm{n}(96,100)$ | $\mathrm{n}(96,101)$ | $\mathrm{n}(96,102)$ | $\mathrm{n}(96,103)$ | ... | n(96,.) |
| 97 | $\ldots$ | 0 | 0 | 0 | 0 | $\mathrm{n}(97,102)$ | $\mathrm{n}(97,103)$ | $\ldots$ | n (97,.) |
| 98 | $\ldots$ | 0 | 0 | 0 | 0 | $\mathrm{n}(98,102)$ | $\mathrm{n}(98,103)$ | $\ldots$ | n (98,.) |
| 99 | $\ldots$ | 0 | 0 | 0 | 0 | 0 | 0 | $\ldots$ | $\mathrm{n}(99 .$. |
| $\cdots$ | $\ldots$ | $\cdots$ | $\cdots$ | $\cdots$ | $\cdots$ | $\cdots$ | $\ldots$ | $\ldots$ | $\cdots$ |
| 200 | $\ldots$ | 0 | 0 | 0 | 0 | 0 | 0 | $\ldots$ | $\mathrm{n}(200,$. |
| Total | $\ldots$ | $\mathrm{n}(., 98)$ | $\mathrm{n}(., 99)$ | $\mathrm{n}(., 100)$ | $\mathrm{n}(., 101)$ | n(.,102) | $\mathrm{n}(., 103)$ | $\ldots$ |  |

b) For Each Downstream Site, Estimate Distribution of McNary Date Contributions

| $\begin{gathered} \hline \text { McN } \\ \text { Date } \\ \text { (Julian) } \\ \hline \end{gathered}$ | p(McN,D.S. ) = n[McN,D.S.)/n(., D.S.) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | D.S. Date (Julian) |  |  |  |  |  |
|  | $\ldots$ | 100 | 101 | 102 | 103 |  |
| 90 | $\ldots$ | $\ldots$ | $\ldots$ | ... | $\ldots$ | $\ldots$ |
| ... | $\ldots$ | $\ldots$ | ... | $\ldots$ | $\ldots$ | $\ldots$ |
| 94 | $\ldots$ | $\mathrm{p}(94,100)$ | $p(94,101)$ | 0 | 0 | $\ldots$ |
| 95 | $\ldots$ | p( 95,100$)$ | $\mathrm{p}(95,101)$ | $\mathrm{p}(95,102)=\mathrm{n}(95,102) / \mathrm{n}(., 102)$ | 0 | $\ldots$ |
| 96 | $\ldots$ | $\mathrm{p}(96,100)$ | $p(96,101)$ | $\mathrm{p}(96,102)=\mathrm{n}(96,102) / \mathrm{n}(., 102)$ | $\mathrm{p}(96,103)$ | $\ldots$ |
| 97 | $\ldots$ | 0 | 0 | $\mathrm{p}(97,102)=\mathrm{n}(97,102) / \mathrm{n}(., 102)$ | $\mathrm{p}(97,103)$ | $\ldots$ |
| 98 | ... | 0 | 0 | $\mathrm{p}(98,102)=\mathrm{n}(98,102) / \mathrm{n}(., 102)$ | $\mathrm{p}(98,103)$ | $\ldots$ |
| 99 | $\ldots$ | 0 | 0 | $\mathrm{p}(99,102)=\mathrm{n}(99,102) / \mathrm{n}(., 102)$ | p(99,103) | ... |
| $\ldots$ | $\ldots$ | $\ldots$ | $\ldots$ | ... | $\ldots$ | $\ldots$ |
| 200 | $\ldots$ | 0 | 0 | 0 | 0 | $\ldots$ |
| Total |  | 1.000 | 1.000 | 1.000 | 1.000 |  |

Table B.1. Conceptual method of estimating detection efficiencies (continued)
c) Allocate Daily Lower Site Counts [N(D.S.)] over McNary Dates using above Distributions and total over Lower Dam Dates within McNary Dates

| McN | N'(McN,D.S. $)=\mathrm{N}(\text { D.S. })^{*} \mathrm{P}(\mathrm{McN}, \mathrm{D} . \mathrm{S}$. |  |  |  |  |  | McN |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date | D.S. Date (Julian) |  |  |  |  |  | Dam |
| (Julian) | $\ldots$ | 100 | 101 | 102 | 103 | $\ldots$ | Total |
| 90 | $\ldots$ | 0 | 0 | 0 | 0 | $\ldots$ | N'(90,.) |
| $\cdots$ | $\ldots$ | ... | ... | $\ldots$ | $\ldots$ | $\cdots$ | $\cdots$ |
| 94 | $\ldots$ | $\mathrm{N}^{\prime}(94,100)$ | $\mathrm{N}^{\prime}(94,101)$ | 0 | 0 | $\ldots$ | $\mathbf{N}^{\prime}(94 .$. |
| 95 | $\ldots$ | $\mathrm{N}^{\prime}(95,100)$ | $\mathrm{N}^{\prime}(95,101)$ | $N^{\prime}(95,102)=p(95,102) * N(., 102)$ | 0 | $\ldots$ | $\mathbf{N}^{\prime}(95,$. |
| 96 | $\ldots$ | $\mathrm{N}^{\prime}(96,100)$ | $\mathrm{N}^{\prime}(96,101)$ | $N^{\prime}(96,102)=p(96,102) * N(., 102)$ | $\mathrm{N}^{\prime}(96,103)$ | $\ldots$ | $\mathbf{N}^{\prime}(96,$. |
| 97 | $\ldots$ | 0 | 0 | $N^{\prime}(97,102)=p(97,102)^{*} \mathrm{~N}(., 102)$ | $\mathrm{N}^{\prime}(97,103)$ | $\cdots$ | N'(97,.) |
| 98 | $\ldots$ | 0 | 0 | $N^{\prime}(98,102)=p(98,102)^{*} N(., 102)$ | $\mathrm{N}^{\prime}(98,103)$ | $\cdots$ | $\mathbf{N}^{\prime}(98,$. |
| 99 | $\ldots$ | 0 | 0 | $N^{\prime}(99,102)=p(99,102) * N(., 102)$ | $\mathrm{N}^{\prime}(99,103)$ | $\ldots$ | N'(99,.) |
| $\ldots$ | $\ldots$ | $\ldots$ | $\ldots$ | $\cdots$ | $\cdots$ | $\cdots$ | $\cdots$ |
| 200 | $\ldots$ | 0 | 0 | 0 | 0 | $\ldots$ | $\mathrm{N}^{\prime}(200,$. |
| Total |  | $\mathrm{N}(100)$ | N (101) | $\mathrm{N}(102)$ | N (103) | ... |  |

d) Use Total Joint McNary and Downstream Dam Detections [Table a)] and Total Downstream Dam Detections [Table c)] to estimate McNary Detection Efficiencies (McN D.E.)

| McNary Dam Date (Julian) | Table a) Total | Table c) $\mathbf{N}^{\prime}$ Total | McNary Detection Efficiency McN D.E. $=\mathrm{n} / \mathbf{N}^{\prime}$ |
| :---: | :---: | :---: | :---: |
| 90 | n(90,.) | N'(90,.) | McN D.E.(90,.)=n(90,.)/N'(90,.) |
|  |  | $\ldots$ |  |
| 94 | n (94,.) | $\mathrm{N}^{\prime}(94,$. | McN D.E.(94,.)=n(94,.)/N'(94,.) |
| 95 | $\mathrm{n}(95,$. | $\mathrm{N}^{\prime}(95,$. | McN D.E.(95,.)=n(95,.)/N'(95,.) |
| 96 | $\mathrm{n}(96,$. | $\mathrm{N}^{\prime}(96,$. | McN D.E.(96,.)=n(96,.)/ $\mathbf{N}^{\prime}(96,$. |
| 97 | $\mathrm{n}(97,$. | N'(97,.) | McN D.E.(97,.)=n(97,.)/N'(97,.) |
| 98 | n(98,.) | $\mathrm{N}^{\prime}(98,$. | McN D.E.(98,.)=n(98,.)/N'(98,.) |
| 99 | n(99,.) | $\mathrm{N}^{\prime}(99,$. | McN D.E.(99,.)=n(99,.)/ $\mathbf{N}^{\prime}(99,$. |
| . |  |  | McN D. E (200 ) $\mathrm{n}(200) \mathrm{N}$ ( 200 ) |
| 200 | $\mathrm{n}(200,$. | $\mathrm{N}(200,$. | McN D.E.(200,.)=n(200,.)/N'(200,.) |

## B.2. 2002 Detection Efficiencies

The detection efficiencies used are given in Table B.2. In the table, the "Augmented" includes downstream detections from days for which there were no associated joint downstream-dam and McNary Dam counts. The detection efficiencies actually used and presented before in Appendix Table A. 2 are those listed under "Augmented" under "Pooled over John Day and Bonneville" in Table B.2. The stepwise logistic analysis of variation leading to the stratification is given in Table B.3.

Assumptions behind the detection efficiency estimation procedures are as follows:

1. Detected and undetected fish passing McNary on a given date are temporally and spatially mixed before reaching the downstream detectors so that their proportional composition at the time of McNary passage will be the same for the surviving fish passing through downstream detectors;
2. Survivals from McNary to downstream detectors are the same for all routes of McNary passage (e.g., survival is the same for fish whether they pass through the bypass, the turbines, or the spillway);
3. The allocations of total downstream dam counts to McNary days of passage are accurate; and
4. The pooled detection rate estimates from John Dam and Bonneville Dams are accurate.

Assumption 2 is unlikely to hold. Assumption 3 is also unlikely to hold, because the method of allocation assumes that the McNary detection efficiencies for a given day of downstream-dam detection are homogeneous. It is unlikely that all fish detected on a given downstream date passed McNary on days for which the detection rates were homogeneous. The estimated detection efficiencies are probably biased, but the bias would be less than assuming a single detection-efficiency value for the whole of McNary passage. For Assumption 4 to hold for the methods used in this report, the probability of a fish being entrained into the bypass at Bonneville would have to be independent of whether or not that fish was entrained into a bypass at John Day or McNary and the probability of a fish being entrained into the bypass at John Day would have to be independent of whether or not that fish was entrained into the bypass at McNary.

The stratification was more effective in explaining variation in daily detection efficiencies than either the daily proportion of McNary flow that is spilled (spill proportion) or the daily proportion of flow that is passed through the turbines (turbine proportion). Table B. 4 gives a logistic analysis of variation that assesses the effects of strata, spill proportion, and turbine proportion on estimated daily McNary detection efficiencies; Table B. 5 gives the data used in the regression. The effect of strata was stronger than those of spill proportion and turbine proportion. Although the effects of spill and turbine proportions are significant ( $\mathrm{P}<0.0001$, Table B.4), their effects when adjusted for the effects of strata are not significant ( $\mathrm{P}=0.347$ for Spill and Turbine separately, $\mathrm{P}=0.672$ for both Spill and Turbine Proportions included in regression, Table B.4.). This does not mean that spill and turbine proportions are not important; rather, it means that the strata differences account for the effects of the spill and turbine proportions. In fact, the effects of strata explain more than the effects of spill and turbine proportions; the strata effects when adjusted for the effects of spill proportion, turbine proportion, and both are still highly significant ( $\mathrm{P}<0.0001$, Table B.4).

Table B.2. Detection efficiencies used to estimate Survival Indices

| Stratum | 1st Date | Last Date | Joint Detection based |  |  | Augmented |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | BO Total | BO-MCJ Total | Estimate | BO Total | BO-MCJ Total | Estimate |
| 1 |  | 4/21/02 | 179.85 | 59 | 0.328046 | 180.85 | 59 | 0.326232 |
| 2 | 4/22/02 | 4/25/02 | 183.85 | 75 | 0.407935 | 183.85 | 75 | 0.407935 |
| 3 | 4/26/02 | 4/29/02 | 109.53 | 50 | 0.456515 | 109.53 | 50 | 0.456515 |
| 4 | 4/30/02 | 5/1/02 | 89.80 | 45 | 0.501104 | 89.80 | 45 | 0.501104 |
| 5 | 5/2/02 | 5/2/02 | 85.68 | 45 | 0.525226 | 85.68 | 45 | 0.525226 |
| 6 | 5/3/02 | 5/5/02 | 448.52 | 265 | 0.590826 | 448.52 | 265 | 0.590826 |
| 7 | 5/6/02 | 5/6/02 | 150.52 | 84 | 0.558056 | 150.52 | 84 | 0.558056 |
| 8 | 5/7/02 | 5/8/02 | 180.00 | 93 | 0.516677 | 180.00 | 93 | 0.516677 |
| 9 | 5/9/02 | 5/14/02 | 150.38 | 68 | 0.452189 | 150.38 | 68 | 0.452189 |
| 10 | 5/15/02 | 5/24/02 | 351.31 | 148 | 0.421278 | 351.31 | 148 | 0.421278 |
| 11 | 5/25/02 | 5/27/02 | 44.55 | 21 | 0.471332 | 44.55 | 21 | 0.471332 |
| 12 | 5/28/02 | 5/30/02 | 34.00 | 9 | 0.264706 | 34.00 | 9 | 0.264706 |
| 13 | 5/31/02 |  | 20.00 | 9 | 0.450000 | 37.00 | 9 | 0.243243 |
|  |  |  | John Day Based |  |  |  |  |  |
|  |  |  | Joint Detection based |  |  |  | Augmented |  |
| Stratum | 1st Date | Last Date | JD Total | JD-MCJ Total | Estimate | JD Total | JD-MCJ Total | Estimate |
| 1 |  | 4/21/02 | 478.68 | 161 | 0.336342 | 480.68 | 161 | 0.334942 |
| 2 | 4/22/02 | 4/25/02 | 388.14 | 147 | 0.378727 | 388.14 | 147 | 0.378727 |
| 3 | 4/26/02 | 4/29/02 | 312.50 | 126 | 0.403194 | 312.50 | 126 | 0.403194 |
| 4 | 4/30/02 | 5/1/02 | 239.24 | 106 | 0.443062 | 239.24 | 106 | 0.443062 |
| 5 | 5/2/02 | 5/2/02 | 288.50 | 147 | 0.509525 | 288.50 | 147 | 0.509525 |
| 6 | 5/3/02 | 5/5/02 | 1334.92 | 768 | 0.575314 | 1334.92 | 768 | 0.575314 |
| 7 | 5/6/02 | 5/6/02 | 241.98 | 129 | 0.533103 | 241.98 | 129 | 0.533103 |
| 8 | 5/7/02 | 5/8/02 | 289.93 | 140 | 0.482877 | 289.93 | 140 | 0.482877 |
| 9 | 5/9/02 | 5/14/02 | 391.24 | 172 | 0.439632 | 391.24 | 172 | 0.439632 |
| 10 | 5/15/02 | 5/24/02 | 410.70 | 147 | 0.357926 | 410.70 | 147 | 0.357926 |
| 11 | 5/25/02 | 5/27/02 | 70.96 | 30 | 0.422790 | 70.96 | 30 | 0.422790 |
| 12 | 5/28/02 | 5/30/02 | 64.87 | 16 | 0.246660 | 64.87 | 16 | 0.246660 |
| 13 | 5/31/02 |  | 37.33 | 11 | 0.294643 | 51.33 | 11 | 0.214286 |
| Stratum | 1st Date | Last Date | Pooled over John Day and Bonneville (Down-stream Dam -based)Joint Detection based |  |  |  |  |  |
|  |  |  | DS Total | DS-MCJ Tota | Estimate | DS Total | DS-MCJ Total | Estimate |
| 1 |  | 4/21/02 | 658.53 | 220 | 0.334076 | 661.53 | 220 | 0.332561 |
| 2 | 4/22/02 | 4/25/02 | 572.00 | 222 | 0.388115 | 572.00 | 222 | 0.388115 |
| 3 | 4/26/02 | 4/29/02 | 422.03 | 176 | 0.417032 | 422.03 | 176 | 0.417032 |
| 4 | 4/30/02 | 5/1/02 | 329.05 | 151 | 0.458902 | 329.05 | 151 | 0.458902 |
| 5 | 5/2/02 | 5/2/02 | 374.18 | 192 | 0.513120 | 374.18 | 192 | 0.513120 |
| 6 | 5/3/02 | 5/5/02 | 1783.45 | 1033 | 0.579215 | 1783.45 | 1033 | 0.579215 |
| 7 | 5/6/02 | 5/6/02 | 392.50 | 213 | 0.542672 | 392.50 | 213 | 0.542672 |
| 8 | 5/7/02 | 5/8/02 | 469.93 | 233 | 0.495823 | 469.93 | 233 | 0.495823 |
| 9 | 5/9/02 | 5/14/02 | 541.62 | 240 | 0.443119 | 541.62 | 240 | 0.443119 |
| 10 | 5/15/02 | 5/24/02 | 762.01 | 295 | 0.387133 | 762.01 | 295 | 0.387133 |
| 11 | 5/25/02 | 5/27/02 | 115.51 | 51 | 0.441514 | 115.51 | 51 | 0.441514 |
| 12 | 5/28/02 | 5/30/02 | 98.87 | 25 | 0.252866 | 98.87 | 25 | 0.252866 |
| 13 | 5/31/02 |  | 57.33 | 20 | 0.348837 | 88.33 | 20 | 0.226415 |

Table B.3. Weighted Stepwise Logistic Analysis of Year 2002 Daily Detection Efficiencies leading to Stratified McNary Detection Efficiency Estimates (weight is estimated number of downstream dam detections associated with McNary date of detection)

| Source |  |  | Deviance (Dev) | Degrees of Freedom (DF) | Dev/DF | F | Type 1 P |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Step | $\begin{array}{r} \text { Stratum } \\ \text { Julian } \end{array}$ | $\begin{aligned} & \mathrm{g} \mathrm{McNa} \\ & \text { Dates } \end{aligned}$ |  |  |  |  |  |
| Step 1 | 119 | 120 | 77.42 | 1 | 77.4200 | 31.28 | 0.0000 |
| Step 2 | 128 | 129 | 88.67 | 1 | 88.6700 | 85.40 | 0.0000 |
| Step 3 | 122 | 123 | 11.33 | 1 | 11.3300 | 13.11 | 0.0006 |
| Step 4 | 147 | 148 | 9.48 | 1 | 9.4800 | 13.25 | 0.0006 |
| Step 5 | 126 | 127 | 9.2 | 1 | 9.2000 | 16.24 | 0.0002 |
| Step 6 | 134 | 135 | 3.28 | 1 | 3.2800 | 6.33 | 0.0148 |
| Step 7 | 111 | 112 | 7.49 | 1 | 7.4900 | 19.14 | 0.0001 |
| Step 8 | 121 | 122 | 2.06 | 1 | 2.0600 | 5.72 | 0.0203 |
| Step 9 | 125 | 126 | 1.75 | 1 | 1.7500 | 5.24 | 0.0261 |
| Step 10 | 144 | 145 | 1.23 | 1 | 1.2300 | 3.88 | 0.0542 |
| Step 11 * | 150 | 151 | 1.61 | 1 | 1.6100 | 5.52 | 0.0227 |
| Step 12 | 115 | 116 | 0.84 | 1 | 0.8400 | 2.99 | 0.0898 |
| Step 13 ** | 109 | 110 | 0.68 | 1 | 0.6800 | 2.50 | 0.1206 |
| ERROR |  |  |  |  |  |  |  |
| Error | Variation | partitio | 228.39 | 62 | 3.6837 |  |  |
| Error For Step 1 | 119 | 120 | 150.97 | 61 | 2.4749 |  |  |
| Error For Step 2 | 128 | 129 | 62.3 | 60 | 1.0383 |  |  |
| Error For Step 3 | 122 | 123 | 50.97 | 59 | 0.8639 |  |  |
| Error For Step 4 | 147 | 148 | 41.49 | 58 | 0.7153 |  |  |
| Error For Step 5 | 126 | 127 | 32.29 | 57 | 0.5665 |  |  |
| Error For Step 6 | 134 | 135 | 29.01 | 56 | 0.5180 |  |  |
| Error For Step 7 | 111 | 112 | 21.52 | 55 | 0.3913 |  |  |
| Error For Step 8 | 121 | 122 | 19.46 | 54 | 0.3604 |  |  |
| Error For Step 9 | 125 | 126 | 17.71 | 53 | 0.3342 |  |  |
| Error For Step 10 | 144 | 145 | 16.48 | 52 | 0.3169 |  |  |
| Error For Step 11 | 150 | 151 | 14.87 | 51 | 0.2916 |  |  |
| Error For Step 12 | 115 | 116 | 14.03 | 50 | 0.2806 |  |  |
| Error For Step 13 | 109 | 110 | 13.35 | 49 | 0.2724 |  |  |

[^24]Table B.4. Weighted Logistic Analysis of Variation of Year 2002 Effects of Strata, Spill Proportion of Flow, and Turbine Proportion of Flow on McNary Daily Detection Efficiencies (weight is estimated number of downstream dam detections associated with McNary date of detection)

| Source | Deviance <br> (Dev) | Degrees of <br> Freedom <br> (DF) | Dev/DF | F-Ratio | Type 1 P |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Spill Proportion (Spill) | 127.88 | 1 | 127.88 | 446.09 | 0.0000 |
| Turbidity Proportion (Turb) | 129.03 | 1 | 129.03 | 450.10 | 0.0000 |
| Spill, Turb | 132.24 | 2 | 66.12 | 230.65 | 0.0000 |
| Turb adjusted for Spill | 4.36 | 1 | 4.36 | 15.21 | 0.0003 |
| Spill adjusted for Turb | 3.21 | 1 | 3.21 | 11.20 | 0.0016 |
| Among Strata (Strata) | 214.36 | 12 | 17.86 | 62.31 | 0.0000 |
| Spill adjusted for Strata | 0.27 | 1 | 0.27 | 0.94 | 0.3367 |
| Turbidity adjusted for Strata | 0.27 | 1 | 0.27 | 0.94 | 0.3367 |
| Spill, Turb adjusted for Strata | 0.27 | 2 | 0.14 | 0.47 | 0.6273 |
| Spill adjusted for Strata, Turb | 0 | 1 | 0.00 | 0.00 | 1.0000 |
| Turb adjusted for Strata, Spill | 0 | 1 | 0.00 | 0.00 | 1.0000 |
| Strata adjusted for Spill | 86.48 | 11 | 7.86 | 27.42 | 0.0000 |
| Strata adjusted for Turb | 85.6 | 12 | 7.13 | 24.88 | 0.0000 |
| Strata adjusted for Spill, Turb | 82.39 | 12 | 6.87 | 23.95 | 0.0000 |
| Error | 13.76 | 48 | 0.29 |  |  |

Table B.5. Data $^{24}$ used in Table B. 4 Analysis

| McNary <br> Date | Estimated <br> Detection <br> Efficiency | Estimated Down-Stream Dam Detections (weight) | Stratum <br> Number | Spill Proportion of McNary Flow | Turbine Proportion of McNary Flow |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 04-Apr-02 | 0.2000 | 5.0 | 1 | 0.0000 | 0.9614 |
| 06-Apr-02 | 0.3695 | 2.7 | 1 | 0.0000 | 0.9614 |
| 07-Apr-02 | 0.3544 | 2.8 | 1 | 0.0000 | 0.9601 |
| 10-Apr-02 | 0.3418 | 5.9 | 1 | 0.1856 | 0.7839 |
| 13-Apr-02 | 0.4000 | 5.0 | 1 | 0.3326 | 0.6454 |
| 14-Apr-02 | 0.2941 | 3.4 | 1 | 0.3782 | 0.5982 |
| 15-Apr-02 | 0.2449 | 4.1 | 1 | 0.5039 | 0.4785 |
| 16-Apr-02 | 0.2993 | 43.4 | 1 | 0.4736 | 0.5117 |
| 17-Apr-02 | 0.3206 | 84.2 | 1 | 0.4773 | 0.5011 |
| 18-Apr-02 | 0.3168 | 85.2 | 1 | 0.4897 | 0.4947 |
| 19-Apr-02 | 0.3336 | 170.8 | 1 | 0.4668 | 0.5173 |
| 20-Apr-02 | 0.3540 | 130.0 | 1 | 0.4738 | 0.5096 |
| 21-Apr-02 | 0.3535 | 116.0 | 1 | 0.4430 | 0.5393 |
| 22-Apr-02 | 0.3847 | 119.6 | 2 | 0.4375 | 0.5446 |
| 23-Apr-02 | 0.3866 | 119.0 | 2 | 0.4797 | 0.5032 |
| 24-Apr-02 | 0.3840 | 127.6 | 2 | 0.5027 | 0.4791 |
| 25-Apr-02 | 0.3935 | 205.8 | 2 | 0.4668 | 0.5137 |
| 26-Apr-02 | 0.4204 | 130.8 | 3 | 0.4380 | 0.5418 |
| 27-Apr-02 | 0.4023 | 101.9 | 3 | 0.4572 | 0.5199 |
| 28-Apr-02 | 0.4332 | 92.3 | 3 | 0.4117 | 0.5632 |
| 29-Apr-02 | 0.4126 | 96.9 | 3 | 0.3663 | 0.6120 |
| 30-Apr-02 | 0.4578 | 107.0 | 4 | 0.3885 | 0.5906 |
| 01-May-02 | 0.4594 | 222.0 | 4 | 0.3609 | 0.6169 |
| 02-May-02 | 0.5131 | 374.2 | 5 | 0.2337 | 0.7444 |
| 03-May-02 | 0.5702 | 477.0 | 6 | 0.2626 | 0.7166 |
| 04-May-02 | 0.5878 | 718.0 | 6 | 0.2852 | 0.6952 |
| 05-May-02 | 0.5761 | 588.5 | 6 | 0.2615 | 0.7169 |
| 06-May-02 | 0.5427 | 392.5 | 7 | 0.3057 | 0.6721 |
| 07-May-02 | 0.5106 | 270.3 | 8 | 0.3644 | 0.6162 |
| 08-May-02 | 0.4758 | 199.7 | 8 | 0.3171 | 0.6639 |
| 09-May-02 | 0.4616 | 136.5 | 9 | 0.3627 | 0.6140 |
| 10-May-02 | 0.4363 | 155.8 | 9 | 0.3666 | 0.6095 |
| 11-May-02 | 0.4526 | 90.6 | 9 | 0.3229 | 0.6545 |
| 12-May-02 | 0.4488 | 44.6 | 9 | 0.3248 | 0.6451 |
| 13-May-02 | 0.4174 | 50.3 | 9 | 0.3637 | 0.6143 |
| 14-May-02 | 0.4231 | 63.8 | 9 | 0.3208 | 0.6568 |
| 15-May-02 | 0.3923 | 61.2 | 10 | 0.2933 | 0.6844 |
| 16-May-02 | 0.3651 | 131.5 | 10 | 0.2966 | 0.6800 |
| 17-May-02 | 0.3790 | 142.5 | 10 | 0.3161 | 0.6622 |
| 18-May-02 | 0.3813 | 94.4 | 10 | 0.3351 | 0.6423 |
| 19-May-02 | 0.3991 | 100.2 | 10 | 0.3391 | 0.6384 |
| 20-May-02 | 0.4132 | 50.8 | 10 | 0.3231 | 0.6561 |
| 21-May-02 | 0.4482 | 42.4 | 10 | 0.3906 | 0.5924 |
| 22-May-02 | 0.4284 | 37.3 | 10 | 0.3987 | 0.5838 |
| 23-May-02 | 0.3483 | 51.7 | 10 | 0.4150 | 0.5678 |
| 24-May-02 | 0.3802 | 50.0 | 10 | 0.4178 | 0.5645 |
| 25-May-02 | 0.4408 | 59.0 | 11 | 0.4264 | 0.5541 |
| 26-May-02 | 0.4484 | 40.1 | 11 | 0.4306 | 0.5491 |
| 27-May-02 | 0.4272 | 16.4 | 11 | 0.4461 | 0.5364 |
| 28-May-02 | 0.2956 | 13.5 | 12 | 0.4158 | 0.5672 |
| 29-May-02 | 0.2598 | 42.3 | 12 | 0.4518 | 0.5322 |
| 30-May-02 | 0.2326 | 43.0 | 12 | 0.4164 | 0.5676 |
| 31-May-02 | 0.2857 | 3.5 | 13 | 0.4638 | 0.5216 |
| 01-Jun-02 | 0.3000 | 6.7 | 13 | 0.4953 | 0.4910 |
| 02-Jun-02 | 0.2158 | 23.2 | 13 | 0.4309 | 0.5535 |
| 04-Jun-02 | 0.5714 | 3.5 | 13 | 0.5391 | 0.4478 |
| 06-Jun-02 | 0.4000 | 2.5 | 13 | 0.5658 | 0.4217 |
| 08-Jun-02 | 0.3333 | 9.0 | 13 | 0.5739 | 0.4134 |
| 14-Jun-02 | 1.0000 | 1.0 | 13 | 0.4303 | 0.5530 |
| 15-Jun-02 | 0.6667 | 3.0 | 13 | 0.4404 | 0.5442 |
| 16-Jun-02 | 0.5000 | 2.0 | 13 | 0.3618 | 0.6209 |
| 17-Jun-02 | 0.5000 | 2.0 | 13 | 0.4739 | 0.5103 |
| 01-Jul-02 | 1.0000 | 1.0 | 13 | 0.4519 | 0.5328 |

[^25]
# APPENDIX D 

Annual Report:
Smolt Survival to McNary of Year- 2002 Coho and Fall Chinook Releases into the Yakima Basin

# IntSTATS 

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# Annual Report: <br> Smolt Survival to McNary of Year-2002 Coho and <br> Fall Chinook Releases into the Yakima Basin 

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Submitted April 1, 2003

## 1. Summary

Coho: As in years 1999, 2000, and 2001, there were early and late releases of Coho into Yakima tributaries in $2002{ }^{1}$. And, as in 2001, there were releases from Willard Hatchery (Willard) and Yakima-return (Yakima) broodstock.

For Yakima-brood smolt released into the Naches Rivers in 2002, those released on May 6 had a significantly lower $(P=0.029)$ survival index to McNary Dam (25\%) than those released on May 28 (60\%). However, for Willard-brood smolt released into the Upper Yakima and Naches, the difference between the early and late survival indices ( $20 \%$ and $24 \%$, respectively) was not significant ( $\mathrm{P}=0.620$ ). In previous years of release (1999, 2000, and 2001) there were no significant differences between early and late releases $(\mathrm{P}>0.20)$.

The above time-of-release x stock interaction is also manifested

1) by the survival indices of late-released Yakima having a significantly higher survival index then that of late-released Willard Stock (respective survival indices of $60 \%$ and $27 \%$; $\mathrm{P}=0.036$ ), but
2) by the early released Yakima and Willard stock survival indices not differing significantly (respective survival indices of $25 \%$ and $27 \% ; \mathrm{P}=$ $0.828)$.

Outmigration- year 2001 was the only other release year when Willard was used, and in that year, the survival index of Yakima stock was significantly greater than that of Willard stock at the $10 \%$ level over both release times (respective survival indices of $19 \%$

[^26]and 7\%; $\mathrm{P}<0.001$ ). In outmigration- year 1999, Yakima and Cascade stock were released, and the Yakima survival index was significantly less than the Cascade stock at the $10 \%$ level over early and late releases (respective survival indices of $43 \%$ and $54 \%$; P $=0.058$ ). In outmigration year 2000, no hatchery stock was available, and only Yakima stock was used.

In 2002 Willard-brood McNary Dam passage was generally later than Yakimabrood passage. The passage distributions of the Willard stock early and late releases were more similar than those for the Yakima stock.

Fall Chinook: As in years 2000 and 2001, there were replicated releases in $2002^{2}$ below Prosser Dam of fish experiencing accelerated rearing and conventional rearing. There were also replicated releases made into Marion drain. Analysis of the 2002 data resulted in no significant survival-index differences among the releases (survival indices were $23 \%, 22 \%$, and $30 \%$ for the accelerated rearing, conventional rearing, and Marion Drain releases, respectively; $\mathrm{P}=0.374$ ). In year 2001, there were also no significant differences among survival indices (2001 survival indices 39\%, 27\%, and 30\% respectively for accelerated, conventional, and Marion Drain; $\mathrm{P}=0.480$ ), but in year 2000, the survival index of the conventional reared significantly exceed those of both the accelerated rearing $(\mathrm{P}=0.033)$ and the Marion drain releases $(\mathrm{P}=0.025)$ [2000 survival indices $43 \%, 82 \%$ and $27 \%$ respectively for accelerated, conventional, and Marion Drain].

## 2. 2002 Coho Survival to McNary Dam and McNary Dam Passage

## 2.a. Tagging to McNary Passage Survival

In year 2002, Coho releases were made at two sites in the Naches (Lost Creek and Stiles) and at one site in the Upper Yakima (Easton). Cle Elum, which served as a second site in the Upper Yakima in previous years, was dropped in 2002. The releases were treatments comprised of two factors with primarily two levels each. The factors were time of release (early release on May 6 and late release on May 25) and stock (Willard and Yakima returns) as an intended complete factorial at each site. There was also an augmented very early release on March 28 at Easton of Willard stock. The intended May 6 early-release of Yakima stock at Easton was erroneously released on March 28. The late release at Easton of the intended Yakima stock was comprised of mixed stock ${ }^{3}$. A portion of all of these releases were PIT-tagged, and survival was estimated using the PIT-tagged fish.

The PIT-tagged releases made in 2002 are listed in Table 2.a. along with the number of fish PIT-tagged per release and the estimated PIT-tagging-to-McNary survival indices.

[^27]The methodology for estimating the survival indices is discussed in Appendix A along with the analysis procedures followed.

Table 2.a. Year 2002 Coho Releases made into the Yakima Basin with Associated Number of Fish PIT-tagged and Estimated Survival Index from Tagging to McNary Passage

| Release | Reason for <br> Omission | Number <br> PIT-tagged | Release-to-McNary <br> Survival Index | File Name <br> Extender |
| ---: | :---: | :---: | :---: | :---: |
| Site, Stock, Release Date* | Released during transport before |  |  |  |
| "Easton", Willard, "Early" | reaching Easton | 1251 | 0.0559 | MBT |
| Easton, Willard, Early |  | 1248 | 0.0580 | EWE |
| Easton, Willard, Late |  | 2497 | 0.1971 | EWL |
| Easton, Yakima, "Early" | Release is Extra Early | 1249 | 0.0154 | EYE |
| Easton, "Yakima", Late** | Actually Mixed Stock | 2500 | 0.1100 | EYL |
| Lost Creek, Willard, Early |  | 1249 | 0.2492 | LWE |
| Lost Creek, Willard, Late |  | 1247 | 0.1317 | LWL |
| Lost Creek, Yakima, Early** |  | 1192 | 0.2297 | LYE |
| Lost Creek, Yakima, Late** |  | 1250 | 0.4002 | LYL |
| Stiles, Willard, Early |  | 1249 | 0.2928 | SWE |
| Stiles, Willard, Late |  | 1251 | 0.4153 | SWL |
| Stiles, Yakima, Early |  | 1250 | 0.2688 | SYE |
| Stiles, Yakima, Late |  | 1250 | 0.8059 | SYL |

*Release Date: Extra Early - March 28
Early - May 6
Late - May 25
** Excessive pre-release escape
Omitted from formal analyses are the following treatments from Table 2.a:

1. "Easton" Site, Willard Brood, "Early" Release (shaded in Table 2.a.)--Omitted because this release was actually dumped on March 28 at Mabton on the Lower Yakima River during its transfer from Willard Hatchery to Easton. Therefore the release site was not Easton and the release date was not May 6, and the fish were presumably neither sufficiently smoltified nor acclimated.
2. Easton Site, Yakima Brood, "Early" Release (shaded in Table 2.a.)--Omitted because the release was at the wrong time (extra early on March 28 instead of early on May 6).
3. Easton Site, "Yakima" Brood, Late Release (shaded in Table 2.a.)--Omitted because the release was made up of mixed brood.

The last of the above treatments will be discussed informally after discussing the formal analysis.

A weighted logistic analysis of variation was performed using the number of fish released as the weighting variable (discussed in Appendix A). In previous years analyses, there was no statistical evidence of two factor interactions between site and
release time, between site and stock, or between stock and release time ( $\mathrm{P}>0.2$; refer to Appendix B). This was not the case for the 2002 outmigrants. As can be seen in the upper portion of Table 2.b., all main effects and two-factor interactions were significant or nearly significant the $10 \%$ significance level except for the two-factor interaction of stock with site. Because the error source, against which the main effects and two-factor interactions were tested, was based on only one degree of freedom, the decision was made to pool all two-way interactions that were not significant at the $10 \%$ significance level with the one-degree-of-freedom error as a new source of error (lower portion of Table 2.b.). All main effects as well as the Stock x Release-Time interactions were significant at the $10 \%$ significance level when using this pooled error's measure of variation (error mean deviance, analogous to error mean square from traditional analyses of variation).

Table 2.b. Weighted Logistic Analysis of Variation among Sites (Easton, Lost Creek, and Stiles), Stock (Willard and Yakima), and Release Times (May 6 and May 25)

|  | Source <br> Deviance <br> (Dev) | Megrees of <br> Freedom (DF) | Mean <br> Deviance <br> (Dev/DF) | F-Ratio | P |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Site (adjusted for Stock and Time) | 676.85 | 2 | 338.43 | 137.57 | 0.0602 |
| Stock (adjusted for Site and Time) | 294.33 | 1 | 294.33 | 119.65 | 0.0580 |
| Time* (adjusted for Site and Stock) | 505.41 | 1 | 505.41 | 205.45 | 0.0443 |
| Site x Stock | 0.66 | 1 | 0.66 | 0.27 | 0.6957 |
| Site x Time | 214.10 | 2 | 107.05 | 43.52 | 0.1066 |
| Stock x Time | 261.57 | 1 | 261.57 | 106.33 | 0.0615 |
| Error(1) | 2.46 | 1 | 2.46 |  |  |
| Site (adjusted for Stock and Time) | 676.85 | 2 | 338.43 | 6.23 | 0.0590 |
| Stock (adjusted for Site and Time) | 294.33 | 1 | 294.33 | 5.42 | 0.0804 |
| Time* (adjusted for Site and Stock) | 505.41 | 1 | 505.41 | 9.31 | 0.0380 |
| Stock x Time | 261.57 | 1 | 261.57 | 4.82 | 0.0932 |
| Error(2) | 217.22 | 4 | 54.31 |  |  |

*Time of release
In the likely presence of a time-of-release interaction with stock, main effect comparisons among times of release and among stock are not meaningful. Instead, statistical comparisons are made between:

1. The Early and Late release times within the different stock
2. The stock within release times.

Because the initial analysis (based on an error with only one degree of freedom) resulted in Stock x Time interaction being nearly significant at the $10 \%$ level, the decision was made to also make the following comparison:
3. The Early and Late release times within the different sites.

For all the comparisons, the pooled four-degree-of-freedom error measure of variation in the lower portion of Table 2.b. was used as the statistical base for comparing survival indices.

Early Release versus Late Release: For time-of-release comparisons within stock, refer to Table 2.c.1). For the Yakima stock, the late-release survival index significantly exceeded that of the early-release ( $\mathrm{P}=0.029$ ), but the time-of-release difference for the Willard stock was not significant $(\mathrm{P}=0.620)$.

For time-of-release comparisons within site, refer to Table 2.c.2). For the Stilesrelease site, the survival index for the late release significantly exceeded that for the early release $(\mathrm{P}=0.036)$. The time-of-release difference was not significant for the other two sites $(\mathrm{P}=0.226$ for Easton and $\mathrm{P}=0.787$ for Lost Creek). Recall that there was no true Yakima release included in the analysis of the Easton releases. These within-site comparisons should be regarded as less certain than indicated by the probabilities presented because, in the original logistic analysis of variation (upper portion of Table 2.a.), the interaction of time of release with site was not significant at the $10 \%$ level ( $\mathrm{P}=$ 0.107).

In previous years (Appendix B), there were no significant differences between early and late releases nor were there significant two-factor interactions among release date, site, and stock.

Yakima Stock versus Willard Stock: Referring to Table 2.c.3), the Yakima-stock survival index to McNary significantly exceeded that of the Willard stock for Late releases $(P=0.036)$. There was no significant difference between the two stock for early releases $(\mathrm{P}=0.828)$.

Referring to Appendix B for previous year's comparisons, the main effect survival index of Yakima stock over both release dates and over all sites significantly exceeded that of the Willard stock for 2001 outmigrants ( $\mathrm{P}<0.001$ ). For 2000 outmigrants, the only broodstock available was Yakima brood. For 1999 outmigrants, the hatchery stock available was Cascade, and the main-effect survival index for that stock exceeded that for Yakima ( $\mathrm{P}=0.026$ ).

Informal Comparisons: Refer to Table 2.a. for the releases omitted from the formal analysis. Regarding the extremely early dumping on March 28 of the Willard stock at Mabton, near Sunnyside on the Lower Yakima, the survival index for this release was low but, surprisingly, barely lower than that of the early Willard release actually made on May 6 at Easton (extremely early dumped Willard stock survival index $=0.056$ and early release Willard survival index $=0.058$ at Easton). The erroneous extremely early release of the Yakima stock at Easton on March 28 had a far lower survival index than the Willard stock dumped into the lower Yakima also on March 28 (Yakima-stock March 28 release survival index $=0.015$ and dumped Willard-stock survival index $=0.056$ ). The
erroneous Yakima-stock extremely early release survival index was by far the lowest of all releases.

Table 2.c.1) For 2002 Coho Outmigrants, Comparison of May 6 (Early) and May 28 (Late) releases' McNary-passage Survival Indices within Stock

|  | Number Released | Survival Index | Logit <br> Transform* | Standard Error of Logit (SE) | Late-Early Difference (Diff) |  | Comparison in Logit Transform |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | $\begin{gathered} \text { SE } \\ \text { Logit Diff } \end{gathered}$ | $\begin{aligned} & \hline \mathrm{t}=\text { Diff/ } \\ & \text { SE(Diff) } \\ & \hline \end{aligned}$ | $\begin{gathered} \hline \text { Type-1 } \\ \text { P } \end{gathered}$ |
|  |  |  |  |  | Survival | Logit |  |  |  |
| Easton, Lost Creek, Stiles Pooled |  |  |  |  | 0.0354 | 0.2079 | 0.3878 | 0.5361 | 0.6203 |
| Early w/in Willard | 3746 | 0.2000 | -1.38603 | 0.3002 |  |  |  |  |  |
| Late w/in Willard | 4995 | 0.2354 | -1.1781 | 0.2455 |  |  |  |  |  |
| Lost Creek, Stiles Pooled |  |  |  |  | 0.3533 | 1.5183 | 0.4572 | 3.3205 | 0.0294 |
| Early w/in Yakima | 2442 | 0.2497 | -1.10002 | 0.3441 |  |  |  |  |  |
| Late w/in Yakima | 2500 | 0.6031 | 0.41825 | 0.3011 |  |  |  |  |  |

Table 2.c.2) For 2002 Coho Outmigrants, Comparison of May 6 (Early) and May 28 (Late) releases' McNary-passage Survival Indices within Site

|  | Number Released | Survival Index | Logit <br> Transform* | Standard <br> Error of <br> Logit (SE) | Late-Early Difference (Diff) |  | Comparison in Logit Transform |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | $\begin{gathered} \text { SE } \\ \text { Logit Diff } \\ \hline \end{gathered}$ | $\begin{aligned} & \hline \mathrm{t}=\text { Diff/ } \\ & \text { SE(Diff) } \end{aligned}$ | $\begin{gathered} \hline \text { Type-1 } \\ \text { P } \\ \hline \end{gathered}$ |
|  |  |  |  |  | Survival | Logit |  |  |  |
| Lost Creek, Stiles Pooled |  |  |  |  |  |  |  |  |  |
| Willard within Early | 2498 | 0.2710 | -0.98947 | 0.3320 | -0.0213 | -0.1106 | 0.4781 | -0.2312 | 0.8285 |
| Yakima within Early | 2442 | -0.2497 | -1.10002 | 0.3441 |  |  |  |  |  |
| Lost Creek, Stiles Pooled |  |  |  |  |  |  |  |  |  |
| Willard within Late | 2498 | 0.2737 | -0.97582 | 0.3304 | 0.3293 | 1.3941 | 0.4470 | 3.1185 | 0.0356 |
| Yakima within Late | 2500 | 0.6031 | 0.41825 | 0.3011 |  |  |  |  |  |

Table 2.c.3) For 2002 Coho Outmigrants, Comparison of Willard- and YakimaStock Releases' McNary-passage Survival Indices within early and late time of release

|  | Number Released | Survival Index | Logit <br> Transform* | Standard Error of Logit (SE) | Late-Early Difference (Diff) |  | Comparison in Logit Transform |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | $\begin{gathered} \hline \text { SE } \\ \text { Logit Diff } \\ \hline \end{gathered}$ | $\begin{aligned} & \hline \mathrm{t}=\text { Diff/ } \\ & \mathrm{SE} \text { (Diff) } \\ & \hline \end{aligned}$ | $\begin{gathered} \hline \text { Type-1 } \\ \text { P } \end{gathered}$ |
|  |  |  |  |  | Survival | Logit |  |  |  |
| Willard Stock |  |  |  |  |  |  |  |  |  |
| Early within Easton | 1248 | 0.0580 | -2.78797 | 0.8926 | 0.1391 | 1.3831 | 0.9665 | 1.4311 | 0.2256 |
| Late within Easton | 2497 | 0.1971 | -1.40483 | 0.3707 |  |  |  |  |  |
| W Willard, Yakima Pooled |  |  |  |  |  |  |  |  |  |
| Early within Lost Creek | 2441 | 0.2397 | -1.1542 | 0.3496 | 0.0264 | 0.1399 | 0.4832 | 0.2895 | 0.7866 |
| Late within Lost Creek | 2497 | 0.2661 | -1.0144 | 0.3337 |  |  |  |  |  |
| Willard, Yakima Pooled |  |  |  |  |  |  |  |  |  |
| Early within Stiles | 2499 | 0.2808 | -0.9405 | 0.3279 | 0.3297 | 1.3900 | 0.4458 | 3.1178 | 0.0356 |
| Late within Stiles | 2501 | 0.6105 | 0.44947 | 0.3020 |  |  |  |  |  |

[^28]
## 2.b. McNary Passage Summaries

Smolt Leakage from Rearing Ponds: As in previous years, there is evidence of fish leaving the rearing ponds prior to the release date (referred to here as leakage). The measure used for this leakage is the proportion of the McNary PIT-tagged passage that was detected on or before the release date. These proportions are presented in Table 2.d. for the various releases. For the Willard 2002 outmigrants, the evidence of leakage was confined to the late releases with some fish from each of the three release sites being detected at McNary on or before the late release date (May 25). In the case of Easton late releases of Willard stock, an estimated $14 \%$ of the PIT-tagged passage was detected on or before May 25. In the case of Yakima stock, there was evidence of leakage for both early and late releases ${ }^{4}$. For three of the Yakima releases, the proportion of PIT-tagged passage detected before release date exceeded $25 \%$ ( $39 \%$ for the late Easton release ${ }^{5}$; $41 \%$ for the early Lost Creek release; and $27 \%$ for the late Lost Creek release). For all Yakima late releases, some of the fish passed McNary on or before the early release date (May 6). The lack of evidence of leakage for early release Willard stock may be due to these fish leaving the Yakima later than Yakima stock rather than due to lack of leakage; evidence for such late outmigration of Willard stock is presented below.

Passage time: Table 2.e. presents the estimated Julian dates of quartile passage ${ }^{6}$ of PIT-tagged fish for each of the releases. There are several points worth making:

1. Passage of Willard brood is generally later than that of Yakima brood. The exception is the late releases from Stiles where the Julian dates of $25 \%, 50 \%$, and $75 \%$ are almost identical for the Willard and Yakima brood.
2. Although the early releases ge nerally have earlier passage than the late releases, the differences in quartile passage dates between the early and late releases for the Willard brood are substantially less than those for the Yakima brood.
3. When comparing within release pairs, the difference in the Julian dates for $75 \%$ passage between early and late releases is substantially less than those for the other quartile passages ${ }^{7}$. This suggests that, within a stock, some of the fish tend to hold in the Yakima whether the releases were made on May 6 or May 25.
[^29]Table 2.d. Proportion of PIT-Tagged Fish passing McNary on or before Designated Release Date (bold faced proportions).

| Easton (Upper Yakima) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Willard Stock |  |  | Yakima Stock |  |
| Release Date | Cumulative Proportion Detected at McNary on or before "Release" Date |  |  |  |  |
| 03/28/02 (extra early) | 0.0000 | 0.0000 |  | 0.0000 |  |
| 05/06/02 (early) | 0.9412 | 0.0000 | 0.0000 | 0.7506 | 0.0225 |
| 05/25/02 (late) | 1.0000 | 0.4630 | 0.1434 | 1.0000 | 0.3933 |
| Lost Creek (Naches) |  |  |  |  |  |
|  | Willard Stock |  |  | Yakima Stock |  |
| Release Date | Cumulative Proportion Detected at McNary on or before "Release" Date |  |  |  |  |
| 03/28/02 (extra early) | 0.0000 -------- |  |  | 0.0000 |  |
| 05/06/02 (early) |  | 0.0000 | 0.0000 | 0.4058 | 0.0535 |
| 05/25/02 (late) |  | 0.0743 | 0.0719 | 0.7773 | 0.2674 |
| Stiles (Naches) |  |  |  |  |  |
|  | Willard Stock |  |  | Yakima Stock |  |
| Release Date | Cumulative Proportion Detected at McNary on or before "Release" Date |  |  |  |  |
|  |  | Early | Late | Early | Late |
| 03/28/02 (extra early) | 0.0000 |  |  | 0.0000 |  |
| 05/06/02 (early) |  | 0.0000 | 0.0000 | 0.0061 | 0.0082 |
| 05/25/02 (late) |  | 0.5929 | 0.0276 | 0.9164 | 0.0309 |

* Actually Extra Early Release
** Actually mixed Willard and Yakima broodstock

Table 2.e. Estimated Julian Date for $\mathbf{2 5 \%}$, $\mathbf{5 0 \%}$, and $\mathbf{7 5 \%}$ of Total McNary Smolt Passage of PIT-tagged Fish for each 2002 Coho Release into the Yakima


[^30]
## 3. 2002 Fall Chinook Survival to McNary Dam and McNary Dam Passage

## 3.a. Tagging to McNary Passage Survival

There were three major Fall Chinook release groups made into the Yakima River in 2002:

1. Below Prosser Dam release, accelerated rearing
2. Below Prosser Dam release, conventional rearing
3. Marion Drain release

There were two releases (replicates) for each of the release groups.
Table 3.a. presents the number of fish PIT-tagged per release and the estimated PIT-tagging-to-McNary survival indices. The group survival index means, associated logit transforms, and their standard errors are also presented for each of the three release groups. The standard errors utilize the error mean deviance given in the logistic analysis of variation table (Table 3.b.). The methodology for estimating the survival indices is discussed in Appendix A along with the analysis procedures followed.

As can be seen from the logistic analysis of variation table (Table 3.b.). There were no significant differences among the survival indices for the three groups $(\mathrm{P}=0.374)$.

Table 3.a. Year 2002 Fall Chinook Releases made into the Yakima Basin with Associated Number of Fish Released and Estimated Survival Index from Tagging to McNary Passage

| Release | $\begin{gathered} \hline \text { Date } \\ \text { Released } \\ \hline \end{gathered}$ | Number PIT-tagged | Survival Index |
| :---: | :---: | :---: | :---: |
| Accelerated Rearing |  |  |  |
| Replication 1 | 04/15/02 | 1001 | 0.2331 |
| Replication 2 | 04/16/02 | 1000 | 0.2286 |
| Pooled Mean |  | 2001 | 0.2308 |
| Conventional Rearing |  |  |  |
| Replication 1 | 05/15/02 | 1000 | 0.2662 |
| Replication 2 | 05/16/02 | 1000 | 0.1791 |
| Pooled Mean |  | 2000 | 0.2227 |
| Marion Drain |  |  |  |
| Replication 1 | 04/01/02 | 500 | 0.3208 |
| Replication 2 | 04/01/02 | 500 | 0.2777 |
| Pooled Mean |  | 1000 | 0.2993 |

Table 3.b. Weighted Logistic Analysis of Variation among Release Groups (Prosser Accelerated Rearing, Prosser Conventional Rearing, and Marion Drain Releases).

|  | Deviance <br> $($ Dev $)$ | Degrees of <br> Freedom $(\mathrm{DF})$ | Mean Deviance <br> $(\mathrm{Dev} / \mathrm{DF})$ | F-Ratio | Type 1 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Source | 22.54 | 2 | 11.27 | 1.39 | 0.3739 |
| Release Group | 3 | 8.11 |  |  |  |
| Error | 24.32 | 3 |  |  |  |

Referring to Appendix B, there were also no significant differences among similar releases made in year 2001 ( $\mathrm{P}=0.480$ ); however, for releases made in year 2000, the conventional rearing survival index significantly exceeded that of the accelerated ( $\mathrm{P}=$ 0.033 ) and also exceeded that of fish released at Marion Drain $(\mathrm{P}=0.025)$.

## 3.b. McNary Passage Summaries

Smolt Leakage from Rearing Ponds: Based on McNary detections, there was no evidence of fish leakage from any of the six Fall Chinook releases; i.e., none of the PITtagged fish were detected at McNary before the associated release date.

Passage time: Table 3.c. presents the estimated Julian dates of quartile passage of PIT-tagged fish for each of the releases. As might be expected, the conventional rearing treatment, which was released in mid-May, passed later than that of the accelerated rearing treatment, which was released in mid-April, a month earlier than the conventionally reared fish. The first quartile passage of the Marion Drain release, which was released on April 1, was comparable to that of the Prosser released conventionally reared fish, which were released on May 15 and 16. However, the passage of the Marion Drain release was more compressed than that of the Prosser releases. There is between 4 and 5 days separating the $25 \%$ and $75 \%$ McNary passage dates for the Marion Drain releases. For the below-Prosser Dam releases, the differences between the 25\% and 75\% passage dates were 12 and 16 days for the accelerated treatment and were 7 and 15 days for the conventional treatment.

Table 3.c. Estimated Julian Date for $\mathbf{2 5 \%}$, $\mathbf{5 0 \%}$, and $\mathbf{7 5 \%}$ of Total McNary Smolt Passage of PIT-tagged Fish for each 2002 Fall Chinook Release into the Yakima

| Quartile (when \% passage attained) | Julian Date when \% Passge AttainedBelow-Prosser Dam Release |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Accelerated Rearing | Conventional Rearing |  | Marion Drain Releases |  |
|  | Release 1 Release 2 | Release 1 | Release 2 | Release 1 | Release 2 |
| Q1 (25\% passage) | 140 138 | 152 | 153 | 150 | 151 |
| Q2 (50\% passage) | 150 144 | 156 | 157 | 153 | 153 |
| Q3 (75\% passage) | 152 : 154 | 159 | 168 | 155 | 155 |
| Estimated Total Passage | 233 - 229 | 266 | 179 | 160 | 139 |

## Appendix A. Estimated Survival Index and Logistic Analysis

A weighted logistic analysis of variation was undertaken using release number as the weighting variable. The basic nature of and justification for the use of the logistic analysis of variation is discussed in Appendix A. of 2002 Annual Report: OCT-SNT Survival by Doug Neeley.

The 2002-release estimators for Release-to-McNary survival indices are somewhat different than those for previous brood years. For the 2002 releases, McNary detection efficiencies were based on detections of all Coho and all Fall Chinook released into the Upper Yakima; in previous release years the efficiencies were based on the releases within the specific study (early and late releases for Coho, and accelerated, conventional, and Marion Drain releases for Fall Chinook). Efforts will be made next year to standardize the estimation procedures over all release years. Alternative survival estimation procedures will also be investigated.

## A.1. Coho

The total number of Coho detections per stratum, the expanded number (detected number divided by detection efficiency) per stratum, the total expanded number over strata, the total number tagged, and the survival index are given for 2002 Coho releases in Table A.1.

The detection efficiencies used for Coho are given in Table A.2. In the table, the "Augmented" includes downstream detections from days for which there were no associated joint downstream-dam and McNary Dam counts. The detection efficiencies actually used are those listed under "Augmented" under "Pooled over John Day and Bonneville" in Table A.2. The stepwise logistic analysis of variation leading to the stratification is given in Table A. 3

The stratification explained more of the variation in McNary daily detection efficiencies for Coho than was explained by the daily proportion of McNary flow that is spilled (spill proportion) or the daily proportion of flow that is passed through the turbines (turbine proportion). This is reflected in the significance levels in Table A. 4 for strata adjusted 1) for spill proportion, 2) for turbidity proportion, and 3) for both (respectively $\mathrm{P}=0.001, \mathrm{P}=0.003, \mathrm{P}=0.002$ ). However, stratification does not explain all of the variability explained by the effects of spill proportion and of turbine proportion; the effect of spill proportion and turbine proportion being significant. (Type 1error probability estimates associated with strata adjusted for spill proportion, for turbine proportion, and for both respectively are $\mathrm{P}=0.090, \mathrm{P}=0.036$, and $\mathrm{P}<0.001$, Table A.4.)

Table A. 5 gives the data used in the analyses.

Table A.1. Numbers used to estimate Survival Indices for 2002 Coho releases into the Upper Yakima and Naches

| Site > | Easton |  |  |  |  | Lost Creek |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Stock > | Willard |  | Mixed | Yakima |  | Willard |  | Yakima |  |
| Release Time > | Very Early | Early | Late | Very <br> Early | Late | Early | Late | Early | Late |
| Stratum Number of Detections |  |  |  |  |  |  |  |  |  |
| 1 Beginning - 5/8/02 | 34 | 0 | 0 | 7 | 4 | 1 | 0 | 55 | 17 |
| 2 5/9/02-5/18/02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 9 |
| 3 5/19/02-5/22/02 | 0 | 0 | 8 | 0 | 4 | 3 | 1 | 7 | 11 |
| 4 5/23/02-5/29/02 | 0 | 7 | 13 | 1 | 23 | 8 | 3 | 13 | 17 |
| 5 5/30/02 - End | 0 | 5 | 53 | 0 | 19 | 34 | 19 | 6 | 41 |
| Total over Strata | 34 | 12 | 74 | 8 | 50 | 46 | 23 | 89 | 95 |
| Detection |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |
| Stratum Efficiency |  |  |  |  |  |  |  |  |  |
| 10.48591 | 70.0 | 0.0 | 0.0 | 14.4 | 8.2 | 2.1 | 0.0 | 113.2 | 35.0 |
| 20.22148 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 36.1 | 40.6 |
| 30.44602 | 0.0 | 0.0 | 17.9 | 0.0 | 9.0 | 6.7 | 2.2 | 15.7 | 24.7 |
| 40.20896 | 0.0 | 33.5 | 62.2 | 4.8 | 110.1 | 38.3 | 14.4 | 62.2 | 81.4 |
| 50.12868 | 0.0 | 38.9 | 411.9 | 0.0 | 147.7 | 264.2 | 147.7 | 46.6 | 318.6 |
| Total Expanded Detections | 70.0 | 72.4 | 492. | 19.2 | 274.9 | 311.3 | 164.3 | 273.8 | 500.3 |
| Number Released | 1251 | 1248 | 2497 | 1249 | 2500 | 1249 | 1247 | 1192 | 1250 |
| Survival Index = <br> (Total Expanded Detections)/ <br> (Number Released) |  |  |  |  |  |  |  |  |  |
|  | 0.0559 | 0.0580 | 0.1971 | 0.0154 | 0.1100 | 0.2492 | 0.1317 | 0.2297 | 0.4002 |



Table A.2. Detection efficiencies used to estimate 2002-Outmigrant Coho Survival Indices

| Stratum | 1st Date | Last Date | Joint Detection Based |  |  | ased | Augmented BO-McN Total | Estimate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | BO Total | BO-McN Total | Estimate | BO Total |  |  |
| 1 |  | 5/8/02 | 56.00 | 36 | 0.642857 | 65.52 | 36 | 0.549481 |
| 2 | 5/9/02 | 5/18/02 | 49.66 | 13 | 0.261768 | 53.80 | 13 | 0.241631 |
| 3 | 5/19/02 | 5/22/02 | 20.23 | 10 | 0.494438 | 20.57 | 10 | 0.486149 |
| 4 | 5/23/02 | 5/29/02 | 125.52 | 25 | 0.199164 | 125.52 | 25 | 0.199164 |
| 5 | 5/30/02 |  | 298.59 | 43 | 0.144011 | 351.59 | 43 | 0.122302 |
|  |  |  | John Day-based |  |  |  |  |  |
|  |  |  | Joint Detection Based |  |  | Augmented |  |  |
| Stratum | 1st Date | Last Date | JD Total | JD-McN Total | Estimate | JD Total | JD-McN Total | Estimate |
| 1 |  | 5/16/02 | 48.33 | 26 | 0.537931 | 62.08 | 26 | 0.418818 |
| 2 | 5/17/02 | 5/26/02 | 11.90 | 3 | 0.252101 | 18.44 | 3 | 0.162690 |
| 3 | 5/27/02 | 5/30/02 | 27.97 | 12 | 0.429082 | 28.76 | 12 | 0.417316 |
| 4 | 5/31/02 | 6/6/02 | 77.08 | 18 | 0.233516 | 80.25 | 18 | 0.224290 |
| 5 | 6/7/02 |  | 350.72 | 53 | 0.151119 | 394.47 | 53 | 0.134357 |
|  |  |  | Pooled over John Day and Bonneville |  |  |  |  |  |
|  |  |  | Joint Detection Based |  |  | Augmented (Actually Used) |  |  |
| Stratum | 1st Date | Last Date | JD Total | JD-McN Total | Estimate | JD Total | JD-McN Total | Estimate |
| 1 |  | 5/24/02 | 104.33 | 62 | 0.594249 | 127.60 | 62 | 0.485910 |
| 2 | 5/25/02 | 6/3/02 | 61.56 | 16 | 0.259900 | 72.24 | 16 | 0.221481 |
| 3 | 6/4/02 | 6/7/02 | 48.19 | 22 | 0.456510 | 49.32 | 22 | 0.446021 |
| 4 | 6/8/02 | 6/14/02 | 202.61 | 43 | 0.212234 | 205.78 | 43 | 0.208963 |
| 5 | 6/15/02 |  | 649.31 | 96 | 0.147850 | 746.06 | 96 | 0.128676 |

Table A.3. Weighted Stepwise Logistic Analysis of Year 2002 Yakima-Released Coho Detection Efficiencies leading to Stratified McNary Detection Efficiency Estimates (weight is estimated number of downstream dam detections associated with McNary date of detection)

| Source |  | Deviance (Dev) | Degrees of Freedom (DF) | Dev/DF | F | Type 1 P |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Step | Stratum Separating McNary Julian Detection Dates |  |  |  |  |  |
| Step 1 | 142143 | 80.74 | 1 | 80.74 | 68.46 | 0.0000 |
| Step 2 * | 126127 | 13.91 | 1 | 13.91 | 15.22 | 0.0003 |
| Step 3 | 149150 | 4.47 | 1 | 4.47 | 5.33 | 0.0254 |
| Step 4 ** | 139140 | 3.73 | 1 | 3.73 | 4.81 | 0.0334 |
| Step 5 *** | 150151 | 1.08 | 1 | 1.08 | 1.40 | 0.2422 |
| ERROR |  | Deviance | DF | Mean Dev |  |  |
| Error | Variation with no partition | 138.53 | 50 | 2.77 |  |  |
| Error For Step 1 |  | 57.79 | 49 | 1.18 |  |  |
| Error For Step 2 |  | 43.88 | 48 | 0.91 |  |  |
| Error For Step 3 |  | 39.41 | 47 | 0.84 |  |  |
| Error For Step 4 |  | 35.68 | 46 | 0.78 |  |  |
| Error For Step 5 |  | 34.6 | 45 | 0.77 |  |  |

* Shifting of strata after all steps completed actually produced smaller error for separation
between Julian dates 128 and 129 which was the separation selected
** Partitioning produced a stratum that had less than 20 total joint McNary, down-stream dam detections, separation actually selected was between Julian dates 138 and 139
*** Omitted and terminated stepwise process because $10 \%$ significance level not attained ( $P=0.242$ )

Table A.4. Weighted Logistic Analysis of Variation of Year 2002 Effects of Strata, Spill Proportion of Flow, and Turbine Proportion of Flow on Yakima-Released Coho McNary Daily Detection Efficiencies (weight is estimated number of downstream dam detections associated with McNary date of detection)

| Source | Deviance <br> Devees of <br> (Dev) | Freedom <br> (DF) | Dev/DF | F-Ratio | Type 1 P |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Spill Proportion (Spill) | 1.12 | 1 | 1.12 | 20.97 | 0.0000 |
| Turbidity Proportion (Turb) | 1.35 | 1 | 1.35 | 25.28 | 0.0000 |
| Spill, Turb | 1.45 | 2 | 0.73 | 13.57 | 0.0000 |
| Spill adjusted for Turb | 0.1 | 1 | 0.10 | 1.87 | 0.1782 |
| Turb adjusted for Spill | 0.33 | 1 | 0.33 | 6.18 | 0.0168 |
| Among Strata (Strata) | 1.25 | 4 | 0.31 | 5.85 | 0.0007 |
| Spill adjusted for Strata | 1.03 | 1 | 1.03 | 19.29 | 0.0001 |
| Turb adjusted for Strata | 1.12 | 1 | 1.12 | 20.97 | 0.0000 |
| Spill adjusted for Strata, Turb | 0.16 | 1 | 0.16 | 3.00 | 0.0905 |
| Turb adjusted for Strata, Spill | 0.25 | 1 | 0.25 | 4.68 | 0.0360 |
| Spill, Turb adjusted for Strata | 1.28 | 2 | 0.64 | 11.98 | 0.0001 |
| Strata adjusted for Spill | 1.16 | 4 | 0.29 | 5.43 | 0.0012 |
| Strata adjusted for Turb | 1.02 | 4 | 0.26 | 4.77 | 0.0028 |
| Strata adjusted for Spill, Turb | 1.08 | 4 | 0.27 | 5.06 | 0.0019 |
| Error | 2.35 | 44 | 0.05 |  |  |

Table A.5. Data ${ }^{8}$ used in Table A. 4 Analysis

| McNary Date | Estimated <br> Detection <br> Efficiency | Estimated <br> Downstream Detections (weight) | Stratum Number | Spill <br> Propotion of McNary Flow | Turbine Propotion of McNary Flow |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 4/7/02 | 0 | 0.388888889 | 1 | 0.0000 | 0.9601 |
| 4/9/02 | 1 | 1 | 1 | 0.0000 | 0.9660 |
| 4/10/02 | 0 | 0.333333333 | 1 | 0.1856 | 0.7839 |
| 4/11/02 | 0 | 0.433333333 | 1 | 0.3572 | 0.6151 |
| 4/12/02 | 0.639593909 | 1.563492063 | 1 | 0.3779 | 0.5950 |
| 4/13/02 | 0 | 1.111111111 | 1 | 0.3326 | 0.6454 |
| 4/14/02 | 0 | 0.748917749 | 1 | 0.3782 | 0.5982 |
| 4/15/02 | 0.583333333 | 5.142857143 | 1 | 0.5039 | 0.4785 |
| 4/16/02 | 0.328125 | 3.047619048 | 1 | 0.4736 | 0.5117 |
| 4/17/02 | 0.61192053 | 4.902597403 | 1 | 0.4773 | 0.5011 |
| 4/18/02 | 0.394551433 | 2.53452381 | 1 | 0.4897 | 0.4947 |
| 4/19/02 | 0.481305707 | 6.233044733 | 1 | 0.4668 | 0.5173 |
| 4/20/02 | 0.431692769 | 9.265848966 | 1 | 0.4738 | 0.5096 |
| 4/21/02 | 0.267423015 | 3.739393939 | 1 | 0.4430 | 0.5393 |
| 4/22/02 | 0 | 2.004184704 | 1 | 0.4375 | 0.5446 |
| 4/23/02 | 0 | 1.015305942 | 1 | 0.4797 | 0.5032 |
| 4/24/02 | 0.207385684 | 4.821933622 | 1 | 0.5027 | 0.4791 |
| 4/25/02 | 0.687159147 | 1.455266955 | 1 | 0.4668 | 0.5137 |
| 4/26/02 | 0.391752577 | 2.552631579 | 1 | 0.4380 | 0.5418 |
| 4/27/02 | 0.623465601 | 3.207875458 | 1 | 0.4572 | 0.5199 |
| 4/28/02 | 0.681572304 | 4.401587302 | 1 | 0.4117 | 0.5632 |
| 4/29/02 | 0.326213592 | 3.06547619 | 1 | 0.3663 | 0.6120 |
| 4/30/02 | 0.194895592 | 5.130952381 | 1 | 0.3885 | 0.5906 |
| 5/1/02 | 0 | 0.763888889 | 1 | 0.3609 | 0.6169 |
| 5/2/02 | 0.491848074 | 10.16574074 | 1 | 0.2337 | 0.7444 |
| 5/3/02 | 0.885664592 | 5.645478036 | 1 | 0.2626 | 0.7166 |
| 5/4/02 | 0.719212937 | 11.12327044 | 1 | 0.2852 | 0.6952 |
| 5/5/02 | 0.64516129 | 9.3 | 1 | 0.2615 | 0.7169 |
| 5/6/02 | 0.475038487 | 12.63055556 | 1 | 0.3057 | 0.6721 |
| 5/7/02 | 0.375782881 | 5.322222222 | 1 | 0.3644 | 0.6162 |
| 5/8/02 | 0.4400978 | 4.544444444 | 1 | 0.3171 | 0.6639 |
| 5/9/02 | 0.139878254 | 7.149074074 | 2 | 0.3627 | 0.6140 |
| 5/10/02 | 0 | 1.962287664 | 2 | 0.3666 | 0.6095 |
| 5/11/02 | 0 | 0.846075124 | 2 | 0.3229 | 0.6545 |
| 5/12/02 | 0 | 0.924958229 | 2 | 0.3248 | 0.6451 |
| 5/13/02 | 0 | 1.235779269 | 2 | 0.3637 | 0.6143 |
| 5/14/02 | 0 | 0.848183678 | 2 | 0.3208 | 0.6568 |
| 5/15/02 | 0.220478736 | 9.071169586 | 2 | 0.2933 | 0.6844 |
| 5/16/02 | 0.14556636 | 13.73943818 | 2 | 0.2966 | 0.6800 |
| 5/17/02 | 0.305508506 | 9.819693859 | 2 | 0.3161 | 0.6622 |
| 5/18/02 | 0.300250536 | 26.6444154 | 2 | 0.3351 | 0.6423 |

[^31]Table A.5. Data used in Table A. 4 Analysis (continued)

| McNary Date | Estimated |  |  | Spill <br> Propotion of McNary Flow | Turbine Propotion of McNary Flow |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Estimated <br> Detection <br> Efficiency | Downstream Detections (weight) | Stratum <br> Number |  |  |
| 05/19/02 | 0.361620411 | 13.83 | 3 | 0.3391 | 0.6384 |
| 05/20/02 | 0.514096494 | 11.67 | 3 | 0.3231 | 0.6561 |
| 05/21/02 | 0.507502951 | 13.79 | 3 | 0.3906 | 0.5924 |
| 05/22/02 | 0.398630273 | 10.03 | 3 | 0.3987 | 0.5838 |
| 05/23/02 | 0.209957069 | 47.63 | 4 | 0.4150 | 0.5678 |
| 05/24/02 | 0.199390789 | 50.15 | 4 | 0.4178 | 0.5645 |
| 05/25/02 | 0.254258452 | 35.40 | 4 | 0.4264 | 0.5541 |
| 05/26/02 | 0.171625177 | 23.31 | 4 | 0.4306 | 0.5491 |
| 05/27/02 | 0.333333333 | 3.00 | 4 | 0.4461 | 0.5364 |
| 05/28/02 | 0.22613769 | 4.42 | 4 | 0.4158 | 0.5672 |
| 05/29/02 | 0.191064906 | 41.87 | 4 | 0.4518 | 0.5322 |
| 05/30/02 | 0.166872117 | 239.70 | 5 | 0.4164 | 0.5676 |
| 05/31/02 | 0.125669443 | 167.11 | 5 | 0.4638 | 0.5216 |
| 06/01/02 | 0.137381275 | 65.51 | 5 | 0.4953 | 0.4910 |
| 06/02/02 | 0.12189195 | 82.04 | 5 | 0.4309 | 0.5535 |
| 06/03/02 | 0.133489461 | 22.47 | 5 | 0.4460 | 0.5390 |
| 06/04/02 | 0 | 7.87 | 5 | 0.5391 | 0.4478 |
| 06/05/02 | 0.07640897 | 26.17 | 5 | 0.5876 | 0.3999 |
| 06/06/02 | 0.078428974 | 38.25 | 5 | 0.5658 | 0.4217 |
| 06/07/02 | 0.095846645 | 20.87 | 5 | 0.5340 | 0.4523 |
| 06/08/02 | 0 | 7.10 | 5 | 0.5739 | 0.4134 |
| 06/09/02 | 0 | 6.71 | 5 | 0.4915 | 0.4943 |
| 06/10/02 | 0 | 7.59 | 5 | 0.4359 | 0.5482 |
| 06/11/02 | 0 | 12.49 | 5 | 0.4980 | 0.4882 |
| 06/12/02 | 0 | 12.39 | 5 | 0.4877 | 0.4965 |
| 06/13/02 | 0 | 8.12 | 5 | 0.4664 | 0.5183 |
| 06/14/02 | 0 | 4.15 | 5 | 0.4303 | 0.5530 |
| 06/15/02 | 0.148148148 | 6.75 | 5 | 0.4404 | 0.5442 |
| 06/16/02 | 0.558139535 | 3.58 | 5 | 0.3618 | 0.6209 |
| 06/17/02 | 0.52173913 | 3.83 | 5 | 0.4739 | 0.5103 |
| 06/18/02 | 0 | 0.67 | 5 | 0.5078 | 0.4778 |
| 06/19/02 | 0 | 0.67 | 5 | 0.5191 | 0.4673 |
| 06/20/02 | 0.6 | 1.67 | 5 | 0.5226 | 0.4644 |
| 06/21/02 | 0 | 0.33 | 5 | 0.5037 | 0.4829 |

## A.2. Fall Chinook

The total number of Fall Chinook detections per stratum, the expanded number (detected number divided by detection efficiency) per stratum, the total expanded number over strata, the total number tagged, and the survival index are given for 2002 Fall Chinook releases in Table A.6.

The detection efficiencies used for Fall Chinook are given in Table A.7. In the table, the "Augmented" includes downstream detections from days for which there were no associated joint downstream-dam and McNary Dam counts. The detection efficiencies actually used are those listed under "Augmented" under "Pooled over John Day and Bonneville" in Table A.7. The stepwise logistic analysis of variation leading to the stratification is given in Table A.8.

The stratification was more effective in explaining variation in daily detection efficiencies than either the daily proportion of McNary flow that was spilled (spill proportion) or the daily proportion of flow that was passed through the turbines (turbine proportion). Table A. 9 gives a logistic analysis of variation that assesses the effects of strata, spill proportion, and turbine proportion on estimated daily McNary detection efficiencies. Table A. 10 gives the data used in the regression. The effect of strata was stronger than those of spill proportion and turbine proportion. In fact, the effects of spill and turbine proportions are not significant ( $\mathrm{P}>0.2$, Table A.9), nor are their effects when adjusted for the effect of strata. However, the effects of strata are significant, whether or not the adjusted for the effects of spill proportion, turbine proportion, and both are still highly significant ( $\mathrm{P}<0.001$, Table A.9).

Table. A.6. Numbers used to estimate Survival Indices for 2002 Fall Chinook releases into the Upper Yakima and Naches

| Release > | Below-Prosser Releases |  |  |  | Marion Drain |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Accelerated |  | Conventional |  |  |  |
|  | Release 1 | Release 2 | Release 1 | Release? | Release 1 | Release 2 |
| Stratum | Number of Detections |  |  |  |  |  |
| 1 Beginning - 5/31/02 | 48 | 50 | 20 | - 10 | 16 | 12 |
| 2 6/1/00 - 6/4/02 | 13 | 6 | 15 | 10 | 18 | 16 |
| 3 6/5/02 - End | 7 | 13 | 41 | 31 | 9 | 9 |
| Total over Strata | 68 | 69 | 76 | - 51 | 43 | 37 |
| Detection |  |  |  |  |  |  |
|  |  |  |  |  |  |  |
| Stratum Efficiency | 152.497676 | 158.851745 | 63.5406982 | - | 50.8325586 | 38.1244189 |
| 10.3148 |  |  |  | 31.7703491 |  |  |
| 20.2260 | 57.5279659 | 26.5513689 | 66.3784222 | 44.2522814 | 79.6541066 | 70.8036503 |
| $3 \quad 0.3008$ | 23.2748577 | 43.2247356 | $\left\lvert\, \frac{136.324166 \mid}{266.243287!}\right.$ | 103.07437 | 29.924817 | 29.924817 |
| Total Expanded Detections | 233.300499 | $\underline{228.62785}$ |  | 179.097 | 160.411482 | 138.852886 |
| Number Released | 1001 | 1000 | 1000 | 1000 | 500 | 500 |
| Survival Index = |  |  |  |  |  |  |
| (Total Expanded Detections)/ |  |  |  |  |  |  |
| (Number Released) | 0.2331 | 0.2286 | 0.2662 | 0.1791 | 0.3208 | 0.2777 |

Table A.7. Detection efficiencies used to estimate 2002-Outmigrant Fall Chinook Survival Indices

| Stratum | 1st Date Last Date |  | Joint Detection Based Bonnville-based |  |  |  | Augmented BO-McN Total | Estimate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | BO Total | BO-McN Total | Estimate | BO Total |  |  |
| 1 |  | 5/31/02 | 49.33 | 22 | 0.445946 | 65.74 | 22 | 0.334634 |
| 2 | 6/1/02 | 6/4/02 | 27.67 | 9 | 0.325301 | 30.78 | 9 | 0.292394 |
| 3 | 6/5/02 | 5/22/02 | 18.00 | 8 | 0.444444 | 26.93 | 8 | 0.297083 |
|  |  |  |  | int Detection Bas | John | based | Augmented |  |
| Stratum | 1st Date | Last Date | JD Total | JD-McN Total | Estimate | JD Total | JD-McN Total | Estimate |
| 1 |  | 5/16/02 | 105.81 | 34 | 0.321332 | 112.17 | 34 | 0.303110 |
| 2 | 5/17/02 | 5/26/02 | 60.83 | 12 | 0.197260 | 62.15 | 12 | 0.193083 |
| 3 | 5/27/02 | 5/30/02 | 65.36 | 24 | 0.367213 | 79.47 | 24 | 0.301998 |
|  |  |  | Pooled over John Day and Bonneville |  |  |  |  |  |
|  |  |  | Joint Detection Based |  |  | Augmented (Actually Used) |  |  |
| Stratum | 1st Date | Last Date | JD Total | JD-McN Total | Estimate | JD Total | JD-McN Total | Estimate |
| 1 |  | 5/24/02 | 155.14 | 56 | 0.360958 | 177.91 | 56 | 0.314759 |
| 2 | 5/25/02 | 6/3/02 | 88.50 | 21 | 0.237288 | 92.93 | 21 | 0.225977 |
| 3 | 6/4/02 | 6/7/02 | 83.36 | 32 | 0.383890 | 106.40 | 32 | 0.300754 |

Table A.8. Weighted Stepwise Logistic Analysis of Year 2002 Yakima-Released Fall Chinook Daily Detection Efficiencies leading to Stratified McNary Detection Efficiency Estimates (weight is estimated number of downstream dam detections associated with McNary date of detection)

| Source |  |  | Deviance (Dev) | Degrees of <br> Freedom (DF) | Dev/DF | F | Type 1 P |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Step | Stratum Separating McNary Julian Detection Dates |  |  |  |  |  |  |
| Step 1 * | 143 | 144 | 2.00 | 1 | 2 | 5.25 | 0.0291 |
| Step 2 | 155 | 156 | 2.59 | 1 | 2.59 | 8.51 | 0.0068 |
| Step 3 ** | 151 | 152 | 1.87 | 1 | 1.87 | 7.52 | 0.0105 |
| Step 3 *** | 152 | 153 | 1.56 | 1 | 1.56 | 6.01 | 0.0207 |
| Step 3 *** | 154 | 155 | 1.24 | 1 | 1.24 | 4.57 | 0.0413 |
| ERROR |  |  | Deviance | DF | Mean Dev |  |  |
| Error | Variation | arition | 13.42 | 31 | 0.43 |  |  |
| Error Step 1 * |  |  | 11.42 | 30 | 0.38 |  |  |
| Error Step 2 |  |  | 8.83 | 29 | 0.30 |  |  |
| Error Step 3 ** |  |  | 6.96 | 28 | 0.25 |  |  |
| Error Step 4 *** |  |  | 7.27 | 28 | 0.26 |  |  |
| Error Step 5 *** |  |  | 7.59 | 28 | 0.27 |  |  |

* Shifting of strata after all steps completed actually produced smaller error for separation

$$
\text { between Julian dates } 151 \text { and } 152 \text { which was the separation selected (same as first Step 3) }
$$

with elimination of Step 143 and 144 Julian date separation, first Step 3 inconsistency disappears
** Inconsistency between John Data-based and Bonneville-based detection efficiencies
*** Partitioning produced a stratum that had less than 20 total joint McNary, down-stream dam detections
Table A.9. Weighted Logistic Analysis of Variation of Year 2002 Effects of Strata, Spill Proportion of Flow, and Turbine Proportion of Flow on Yakima-Released Fall Chinook McNary Daily Detection Efficiencies (weight is estimated number of downstream dam detections associated with McNary date of detection)

|  | Degrees of <br> Deviance <br> Freedom <br> (Dev) |  |  |  | (DF) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Source | 0.34 | 1 | 0.34 | 1.27 | 0.2691 |
| Spill Proportion (Spill) | 0.33 | 1 | 0.33 | 1.24 | 0.2761 |
| Turbidity Proportion (Turb) | 0.45 | 2 | 0.23 | 0.84 | 0.4416 |
| Spill, Turb | 0.12 | 1 | 0.12 | 0.45 | 0.5083 |
| Spill adjusted for Turb | 0.11 | 1 | 0.11 | 0.41 | 0.5264 |
| Turb adjusted for Spill | 5.35 | 2 | 2.68 | 10.02 | 0.0006 |
| Among Strata (Strata) | 0.48 | 1 | 0.48 | 1.80 | 0.1912 |
| Spill adjusted for Strata | 0.5 | 1 | 0.50 | 1.87 | 0.1825 |
| Turb adjusted for Strata | 0.36 | 1 | 0.36 | 1.35 | 0.2558 |
| Spill adjusted for Strata, Turb | 0.38 | 1 | 0.38 | 1.42 | 0.2433 |
| Turb adjusted for Strata, Spill | 0.86 | 2 | 0.43 | 1.61 | 0.2184 |
| Spill, Turb adjusted for Strata | 5.49 | 2 | 2.75 | 10.28 | 0.0005 |
| Strata adjusted for Spill | 5.52 | 2 | 2.76 | 10.34 | 0.0005 |
| Strata adjusted for Turb | 5.76 | 2 | 2.88 | 10.79 | 0.0004 |
| Strata adjusted for Spill, Turb | 5.76 | 27 | 0.27 |  |  |
| Error | 7.21 |  |  |  |  |

Table A.10. Data ${ }^{9}$ used in Table B. 9 Analysis

| McNary Date | Estimated <br> Detection Efficiency | Estimated <br> Downstream <br> Detections (weight) | Stratum <br> Number | Spill <br> Propotion of McNary Flow | Turbine Propotion of McNary Flow |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 05/07/02 | 0.5 | 2.00 | 1 | 0.3644 | 0.6162 |
| 05/08/02 | 0.4 | 2.50 | 1 | 0.3171 | 0.6639 |
| 05/10/02 | 0 | 1.00 | 1 | 0.3666 | 0.6095 |
| 05/11/02 | 0 | 1.00 |  | 0.3229 | 0.6545 |
| 05/12/02 | 0 | 0.25 | 1 | 0.3248 | 0.6451 |
| 05/13/02 | 0 | 0.95 | 1 | 0.3637 | 0.6143 |
| 05/14/02 | 0 | 0.91 | , | 0.3208 | 0.6568 |
| 05/15/02 | 0.24742268 | 4.04 | 1 | 0.2933 | 0.6844 |
| 05/16/02 | 0 | 0.49 | 1 | 0.2966 | 0.6800 |
| 05/17/02 | 0.208309223 | 4.80 | 1 | 0.3161 | 0.6622 |
| 05/18/02 | 0.328295389 | 6.09 | 1 | 0.3351 | 0.6423 |
| 05/19/02 | 0.378511154 | 13.21 | 1 | 0.3391 | 0.6384 |
| 05/20/02 | 0.324559366 | 12.32 | 1 | 0.3231 | 0.6561 |
| 05/21/02 | 0.478712822 | 10.44 | 1 | 0.3906 | 0.5924 |
| 05/22/02 | 0.356422564 | 5.61 | 1 | 0.3987 | 0.5838 |
| 05/23/02 | 0.409223154 | 7.33 | 1 | 0.4150 | 0.5678 |
| 05/24/02 | 0.151538701 | 13.20 | 1 | 0.4178 | 0.5645 |
| 05/25/02 | 0.325540924 | 15.36 | 1 | 0.4264 | 0.5541 |
| 05/26/02 | 0.227757017 | 8.78 | 1 | 0.4306 | 0.5491 |
| 05/27/02 | 0.315459678 | 9.51 | 1 | 0.4461 | 0.5364 |
| 05/28/02 | 0.31700832 | 12.62 | 1 | 0.4158 | 0.5672 |
| 05/29/02 | 0.239103362 | 16.73 | 1 | 0.4518 | 0.5322 |
| 05/30/02 | 0.455840456 | 8.78 | 1 | 0.4164 | 0.5676 |
| 05/31/02 | 0.350355474 | 19.98 | 1 | 0.4638 | 0.5216 |
| 06/01/02 | 0.251927628 | 19.85 | 2 | 0.4953 | 0.4910 |
| 06/02/02 | 0.228426851 | 30.64 | 2 | 0.4309 | 0.5535 |
| 06/03/02 | 0.254602734 | 19.64 | 2 | 0.4460 | 0.5390 |
| 06/04/02 | 0.175438596 | 22.80 | 2 | 0.5391 | 0.4478 |
| 06/05/02 | 0.208742013 | 4.79 | 3 | 0.5876 | 0.3999 |
| 06/06/02 | 0.349768146 | 22.87 | 3 | 0.5658 | 0.4217 |
| 06/07/02 | 0.304477724 | 22.99 | 3 | 0.5340 | 0.4523 |
| 06/08/02 | 0.305546799 | 16.36 | 3 | 0.5739 | 0.4134 |
| 06/09/02 | 0 | 1.10 | 3 | 0.4915 | 0.4943 |
| 06/10/02 | 0 | 0.48 | 3 | 0.4359 | 0.5482 |
| 06/11/02 | 0 | 2.14 | 3 | 0.4980 | 0.4882 |
| 06/12/02 | 0 | 2.33 | 3 | 0.4877 | 0.4965 |
| 06/13/02 | 0 | 1.55 | 3 | 0.4664 | 0.5183 |
| 06/14/02 | 0 | 2.89 | 3 | 0.4303 | 0.5530 |
| 06/15/02 | 0.296296296 | 3.38 | 3 | 0.4404 | 0.5442 |
| 6/16/02 | 0.313017306 | 6.389423077 | 3 | 0.3618 | 0.6209 |
| 6/17/02 | 0.440366972 | 9.083333333 | 3 | 0.4739 | 0.5103 |
| 6/18/02 | 0.352941176 | 2.833333333 | 3 | 0.5078 | 0.4778 |
| 6/19/02 | 0.393442623 | 5.083333333 | 3 | 0.5191 | 0.4673 |
| 6/20/02 | 0.466666667 | 2.142857143 | 3 | 0.5226 | 0.4644 |

[^32]
# Appendix B: Corrected 2001 Annual Report 

# Release-to-McNary Survival Indices of Coho and Fall Chinook 2001 Releases (Review of Coho Brood Year 1999 and Fall Chinook Brood Year 2000) 

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Submitted April 30, 2002
Corrected April 1, 2003

## 1. Coho

## 1.a. Summary

In 2001, releases of juvenile Coho were made from two sites, Cle Elum and Easton, in the Upper Yakima and from two sites, Lost Creek and Stiles, in the Naches. There were two treatment factors assessed at each release site: 1) Release time comprised of two levels--a May 7 early release and a May 25 late release, and 2) bloodstock comprised of two levels-a Yakima-return brood and a Willard Hatchery brood.

The survival index from release to McNary Dam passage varied dramatically over subbasins and between sites within subbasins, the lowest survival indices being from the upper Yakima (as survival-index proportions, 0.014 from Cle Elum and 0.064 from Easton) and the highest being from the Naches ( 0.12 from Lost Creek and 0.30 from Stiles). Pre-release escape from the rearing ponds was high. The proportion of fish detected at McNary before the official release date was 0.21 from Cle Elum, 0.06 from Easton, 0.25 from Lost Creek, and 0.13 from Stiles. These escape estimates will be under-estimates because of the time required to migrate from the release site to McNary Dam. For the year- 2000 releases, which were made at the same sites, there was evidence that fish designated for early release and late release at Stiles intermingled prior to release date. For the 1999 releases, there was little evidence of pre-release mixing of the early and late release groups; however, there were major adjustments made in the 1999 databases that were never discussed in the 1999 Annual Report but are discussed in this report.

The 2001 early and late release-to-McNary-passage survival indices respectively were 0.13 and 0.12 and did not significantly differ $(P=0.61)$. This mirrors the lack of a significant difference from the analysis of the 2000 outmigrants, the 2000 main effect survivalindex means for early and late releases respectively being 0.24 and 0.18 and not differing significantly $(\mathrm{P}=0.53)$. Based on the analysis of the 1999 outmigrants, there was statistical evidence of a difference between that year's early- and late-releases (the main effect survival-index means of 0.53 and 0.44 respectively for early and late releases did differ significantly, $\mathrm{P}=0.05$ ); however, the adjustments to the 1999 data basis, alluded to above, would have affected the early versus late comparisons.

Yakima-brood 2001 outmigrants had greater survival indices than did Willard-brood outmigrants ( $\mathrm{P}<0.001$, survival indices being 0.051 for Yakima and 0.028 for Willard broodstock smolt released from the Upper Yakima and being 0.32 for Yakima and 0.10 for Willard broodstock released from the Naches). Only the Willard broodstock was available for 2000 outmigrants. For 1999 outmigrants, Yakima and Cascade Hatchery broodstock were available, and the Cascade brood survival index was significantly greater than that of the Yakima broodstock $(\mathrm{P}=0.02)$.

It should be noted that the method of estimating the smolt survival index from release to McNary passage in 2001 differed from the methods used in 1999 and 2000. The 1999 and 2000 data sets will be re-analyzed in the future to make the methods consistent.

## 1.b. Release to McNary Smolt-Passage Survival Indices

Tables 1.a.1, 1.a.2, and 1.a. 3 respectively give the 1999, 2000, and 2001 main effect means for time of release and for stock for each site. Tables 1.b.1, 1.b.2, and 1.b. 3 give the associated logistic analyses of variation. Only one brood source, that from Willard Hatchery, was available in 2001. In 1999, Jack Creek was one of the release sites in the Upper Yakima. In 2000 and 2001, Easton was used in place of Jack Creek. In 1999, Lost Creek was dropped from the analysis because of high disease incidence.

As can be seen in the logistic analyses of variation for the 1999 outmigrants (Table 1.b.1, means given in Table 1.a.1), the hatchery brood source (Cascade hatchery) had a significantly greater smolt survival index than did the Yakima-return brood. However, referring to Tables 1.b. 3 and 1.a.3, the 2001 smolt survival index for the Yakima brood was significantly higher than the survival index for the hatchery source (Willard hatchery was used for the broodstock for the 2001 outmigrants instead of the Cascade hatchery which was the broodstock which was used for the 1999 outmigrants).

Only for the 1999 outmigrants was there a significant difference between the early versus late ${ }^{10}$ release survival indices (analysis of variation in Table 1.b.1, means in Table 1.a.1). The method of analysis for the 1999 outmigrants was different than that used for the 2000 and 2001 outmigrants, and the analysis of variation method used for the 1999 outmigrants needs to be revised to correspond to the analysis procedure used in the subsequent outmigrant years' analyses ${ }^{11}$. There are other issues regarding the 1999 analysis procedures. The tagging files identified as early- and late-releases for sites Jack Creek

10 Early releases: May 17, May 7, and May 7 respectively in 1999, 2000, and 2001 Late releases: May 27, May 31, and May 25 respectively in 1999, 2000, and 2001
${ }^{11}$ The currently used method is that of fitting constants in which, when analyzing main effects, first a weighted regression is run which includes all main effect factors, then a weighted regression is run that drops the main effect factor of interest, after which the differences in the deviance and degrees of freedom between the two regression fits are used to evaluate the main effect of interest. Analogous procedures are utilized in evaluating two-factor interactions. In 1999, the main effect of interest was included and not adjusted for the other factors. In the presence of varying release numbers, this procedure can produce biased results. In all analyses, weighted logistic analyses of variation on the survival indices were used where the weight was the number of PIT tags released.
(one of the two Upper Yakima subbasin sites) and Stiles (one of the two Naches subbasin sites) were apparently misidentified since the "early"-release McNary-passage-time distribution was later than was the "late"-release passage. With the agreement of the Coho researcher and the field staff, the "early" releases were treated as late, and the "late" as early in the analysis. This was not pointed out in my 1999 Annual Report on Coho. If any of the reassignments were erroneous, then the early and late comparisons will be biased.

There are also problems associated with the 2000 and 2001 releases. In 2000, the proportion of PIT-tagged fish "released" from Stiles on May 31 (late-release date) that were detected at McNary on or before May 27 was 0.59 . Either there was excessive prerelease escapement from the Stiles late-release pond or there was excessive leakage between the early and late group at Stiles, the early and late groups being separated by a net in the same pond. In 2001 there was evidence of pre-release date escapement from most ponds (Table 1.c). Under these conditions, it is doubtful that accurate early and late release comparisons can be made.

It should be noted from Tables 1.a. 1 through 1.a. 3 that the overall survival indices decreased over years ( 0.485 in 1999, 0.129 in 2000, and 0.129 in 201), perhaps reflecting the reduction in flows over those years. In 2001, there were significant and substantial differences among sites. An overview over all sites suggests that survival out of the Naches may be greater than out of the Upper Yakima. The difference among sites within subbasins is not necessarily associated with distance from McNary. In the Upper Yakima, Cle Elum, with the lowest survival index, is closer to McNary than is Easton. There may be issues with rearing conditions at the different sites.

Table 1.a.1. Weighted Release-to-McNary Survival-Index (S.I.) Main Effect Means for Year 1999 Coho Releases (weight = release number).

| Subbasin | Site | Survival Index (S.I.) | Time of Release |  | Stock |  | Site Mean |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Early | Late | Cascade | Yakima |  |
| UpperYakima | Cle Elum | S.I. | 0.503 | 0.373 | 0.459 | 0.423 | 0.438 |
|  |  | Release Number | 1958 | 1995 | 1608 | 2345 | 3953 |
|  | Jack Creek | S.I. | 0.522 | 0.452 | 0.587 | 0.386 | 0.487 |
|  |  | Release Number | 2491 | 2477 | 2493 | 2475 | 4968 |
| Naches | Lost Creek | S.I. | 0.265 | 0.043 | 0.212 | 0.083 | 0.150 |
|  | (Omitted)* | Release Number | 2209 | 2364 | 2380 | 2193 | 4573 |
|  | Stiles | S.I. | 0.561 | 0.489 | 0.552 | 0.486 | 0.522 |
|  |  | Release Number | 2135 | 2493 | 2527 | 2101 | 4628 |
| Yakima | Pooled | S.I. | 0.529 | 0.443 | 0.543 | 0.429 | 0.485 |
| Basin | Mean | Release Number | 6584 | 6965 | 6628 | 6921 | 13549 |

* Lost Creek Omitted from analysis and pooled Yakima Basin means because of high disease incidence

Table 1.a.2. Weighted Release-to-McNary Survival-Index (S.I.) Main Effect
Means for Year 2000 Coho Releases (weight = release number).

| Subbasin | Site | Survival Index (S.I.) | Time of Release Early Late |  | Site <br> Mean |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{gathered} \hline \text { Upper } \\ \text { Yakima } \end{gathered}$ | Cle Elum | S.I. | 0.136 | 0.020 | 0.078 |
|  |  | Release Number | 2487 | 2462 | 4949 |
|  | Easton | S.I. | 0.278 | 0.182 | 0.230 |
|  |  | Release Number | 2476 | 2476 | 4952 |
| Naches | Lost Creek | S.I. | 0.271 | 0.148 | 0.209 |
|  |  | Release Number | 2489 | 2488 | 4977 |
|  | Stiles | S.I. | 0.259 | 0.358 | 0.31 |
|  |  | Release Number | 2488 | 2493 | 4981 |
| YakimaBasin | Pooled | S.I. | 0.236 | 0.177 | 0.207 |
|  | Mean | Release Number | 9940 | 9919 | 19859 |

* Insufficient 1998 trapped adult returns to produce Yakima broodstock, Willard hatchery only broodstock.

Table 1.a.3. Weighted Release-to-McNary Survival-Index (S.I.) Main Effect Means for Year 2001 Coho Releases (weight = release number).

| Subbasin | Site | Survival Index (S.I.) | Time of Release |  | Stock |  | $\begin{gathered} \text { Site } \\ \text { Mean } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Early | Late | Willard | Yakima |  |
| Upper <br> Yakima | Cle Elum | S.I. | 0.012 | 0.017 | 0.014 | 0.015 | 0.015 |
|  |  | Release Number | 2404 | 2459 | 2416 | 2447 | 4863 |
|  | Easton | S.I. | 0.070 | 0.061 | 0.043 | 0.088 | 0.065 |
|  |  | Release Number | 2483 | 2481 | 2468 | 2496 | 4964 |
| Naches | Lost Creek | S.I. | 0.145 | 0.109 | 0.028 | 0.226 | 0.127 |
|  |  | Release Number | 2490 | 2496 | 2485 | 2501 | 4986 |
|  | Stiles | S.I. | 0.307 | 0.308 | 0.183 | 0.430 | 0.307 |
|  |  | Release Number | 2485 | 2486 | 2473 | 2498 | 4971 |
| Yakima Basin | Pooled | S.I. | 0.135 | 0.124 | 0.067 | 0.191 | 0.129 |
|  | Mean | Release Number | 9862 | 9922 | 9842 | 9942 | 19784 |

Table 1.b.1. Logistic Analysis of Variation on Release-to-McNary Survival Indices of Year 1999 Coho Releases (bottom segment of table is a pooling of top segment, all two- and three-factor interactions being pooled into a common error).

| Source | Deviance <br> $($ Dev $)$ | Freedom <br> $(\mathrm{DF})$ | Deviance <br> $($ Dev/DF) | F-Ratio | Type 1 P |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Subbasin | 39.65 | 1 | 39.65 | 1.36 | 0.2957 |
| Subbasin $\times$ Release Time* | 6.98 | 1 | 6.98 | 0.24 | 0.6450 |
| Subbasin $\times$ Stock | 1.4 | 1 | 1.40 | 0.05 | 0.8350 |
| Site (within Subbasin) | 21.61 | 1 | 21.61 | 0.74 | 0.4282 |
| Site (within Subbasin) $\times$ Release Time* | 8.01 | 1 | 8.01 | 0.28 | 0.6222 |
| Site (within Subbasin) $\times$ Stock | 50.32 | 1 | 50.32 | 1.73 | 0.2455 |
| Release Time* | 100.71 | 1 | 100.71 | 3.46 | 0.1219 |
| Stock | $\mathbf{1 7 6 . 3 5}$ | $\mathbf{1}$ | $\mathbf{1 7 6 . 3 5}$ | 6.06 | $\mathbf{0 . 0 5 7 1}$ |
| Release Time* $\times$ Stock (adj for Site) | 8.01 | 1 | 8.01 | 0.28 | 0.6222 |
| Error (3 factor interactions) | 145.47 | 5 | 29.09 |  |  |
| Subbasin | 39.65 | 1 | 39.65 | 1.8007 | 0.2093 |
| Site (within Subbasin) | 21.61 | 1 | 21.61 | 0.9814 | 0.3452 |
| Release Time* | $\mathbf{1 0 0 . 7 1}$ | $\mathbf{1}$ | $\mathbf{1 0 0 . 7 1}$ | $\mathbf{4 . 5 7 3 8}$ | $\mathbf{0 . 0 5 8 2}$ |
| Stock | $\mathbf{1 7 6 . 3 5}$ | $\mathbf{1}$ | $\mathbf{1 7 6 . 3 5}$ | $\mathbf{6 . 8 6 7 7}$ | $\mathbf{0 . 0 2 5 6}$ |
| Error (All 3 and 2 factor interactions) | 220.19 | 10 | 22.019 |  |  |

* Release Time: Early--May 17, Late--May 27

Table 1.b.2. Logistic Analysis of Variation on Release-to-McNary Survival Indices of Year 2000 Coho Releases (bottom segment of table is a pooling of top segment, all two-factor interactions being pooled into a common error).

| Source | Deviance <br> (Dev) | Freedom <br> (DF) | Deviance <br> (Dev/DF) | F-Ratio | Type 1 P |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Subbasin | 337.57 | 1 | 337.57 | 2.53 | 0.2524 |
| Site (within Subbasin) | 584.88 | 2 | 292.44 | 2.20 | 0.3129 |
| Release Time $^{*}$ (adj for Site and Stock) | 110.31 | 1 | 110.31 | 0.83 | 0.4589 |
| Subbasin x Release Time | 112.02 | 1 | 112.02 | 0.84 | 0.4559 |
| Error (Site $\times$ Release Time) | 266.4 | 2 | 133.2 |  |  |
| Subbasin | 337.57 | 1 | 337.57 | 2.68 | 0.2004 |
| Site (within Subbasin) | 584.88 | 2 | 292.44 | 2.32 | 0.2462 |
| Release Time (adj for Site and Stock) | 110.31 | 1 | 110.31 | 0.87 | 0.4187 |
| Error (All 2 factor interctions) | 378.42 | 3 | 126.14 | 1.00 | 0.5000 |

* Release Time: Early--May 7, Late--May 31

Table 1.b.3. Logistic Analysis of Variation on Release-to-McNary Survival Indices of Year 2001 Coho Releases (bottom segment of table is a pooling of top segment, all two-factor and three-factor interactions being pooled into a common error).

| Source | Deviance (Dev) | Freedom (DF) | Deviance (Dev/DF) | F-Ratio | Type 1 P |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Subbasin | 1551.51 | 1 | 1551.51 | 37.66 | 0.0087 |
| Subbasin x Release Time* | 0.02 | 1 | 0.02 | 0.00 | 0.9838 |
| Subbasin x Stock | 41.87 | 1 | 41.87 | 1.02 | 0.3877 |
| Site (within Subbasin) | 697.54 | 2 | 348.77 | 8.47 | 0.0584 |
| Site (within Subbasin) x Release Time* | 9.68 | 2 | 4.84 | 0.12 | 0.8931 |
| Site (within Subbasin) x Stock | 76.55 | 2 | 38.28 | 0.93 | 0.4853 |
| Release Time* (adj for Site and Stock) | 5.53 | 1 | 5.53 | 0.13 | 0.7384 |
| Stock (adj for Site and Release Time*) | 784.31 | 1 | 784.31 | 19.04 | 0.0223 |
| Release Time*x Stock (adj for Site) | 5.07 | 1 | 5.07 | 0.12 | 0.7489 |
| Error (3 factor interactions) | 123.59 | 3 | 41.20 |  |  |
| Subbasin | 1551.51 | 1 | 1551.51 | 60.422 | 0.0000 |
| Site (within Subbasin) | 697.54 | 2 | 348.77 | 13.582 | 0.0014 |
| Release Time* (adj for Site and Stock) | 5.53 | 1 | 5.53 | 0.2154 | 0.6525 |
| Stock (adj for Site and Release Time*) | 784.31 | 1 | 784.31 | 30.544 | 0.0003 |
| Error (All 3 and 2 factor interactions) | 256.78 | 10 | 25.678 |  |  |

* Release Time: Early--May 7, Late--May 25

Table 1.c. Mean McNary Detection Date, Mean Release-to-McNary "Travel Time" = Mean McNary Detection Date - "Release Date, Proportion of McNary Detections Prior to "Release Date", and Related Information.

| Stock > | Willard |  | Yakima |  |
| :---: | :---: | :---: | :---: | :---: |
| Date of "Release" > | $\begin{gathered} \text { Early } \\ \text { 07-May-01 } \end{gathered}$ | Late 25-May-0 | $\begin{gathered} \text { Early } \\ \text { 07-May-0 } \end{gathered}$ | Late 25-May-01 |
| Cle Elum |  |  |  |  |
| Mean Detection Date | 06/20/01 | 06/10/01 | 05/12/01 | 06/06/01 |
| Difference: Detection Date - "Release" Date | 44 | 16 | 5 | 12 |
| Total Detected At McNary on/before "release" Date | 0 | 1 | 5 | 2 |
| Total McNary Detections | 8 | 10 | 8 | 12 |
| Proportion seen at McNary on/before "release" Date | 0.0000 | 0.1000 | 0.6250 | 0.1667 |
| Easton |  |  |  |  |
| Mean Detection Date | 06/11/01 | 06/04/01 | 05/29/01 | 06/03/01 |
| Difference: Detection Date - "Release" Date | 35 | 11 | 23 | 10 |
| Total Detected At McNary on/before "release" Date | 0 | 0 | 4 | 6 |
| Total McNary Detections | 8 | 49 | 86 | 32 |
| Proportion seen at McNary on/before "release" Date | 0.0000 | 0.0000 | 0.0465 | 0.1875 |
| Lost Creek |  |  |  |  |
| Mean Detection Date | 06/12/01 | 06/08/01 | 05/22/01 | 05/26/01 |
| Difference: Detection Date - "Release" Date | 36 | 14 | 15 | 1 |
| Total Detected At McNary on/before "release" Date | 0 | 0 | 21 | 65 |
| Total McNary Detections | 18 | 19 | 177 | 128 |
| Proportion seen at McNary on/before "release" Date | 0.0000 | 0.0000 | 0.1186 | 0.5078 |
| Stiles |  |  |  |  |
| Mean Detection Date | 06/05/01 | 05/28/01 | 05/21/01 | 05/31/01 |
| Difference: Detection Date - "Release" Date | 30 | 3 | 14 | 7 |
| Total Detected At McNary on/before "release" Date | 0 | 46 | 17 | 44 |
| Total McNary Detections | 136 | 108 | 275 | 304 |
| Proportion seen at McNary on/before "release" Date | 0.0000 | 0.4259 | 0.0618 | 0.1447 |

## 1.c. Survival-index estimators

The general form of the survival index estimator is:

$$
\text { Survial Index }=\frac{\text { Expanded PIT }- \text { Tag Detections at McNary }}{\text { Number of PIT }- \text { Tagged Fish Released }}
$$

wherein

$$
\text { Expanded PIT }- \text { Tag Detectionsat McNary }=\frac{\text { Number PIT }- \text { Tags Detected at McNary }}{\text { McNary Detection Rate }}
$$

wherein

$$
\text { McNary Detection Rate }=\frac{\text { Number of Joint Detectionsat NcNary and Down }- \text { Stream Dam }}{\text { Number of Detectionsat Down }- \text { Stream Dam }}
$$

In previous years (1999 and 2000 outmigration years), the detections were based on the down-stream dam date of detection; the downstream dams used being either John Day or Bonneville. The downstream dam used was the one that had the greatest number of detections. In 1999 this dam was John Day, but in 2001 the dam used was Bonneville. The daily McNary detection rate was first estimated for each date of detection at the downstream dam, these daily estimates were pooled over dates within strata. The beginning and ending dates of each stratum were established using detections of OCT$\mathrm{SNT}^{12}$ Spring Chinook releases. The strata beginning and ending dates were chosen in a manner such that the variation among Spring Chinook daily detection rates within strata was minimized and the detection-rate variation among strata was maximized. The OCTSNT detections were used to establish the strata instead of Coho because of the large number of daily detections at the Columbia River Dams of the approximately 40,000 OCT-SNT PIT-tagged fish released. The number of Coho detections was insufficient to establish the strata. However, even though the strata were established using OCT-SNT Spring Chinook, the detection rate estimates for Coho were based on the lower dam detections of Coho, not OCT-SNT Spring Chinook.

The lower dam stratum dates were then offset by the mean travel time from McNary to the downstream dam to establish a McNary Day of passage. In 1999, John Day served as the downstream dam because there were more Coho (and Spring Chinook) detections at John Day than at Bonneville; whereas, in 2000, Bonneville served as the downstream dam because there were more Coho detections there than at John Day.

McNary Detections on a given date were allocated to the downstream-dam detection stratum by applying the McNary-to-downstream-dam migration-time relative distribution based on joint McNary and lower dam detections to the total detections at McNary. The allocated counts were then divided by the respective stratum's McNary detection rate

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estimate to get the expanded count; these expanded values were then totaled over strata and McNary detection dates to get the expanded counts for the release; and these totals were then divided by the release numbers to obtain the estimated survival indices.

The stratified estimates for 1999 are given in Appendix Table 1.a. The 2000 downstream dam detections were so small so as to render stratification meaningless, so the total joint downstream and McNary Dam count divided by the total downstream dam count was used as a single, non-stratified estimate of McNary detection rate. The non-stratified estimates for 2000 are given in Appendix Table 1.b.

A different method of detection rate estimation was deve loped in 2001 and was applied to 1999, 2000, and 2001 OCT-SNT Spring Chinook outmigrants. This method was applied to the 2001 Coho releases, and among other things, the method permitted the pooling of the John-Day-based and the Bonneville-based McNary detection rates. The reason for the pooling is that in recent years, notably 2000, experiments were conducted at John Day which varied the proportion of flow spilled in the daytime relative to the proportion spilled at night. To stabilize the electric power available, contravening action was taken at Bonneville; when the relative daytime spill proportion was increased at John Day, the relative daytime spill proportion was decreased at Bonneville ${ }^{13}$. Given this situation, it was deemed more appropriate to pool John Day and Bonneville Dam-based estimates of the McNary detection rate. This was done for the 2001 outmigrants even though there was minimal spill during this outmigration year. Time constraints precluded the new method's application to the 1999 and 2000 Coho outmigration data.

The 2001 downstream dam detections of Coho were also too small for stratification purposes, so the total joint downstream and McNary Dam count, pooled over Bonneville ${ }^{14}$ and John Day dams, divided by the total downstream dam count, pooled over Bonneville and John Day dams, served as a single, non-stratified estimate of the detection rate for the 2001 outmigrants. The non-stratified estimates for 2001 are given in Appendix Table 1.c.

[^33]Next year, the methodology developed for the OCT-SNT Spring Chinook will be applied to all years of Coho outmigration as well, which will require an updating of the 1999 and 2000 estimates. For the purpose of estimating the detection rates in the future, joint McNary and downstream dam detections and total downstream detections will be based all Yakima PIT-tag releases of Coho, not only those associated with this study. The resulting increased number of detections may facilitate the stratification process. If this is done in the future, the 2001 estimates will also be updated. As was the case for the 2001 estimates presented here, future downstream dam estimates of the McNary detection rates will be pooled over the two downstream dams. See the report Release-to-McNary Survival Index of 1999-2001 Spring Chinook Releases for the methodology to be used.

## 2. Fall Chinook

## 2.a. Summary

Three PIT-tagged sets of replicated releases were made of Fall Chinook in 2001, a Marion Drain release set, and two sets in the Yakima River below Prosser Dam-a replicated accelerated rearing treatment set which brought the fish up to release size more rapidly than the other set, a replicated conventional rearing treatment set. The replicates consisted of releases made 1 day apart (the initial release dates for the three sets differed among the release sets). The analyses were combined over all three sets to increase the degrees of freedom associated with the statistical tests. The tests were not powerful because of the low degrees of freedom ( 3 degrees of freedom) and the large variation in the survival-index estimates ${ }^{15}$ between the replicates within the accelerated and conventional rearing treatment sets; therefore there were no significant differences ( $\mathrm{P}=$ 0.48 ) among the release-to-McNary survival indices of the three sets (proportional survival-index estimates being $0.39,0.27$, and 0.30 respectively for below-Prosser accelerated, below-Prosser conventional, and Marion Drain releases).

For the 2000 releases, the accelerated treatment had a significantly lower $(\mathrm{P}=0.03)$ survival index than did the conventional treatment ( 0.43 for the accelerated and 0.82 for the conventional rearing releases). The 0.27 survival index for the Marion Drain release differed significantly from the conventionally reared below-Prosser release $(\mathrm{P}=0.02)$ but not from the accelerated release $(\mathrm{P}=0.29)$.

For the 1999 releases, there was no replication and, therefore, no basis for statistical testing. The survival-index estimate in 1999 was 0.49 for the accelerated, 0.26 for the control, and 0.44 for the Marion Drain releases.

## 2.b. Release to McNary Smolt-Passage Survival Indices

Tables 2.a. 1 and 2.a. 2 respectively give the 2001 and 2000 survival-index means for the three release sets. Tables 2.b.1 and 2.b. 2 give the logistic analysis of variation associated with the means. As stated before, the 1999 data set involved no replication. The

[^34]databases for the 1999, 2000, and 2001 releases are given respectively in Appendix Tables 2.a, 2.b, and 2.3. For the 1999 and 2001 releases, the accelerated treatment survival-index mean was greater than the conventional mean; however, in neither year was the difference significant. For the 1999 releases an assessment of significance was not possible because of the lack of replication. For the 2001 releases, the individual two replicate estimates for each of the two treatments actually overlapped-- 0.33 and 0.45 for the accelerated and 0.18 and 0.35 for the conventional (Appendix Table 2.c). The result is that the 1999 accelerated survival-index estimate ( 0.49 ) and the conventional estimate (0.26) are not judged to differ significantly ( $\mathrm{P}=0.28$ ), nor is the 2001 accelerated estimate (0.39) and 2001 conventional estimate (0.30) (P not estimable for 1999).

The differences between the 2000 accelerated and conventional release estimates is the opposite of that in 1999 and 2001, the convention survival-index mean of 0.82 exceeding accelerated estimate of 0.43 ; and, in this case, the difference was significant $(\mathrm{P}=0.03)$.

Normally, there would be no interest in comparing the Marion Drain releases to the other two treatment sets, the Marion Drain releases being made in a different location upstream of the below-Prosser Dam releases of the accelerated and conventional treatment sets. However, there is one point of concern. In 1999, the Marion Drain release date (May 27) was much closer to that of the below-Prosser conventional release (May 25) than that of the below-Prosser accelerated release (April 25). However, in subsequent years, the Marion Drain release dates were more comparable to those of the below-Prosser accelerate treatment: 2000 Marion Drain--April 10-11, Accelerated -- April 20-21, and Conventional-May 25-26; 2001 Marion Drain--April 12-13, Accelerated--April 19-20, and Conventional-May 16-17. The concern is whether Marion Drain fish were being reared in 2000 and 2001 using the accelerated method because the 1999 accelerated survival index was greater than that of the conventional. If this is the case, then the decision was premature. The only year in which a significant difference could be detected between the Marion Drain release and the other releases is 2000; and in that year, the Marion Drain release mean survival index of 0.27 was significantly less than the 0.82 of the conventional ( $\mathrm{P}=0.02$ ) and did not significantly differ from the 0.43 of the accelerated $(\mathrm{P}=0.29)$. At this time, there is no statistically significant evidence that accelerated rearing improves survival of Fall Chinook; and there is statistically significant evidence that accelerated rearing can lead to a reduction in survival in some years.

## 2.c. Survival-index estimators

The general form of the survival index estimator is the same as presented for Fall Chinook. In previous years (1999 and 2000) outmigration years, the McNary detection rates were based on the John Day Dam detections using the John Day date of detection. The reason for using John Day Dam instead of Bonneville in those two years is that the number of Fall Chinook detections was greater than at Bonneville.

As was the case for Coho, the John Day and Bonneville detections were pooled to obtain the 2001 McNary detection rate estimate. And, again, as was the case for Coho, there
were not enough downstream detections of Fall Chinook to permit stratification. This was true for 2000 outmigrants as well. Strata in 1999 were established separately from those established for 1999 Spring Chinook outmigrants (and used for Coho).

The same future plans discussed for Coho will be applied to Fall Chinook.
Table 2.a.1. Weighted Release-to-McNary Survival-Index (S.I.) Means for Year-2000 Fall Chinook Releases. (weight = release number).

|  | Logistic Estimates |  | Survival Index |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Logistic Coefficient | Standard <br> Error (SE) | Estimate ${ }^{2}$ | $\begin{gathered} \text { Standard } \\ \text { Error }^{3} \text { (SE) } \end{gathered}$ |
| Accelerated Conventional Marion Drain | $\begin{array}{r} -0.2896 \\ 1.4930 \\ -0.9898 \\ \hline \end{array}$ | $\begin{aligned} & 0.29229 \\ & 0.37446 \\ & 0.46329 \\ & \hline \end{aligned}$ | $\begin{aligned} & 0.4281 \\ & 0.8165 \\ & 0.2709 \end{aligned}$ | 0.07156 0.05610 0.09152 |
| Accelerated - Conventional t-test Type 1 P | $\begin{aligned} & -1.7826 \\ & -3.7526 \\ & 0.0331 \end{aligned}$ | 0.47504 |  |  |
| Accelerated - Marion Drain <br> t-test <br> Type 1P | $\begin{aligned} & 0.7002 \\ & 1.2782 \\ & 0.2911 \end{aligned}$ | 0.54779 |  |  |
| Conventional - Marion Drain t-test <br> Type 1 P | $\begin{aligned} & 2.4828 \\ & 4.1679 \\ & 0.0251 \\ & \hline \end{aligned}$ | $0.59570$ |  |  |
| Standard Error from Logistic Output * Square Root (Mean Deviance) <br> ${ }^{2}$ Estimate $=1 /\{1+\exp [-($ Logistic Coefficient) $]\}$ <br> ${ }^{3} \mathrm{SE}($ Estimate $)=$ Estimate $^{2 *} \exp \left(-\right.$ Logistic Coefficient)${ }^{*} \mathrm{SE}($ Logisitic Coefficient) exp in above is exponential constant |  |  |  |  |

Table 2.a.2. Weighted Release-to-McNary Survival-Index (S.I.) Means for Year-2001 Fall Chinook Releases. (weight = release number).

|  | Logistic Estimates |  | Survival Index |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Logistic Coefficient | Standard Error ${ }^{1}$ (SE) | Estimate ${ }^{2}$ | $\begin{gathered} \text { Standard } \\ \text { Error }^{3} \text { (SE) } \end{gathered}$ |
| Accelerated | -0.4659 | 0.26604 | 0.3856 | 0.06303 |
| Conventional | -0.9954 | 0.29532 | 0.2699 | 0.05819 |
| Marion Drain | -0.8550 | 0.39766 | 0.2984 | 0.08325 |
| Accelerated - Conventional t-test Type 1 P | $\begin{aligned} & 0.5295 \\ & 1.3320 \\ & 0.2750 \end{aligned}$ | 0.39748 |  |  |
| Accelerated - Marion Drain t-test <br> Type 1 P | $\begin{aligned} & 0.3891 \\ & 0.8133 \\ & 0.4756 \\ & \hline \end{aligned}$ | 0.47844 |  |  |
| Conventional - Marion Drain t-test Type 1 P | $\begin{aligned} & -0.1404 \\ & -0.2834 \\ & 0.7953 \\ & \hline \end{aligned}$ |  |  |  |
| Standard Error from Logistic Output * Square Root (Mean Deviance) <br> ${ }^{2}$ Estimate $=1 /\{1+\exp [-($ Logistic Coefficient) $]\}$ <br> ${ }^{3}$ SE(Estimate) $=$ Estimate $^{2 *} \exp \left(-\right.$ Logistic Coefficient)${ }^{*}$ SE(Logisitic Coefficient) <br> exp in above is exponential constant |  |  |  |  |
|  |  |  |  |  |

Table 2.b.1. Logistic Analysis of Variation on Release-to-McNary Survival Indices of Year 2000 Fall Chinook Releases.

|  | Meviation |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Source | Degrees of <br> (Dev) | Mean <br> Deviance <br> (Dev/DF) | F- <br> Ratio | Type 1 <br> Error (P) |  |
| Among Release Sets | 1079.89 | 2 | 539.95 | 12.70 | 0.0328 |
| Within Release Sets | 127.59 | 3 | 42.53 |  |  |

Table 2.b.2. Logistic Analysis of Variation on Release-to-McNary Survival Indices of Year 2001 Fall Chinook Releases.

| Source | Deviation <br> (Dev) | Degrees of <br> Freedom (DF) | Mean <br> Deviance <br> (Dev/DF) | F- <br> Ratio | Type 1 <br> Error (P) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Among Release Sets | 64.07 | 2 | 32.04 | 0.95 | 0.4795 |
| Within Release Sets | 101.31 | 3 | 33.77 |  |  |

## Appendix (2001 Annual Report)

Table 1.a. 1999 Coho Release-to-McNary-Passage Smolt-Survival-Index Database

| Cle Elum (UPPER YAKIMA) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | McNary Detections |  |  |  | John Day Detections |  |  | Passage (Expanded Values) |  |  |  |
| SRATUM | Cascade Early | $\begin{gathered} \hline \text { Cascade } \\ \text { Late } \end{gathered}$ | Yakima Early | Yakima Late | J.D. | $\begin{aligned} & \hline \text { J.D. } \\ & \mathrm{McN} \end{aligned}$ | $\begin{gathered} \hline \text { Detection } \\ \text { Rates } \\ \hline \end{gathered}$ | Cascade Early | Cascade Late | Yakima Early | Yakima Late |
| 3 | 8.0 | 0.0 | 13.8 | 0.0 | 55 | 15 | 0.2727 | 29.2 | 0.0 | 50.6 | 0.0 |
| 4 | 55.4 | 24.8 | 74.3 | 33.8 | 1542 | 284 | 0.1842 | 300.8 | 134.9 | 403.3 | 183.5 |
| 5 | 7.5 | 16.4 | 7.4 | 18.4 | 306 | 62 | 0.2026 | 36.9 | 80.8 | 36.3 | 91.0 |
| 6 | 4.4 | 11.0 | 10.5 | 13.9 | 406 | 72 | 0.1773 | 25.0 | 62.3 | 59.0 | 78.4 |
| 7 | 0.7 | 6.7 | 4.1 | 5.9 | 471 | 52 | 0.1104 | 6.6 | 61.0 | 37.1 | 53.2 |
| Total ${ }^{2}$ | 76.0 | 59.0 | 110.0 | 72.0 | 2780 | 485 |  | 398.5 | 339.0 | 586.3 | 406.0 |
| Number Released |  |  |  |  |  |  |  | 799.0 | 809.0 | 1159.0 | 1186.0 |
| Survival Ind |  |  |  |  |  |  |  | 0.50 | 0.42 | 0.51 | 0.34 |
| JACK CREEK (UPPER YAKIMA) |  |  |  |  |  |  |  |  |  |  |  |
|  |  | McNary D | ections ${ }^{\prime}$ |  |  | Day De | tions |  | ssage (Exp | ded Valu |  |
|  | Cascade | Cascade | Yakima | Yakima |  | J.D. | Detection | Cascade | Cascade | Yakima | Yakima |
| SRATUM | Early ${ }^{3}$ | Late ${ }^{4}$ | Early ${ }^{3}$ | Late ${ }^{4}$ | J.D. | McN | Rates | Early | Late | Early | Late |
| 3 | 2.9 | 0.0 | 0.2 | 0.0 | 55 | 15 | 0.2727 | 10.8 | 0.0 | 0.6 | 0.0 |
| 4 | 103.4 | 25.4 | 38.4 | 6.1 | 1542 | 284 | 0.1842 | 561.4 | 137.8 | 208.7 | 33.3 |
| 5 | 19.6 | 31.1 | 14.3 | 6.6 | 306 | 62 | 0.2026 | 97.0 | 153.7 | 70.5 | 32.8 |
| 6 | 16.3 | 20.6 | 11.1 | 18.1 | 406 | 72 | 0.1773 | 91.8 | 116.3 | 62.8 | 102.1 |
| 7 | 5.7 | 26.9 | 16.0 | 33.1 | 471 | 52 | 0.1104 | 52.0 | 243.3 | 144.7 | 300.1 |
| Total ${ }^{2}$ | 151.0 | 104.0 | 80.0 | 64.0 | 2780 | 485 |  | 812.9 | 651.1 | 487.3 | 468.2 |
| Number Released |  |  |  |  |  |  |  | 1246 | 1247 | 1245 | 1230 |
| Survival Index |  |  |  |  |  |  |  | 0.6524 | 0.5221 | 0.3914 | 0.3807 |

LOST CREEK (NACHES) [not used in analyses because of disease]

|  |  | McNary D | tections ${ }^{\text {a }}$ |  |  | Day De | tiions |  | ssage (Exp | ded Valu |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SRATUM | Cascade Early | Cascade <br> Late | Yakima Early | Yakima Late | J.D. | $\begin{aligned} & \mathrm{J} . \mathrm{D} . \\ & \mathrm{McN} \\ & \hline \end{aligned}$ | $\begin{gathered} \hline \text { Detection } \\ \text { Rates } \\ \hline \end{gathered}$ | Cascade Early | $\begin{gathered} \hline \text { Cascade } \\ \text { Late } \\ \hline \end{gathered}$ | Yakima Early | Yakima Late |
| 3 | 3.6 | 0.9 | 0.2 | 0.0 | 55 | 15 | 0.2727 | 13.3 | 3.5 | 0.6 | 0.0 |
| 4 | 37.3 | 1.4 | 7.4 | 0.0 | 1542 | 284 | 0.1842 | 202.5 | 7.4 | 39.9 | 0.0 |
| 5 | 6.3 | 2.4 | 2.9 | 0.0 | 306 | 62 | 0.2026 | 31.0 | 12.0 | 14.2 | 0.0 |
| 6 | 7.1 | 1.3 | 3.8 | 0.0 | 406 | 72 | 0.1773 | 40.2 | 7.1 | 21.6 | 0.0 |
| 7 | 14.7 | 6.0 | 9.8 | 2.0 | 471 | 52 | 0.1104 | 132.8 | 54.3 | 88.4 | 18.1 |
| Total ${ }^{2}$ | 69.0 | 15.0 | 24.0 | 2.0 | 2780 | 485 |  | 419.8 | 84.3 | 164.8 | 18.1 |
| Number Released |  |  |  |  |  |  |  | 1160.0 | 1220.0 | 1049.0 | 1144.0 |
| Survival Index |  |  |  |  |  |  |  | 0.36 | 0.07 | 0.16 | 0.02 |


| STILES (NACHES) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | McNary D | tections |  |  | Day De | ctions |  | ssage (Ex | nded Val |  |
| SRATUM | Cascade Early ${ }^{3}$ | $\begin{gathered} \text { Cascade } \\ \text { Late }^{4} \end{gathered}$ | Yakima Early ${ }^{3}$ | Yakima Late ${ }^{4}$ | J.D. | $\begin{aligned} & \text { J.D. } \\ & \mathrm{McN} \\ & \hline \end{aligned}$ | $\begin{gathered} \text { Detection } \\ \text { Rates } \\ \hline \end{gathered}$ | Cascade Early | $\begin{gathered} \hline \text { Cascade } \\ \text { Late } \end{gathered}$ | Yakima Early | Yakima <br> Late |
| 3 | 23.7 | 0.0 | 7.2 | 0.0 | 55 | 15 | 0.2727 | 87.0 | 0.0 | 26.5 | 0.0 |
| 4 | 95.8 | 63.8 | 64.5 | 27.9 | 1542 | 284 | 0.1842 | 520.2 | 346.6 | 350.4 | 151.2 |
| 5 | 10.3 | 38.5 | 7.8 | 21.4 | 306 | 62 | 0.2026 | 50.6 | 190.1 | 38.6 | 105.5 |
| 6 | 3.7 | 21.7 | 9.5 | 21.7 | 406 | 72 | 0.1773 | 20.8 | 122.6 | 53.3 | 122.3 |
| 7 | 1.5 | 4.9 | 4.0 | 15.1 | 471 | 52 | 0.1104 | 13.9 | 44.4 | 35.9 | 136.5 |
| Total ${ }^{2}$ | 136.0 | 132.0 | 93.0 | 86.0 | 2780 | 485 |  | 692.4 | 703.7 | 504.7 | 515.6 |
| Number Released |  |  |  |  |  |  |  | 1277.0 | 1250.0 | 858.0 | 1243.0 |
| Survival Index |  |  |  |  |  |  |  | 0.54 | 0.56 | 0.59 | 0.41 |

1 McNary (McN) to John Day (J.D.) Travel-Time Distribution Adjusted
2 Total includes Stratum 1 and 2, the detection numbers for which were so small that they were excluded from the survival index estimattion
3 Presumed early that is identified as late in tagging file
4 Presumed late that is identified as early in tagging file

| SRATUM: | John Day Stratum Dates |  | McNary Stratum Dates |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | $4 / 1 / 99$ | to | $5 / 9 / 99$ | $4 / 1 / 99$ | to | $5 / 3 / 99$ |
| 2 | $5 / 10 / 99$ | to | $5 / 19 / 99$ | $5 / 4 / 99$ | to | $5 / 14 / 99$ |
| 3 | $5 / 20 / 99$ | to | $5 / 28 / 99$ | $5 / 15 / 99$ | to | $5 / 24 / 99$ |
| 4 | $5 / 29 / 99$ | to | $6 / 7 / 99$ | $5 / 25 / 99$ | to | $6 / 4 / 99$ |
| 5 | $6 / 8 / 99$ | to | $6 / 10 / 99$ | $6 / 5 / 99$ | to | $6 / 7 / 99$ |
| 6 | $6 / 11 / 99$ | to | $6 / 17 / 99$ | $6 / 8 / 99$ | to | $6 / 14 / 99$ |
| 7 | $6 / 18 / 99$ | to | $7 / 31 / 99$ | $6 / 15 / 99$ | to | $7 / 31 / 99$ |

Table 1.b. 2000 Coho Release-to-McNary-Passage Smolt-Survival-Index Database

| Site | Stock | Release <br> Time | McNary Detections | $\begin{gathered} \text { Detection } \\ \text { Rate }^{1} \\ \hline \end{gathered}$ | Passage (Expanded Values) | $\begin{gathered} \text { Release } \\ \text { Size } \\ \hline \end{gathered}$ | Survival Index |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Cle Elum (Upper Yakima) | Willard | Early | 70 | 0.2063 | 339.3 | 2487 | 0.1364 |
| Cle_Elum (Upper Yakima) | Willard | Late | 10 | 0.2063 | 48.5 | 2462 | 0.0197 |
| Easton (Upper Yakima) | Willard | Early | 142 | 0.2063 | 688.3 | 2476 | 0.2780 |
| Easton (Upper Yakima) | Willard | Late | 93 | 0.2063 | 450.8 | 2476 | 0.1821 |
| Lost Creek (Naches) | Willard | Early | 139 | 0.2063 | 673.8 | 2489 | 0.2707 |
| Lost Creek (Naches) | Willard | Late | 76 | 0.2063 | 368.4 | 2488 | 0.1481 |
| Stiles (Naches) | Willard | Early | 133 | 0.2063 | 644.7 | 2488 | 0.2591 |
| Stiles (Naches) | Willard | Late | 184 | 0.2063 | 891.9 | 2493 | 0.3578 |
| Detection Rate $=$ (Number Jointly Detected at McNary and Bonneville)/(Number Detected at Bonneville) Joint at McN and Bonn $=59$ <br> Total at Bonn $=286$ <br> Detection Rate $=0.2063$ |  |  |  |  |  |  |  |

Table 1.c. 2000 Coho Release-to-McNary-Passage Smolt-Survival-Index Database

| Site | Stock | Release Time | McNary <br> Detections | $\begin{gathered} \text { Detection } \\ \text { Rate }^{1} \\ \hline \end{gathered}$ | Passage (Expanded Values) | Release Number | Survival Index |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Cle Elum (Upper Yakima) | Willard | Early | 8 | 0.5388 | 14.8 | 1197 | 0.0124 |
| Cle Elum (Upper Yakima) | Willard | Late | 10 | 0.5388 | 18.6 | 1219 | 0.0152 |
| Cle Elum (Upper Yakima) | Yakima | Early | 8 | 0.5388 | 14.8 | 1207 | 0.0123 |
| Cle Elum (Upper Yakima) | Yakima | Late | 12 | 0.5388 | 22.3 | 1240 | 0.0180 |
| Easton (Upper Yakima) | Willard | Early | 8 | 0.5388 | 14.8 | 1234 | 0.0120 |
| Easton (Upper Yakima) | Willard | Late | 49 | 0.5388 | 90.9 | 1234 | 0.0737 |
| Easton (Upper Yakima) | Yakima | Early | 86 | 0.5388 | 159.6 | 1249 | 0.1278 |
| Easton (Upper Yakima) | Yakima | Late | 32 | 0.5388 | 59.4 | 1247 | 0.0476 |
| Lost Creek (Naches) | Willard | Early | 18 | 0.5388 | 33.4 | 1240 | 0.0269 |
| Lost Creek (Naches) | Willard | Late | 19 | 0.5388 | 35.3 | 1245 | 0.0283 |
| Lost Creek (Naches) | Yakima | Early | 177 | 0.5388 | 328.5 | 1250 | 0.2628 |
| Lost Creek (Naches) | Yakima | Late | 128 | 0.5388 | 237.6 | 1251 | 0.1899 |
| Stiles (Naches) | Willard | Early | 136 | 0.5388 | 252.4 | 1236 | 0.2042 |
| Stiles (Naches) | Willard | Late | 108 | 0.5388 | 200.4 | 1237 | 0.1620 |
| Stiles (Naches) | Yakima | Early | 275 | 0.5388 | 510.4 | 1249 | 0.4086 |
| Stiles (Naches) | Yakima | Late | 304 | 0.5388 | 564.2 | 1249 | 0.4517 |
| Detection Rate = (Number Jointly Detected at McNary and Bonneville)/(Number Detected at Bonneville) |  |  |  |  |  |  |  |
| Joint at McN | and JD $=131$ |  | and JD = | 119 |  | Pooled $=250.0000$ |  |
| Total | at JD |  | at JD $=$ | 228 |  | Pooled $=$ | 464.0000 |
| Detection Rate $=$ | 0.5551 |  |  | 0.5219 |  |  | 0.5388 |

Table 2.a. 1999 Fall Chinook Release-to-McNary-Passage Smolt-Survival-Index Database

| Stratum | McNary Detections ${ }^{1}$ |  |  | John Dav Detections |  |  | Passage (Expanded Values) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Accelerated | Conventional | Marion Drain | J.D. | $\begin{aligned} & \text { J.D. } \\ & \mathrm{McN} \end{aligned}$ | Detection Rates | Accelerated | Conventional | Marion Drain |
| 1 | 24.5 | 0.0 | 0.0 | 30 | 2 | 0.0667 | 367.8 | 0.0 | 0.0 |
| 2 | 160.5 | 4.2 | 0.1 | 196 | 62 | 0.3163 | 507.4 | 13.4 | 0.4 |
| 3 | 12.9 | 38.6 | 17.3 | 162 | 22 | 0.1358 | 95.0 | 284.5 | 127.4 |
| 4 | 0.0 | 40.6 | 66.0 | 150 | 32 | 0.2133 | 0.0 | 190.1 | 309.2 |
| 5 | 0.0 | 15.2 | 7.5 | 21 | 10 | 0.4762 | 0.0 | 31.9 | 15.7 |
| Total | 197.9 | 98.6 | 90.9 | 559 | 128 | 0.2290 | 970.3 | 519.9 | 452.7 |
| Number Released |  |  |  |  |  |  | 2000 | 1973 | 1032 |
| Survival Index |  |  |  |  |  |  | 0.4851 | 0.2635 | 0.4386 |
| Rlease Date |  |  |  |  |  |  | 4/25/99 | 5/25/99 | 5/26/99 |

McNary (McN) to John Day (J.D.) Travel-Time Distribution Adjusted

| Stratum | John Day |  | Stratum Dates | McNary Stratum Dates |  |  |
| :---: | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | $05 / 06 / 99$ | to | $05 / 27 / 99$ |  | to | $05 / 14 / 99$ |
| 2 | $05 / 28 / 99$ | to | $06 / 08 / 99$ | $05 / 15 / 99$ | to | $05 / 29 / 99$ |
| 3 | $06 / 09 / 99$ | to | $06 / 22 / 99$ | $05 / 30 / 99$ | to | $06 / 14 / 99$ |
| 4 | $06 / 23 / 99$ | to | $07 / 06 / 99$ | $06 / 15 / 99$ | to | $07 / 01 / 99$ |
| 5 | $07 / 07 / 99$ | to | $07 / 27 / 99$ | $07 / 02 / 99$ | to |  |

Table 2.b. 2000 Fall Chinook Release-to-McNary-Passage Smolt-Survival-Index Database

| Treatment/ Release Site | Replicate | Release Date | McNary <br> Detections | Detection Rate1 | Passage (Expanded Value) | Release <br> Number | Survival Index |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Accelerated Accelerated | Replicate 1 Replicate 2 | $\begin{array}{r} 4 / 20 / 00 \\ 4 / 21 / 00 \\ \hline \end{array}$ | $\begin{array}{r} 126 \\ 127 \end{array}$ | $\begin{aligned} & 0.2907 \\ & 0.2907 \end{aligned}$ | $\begin{array}{r} 433.44 \\ 436.88 \end{array}$ | $\begin{aligned} & 1000 \\ & 1033 \end{aligned}$ | $\begin{array}{r} 0.4334 \\ 0.4229 \\ \hline \end{array}$ |
| Conventional Conventional | Replicate 1 Replicate 2 | $\begin{aligned} & 5 / 25 / 00 \\ & 5 / 26 / 00 \\ & \hline \end{aligned}$ | $\begin{aligned} & 233 \\ & 246 \\ & \hline \end{aligned}$ | $\begin{aligned} & 0.2907 \\ & 0.2907 \end{aligned}$ | $\begin{aligned} & 801.52 \\ & 846.24 \end{aligned}$ | $\begin{aligned} & 1008 \\ & 1010 \end{aligned}$ | $\begin{array}{r} 0.7952 \\ 0.8379 \\ \hline \end{array}$ |
| Marion Drain Marion Drain | Replicate 1 Replicate 2 | $\begin{aligned} & 4 / 11 / 00 \\ & 4 / 10 / 00 \\ & \hline \end{aligned}$ | 17 | $\begin{aligned} & 0.2907 \\ & 0.2907 \end{aligned}$ | $\begin{array}{r} 58.48 \\ 213.28 \end{array}$ | $\begin{aligned} & 495 \\ & 508 \\ & \hline \end{aligned}$ | $\begin{aligned} & 0.1181 \\ & 0.4198 \end{aligned}$ |
| ```1 Detection Number = (Number Jointly Detected at McN and J.D.)/(Number Detected at J.D.) Joint at McN and J.D. = 75 Total at J.D. = 258 Detection Rate = 0.2907``` |  |  |  |  |  |  |  |

Table 2.c. 2001 Fall Chinook Release-to-McNary-Passage Smolt-Survival-Index Database
$\left.\begin{array}{|c|c|c|c|c|c|c|c|}\hline \begin{array}{c}\text { Treatment/ } \\ \text { Release } \\ \text { Site }\end{array} & & \text { Replicate } & \begin{array}{c}\text { Release } \\ \text { Date }\end{array} & \begin{array}{c}\text { McNary } \\ \text { Detections }\end{array} & \begin{array}{c}\text { Detection } \\ \text { Rate }\end{array} & \begin{array}{c}\text { Passage } \\ \text { (Expanded } \\ \text { Value) }\end{array} & \begin{array}{c}\text { Release } \\ \text { Number }\end{array} \\ \hline \text { Accelerated } & \text { Replicate 1 } & 4 / 19 / 01 & 285 & 0.6374 & 447.1 & 1002 & 0.4462 \\ \text { Index }\end{array}\right]$

## APPENDIX E

## Yakima Spring Chinook J uvenile Behavior:

Comparisons of Hatchery and Wild Spring Chinook Oncorhynchus tshawytsha Smolts in Cover Utilization and Avoidance of Predation by Northern Pikeminnows, Ptychocheilus oregonensis

## Yakima Spring Chinook Juvenile Behavior:

# Comparisons of Hatchery and Wild Spring Chinook Oncorhynchus tshawytsha Smolts in Cover Utilization and Avoidance of Predation by Northern Pikeminnows, Ptychocheilus oregonensis 

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#### Abstract

Yearling Spring Chinook smolts, Onchorhynchus tshawytsha, from two hatchery treatment groups, conventional, (OCT), and semi natural (SNT), rearing treatments were compared to wild smolts in an experiment designed to assess differences in cover utilization, and survival to a predation threat. Survival to Northern Pikeminnows was seen to be size dependant for hatchery fish, $(\mathrm{p}=0.001)$, with the largest fish $(141-158 \mathrm{~mm})$ surviving at over twice the rate as the smallest fish $(90-120 \mathrm{~mm})$. Survival was not size dependant for wild fish however, $(\mathrm{p}=0.713)$. Overall survival rates were similar between the three groups, although wild smolts tended to be smaller. Among the smaller smolts ( $<=130 \mathrm{~mm}$ ), wild smolts survived at higher rates, and rate was significantly different from the OCT group, ( $\mathrm{p}=0.033$ ).

The order of introduction did not significantly affect the time any of the three groups of smolts remained in cover, indicating that the presents or absence of other smolts did not influence a newly introduced smolts decision on how long to remain in cover. No signific ant difference was found between the two hatchery treatments in time spent in cover. The semi-naturally reared smolts spent the least time in cover, and the difference from the wild was significant, $(p=0.023)$. Qualitative observations also revealed little difference between the conventional and semi naturally reared smolts. In comparison to wild smolts, hatchery smolts appeared less adept at finding and concealing themselves in cover. Wild smolts also tended to swim less, i.e. in cover they appeared nearly motionless, whereas hatchery fish were almost always swimming.


## Introduction

Chinook salmon culture efforts in the Northwest have produced only very modest returns. The high cost of operating hatchery facilities, combined with their low returns and uncertainties concerning their impacts on wild stocks have led some to question their effectiveness as a tool in salmon restoration efforts, (Winton and Hilborn, 1994).

The low levels of returning fish of hatchery origin is due largely to the much lower survival of hatchery smolts compared to their wild counterparts. For example, in sub basins of the Columbia River, depending upon the race of the species and the particular sub basin, between 8 and 100 times the numbers of wild chinook salmon return to spawn compared to hatchery-reared fish (Fast, et. all., 1991; Mullan et al., 1992).

The phenotypic expression of wild fish differs from that of their hatchery reared counterparts. In salmonids, a host of investigations have reported differences in behavior, (e.g., Bachman, 1984; DeVietti, 1992; Dickson \& MacCrimmon, 1982; Miller, 1958; Vincent, 1960), physiology (e.g., Miller, Sinclair, \& Hochachka, 1959; Vincent, 1960), and morphology (e.g., Sosiak, 1982; Swain, Riddell, and Murray, 1991) between wild and hatchery-reared fish (See White, Karr, \& Nohlson, 1995 for a review). These differences in traits correlate with a large survival advantage favoring the wild fish. Traits of the wild fish have evolved through natural selection over generations from the genesis of anadromy to the present time. It follows that traits in hatchery-reared fish that differ from the wild fish offer little or no survival advantage. Moreover, some have argued that these different traits established through hatchery rearing are actually counterproductive to survival in the wild (DeVietti, 1992; Dickson \& MacCrimmon, 1982; Fast, et al. 1991; Hilborn, 1992; Mullan, at al. 1992, While et al., 1995).

Variances in behaviors between species and systems are large, however. The recommendation by two reviewers is that the lessons learned in one system are not necessarily transferable to another, thus culture strategies may have to be investigated and tailored to the system they are applied, (Winton and Hilborn, 1994).

There is growing interest in the innovative rearing of hatchery fish in which the specific aim is to increase the return to spawn numbers and thus, explicitly or implicitly, to alter the traits of the cultured fish toward those shown by the wild fish. One such effort is currently underway at the Cle Elum Supplementation and Research Facility, (CESRF). CESRF is a Spring Chinook hatchery located on the Yakima River near the headwaters on the eastern slopes of the Cascade Mountains in South Central Washington, 832 river kilometers from the Pacific Ocean. CESRF is operated by the Yakima Nation, with funding provided by the Bonneville Power Administration. It's mission is help restore runs of Spring Chinook in the Yakima Basin by raising and releasing the progeny of wild origin fish into the Yakima River.

Since the spring of 1999 , this facility has released from400,000 to 800,000 spring chinook (Oncorhynchus tshawytscha) each year. Smolts are reared in one of two treatments, 1) Optimum Conventional Treatment (OCT), or 2) Semi-Natural Treatment (SNT). In brief, OCT smolts are raised according to conventional hatchery practices, in a barren concrete raceway and surface fed by hand. SNT smolts are raised in similar raceways, with the raceway walls and floors painted to provide a varied colored background. Floating plastic hoops approximately 2 m in diameter and small submerged pine trees provide cover. Feeding is accomplished using an underwater feed delivery system. The objective for SNT is to attempt to produce smolts that are more similar to their wild counterparts in terms of coloration, utilization of cover, and feeding behaviors. A detailed description and rational of the OCT and SNT treatments can be found in the Spring Chinook Monitoring Plan, (Busack et. all., 1997).

Only wild reared fish are used for brood stock at CESRF, although the parentage of these fish may be wild, hatchery, or mix of fish that had spawned naturally in the wild. During spawning, each female's eggs are divided into 2 to 4 approximately equal groups, and each group fertilized with sperm from a different male. After fertilization, the eggs are re-combined and incubated in a heath tray system similar to other Pacific Northwest salmon hatcheries. The egg lots from females are summed together to produce groups of approximately 80,000 each, and these are divided equally between two raceways, one OCT, one SNT, thus producing an OCT/SNT raceway pair. Thus the parentage of each OCT/SNT pair is identical.

Northern Pikeminnow (Ptychocheilus oregonensis) have been identified as a major predator on outmigrating salmonid smolts in the Columbia River (NMFS, 2000.), and so were selected as the predator threat in this study. They are also abundant in the Yakima River, relatively easy to collect, and fairly tolerant of repeated handling and confinement in a laboratory setting.

As part of the overall monitoring efforts associated with the Yakima spring chinook supplementation effort, it is important to quantify and qualify differences produced by the various experimental regimes being tested at the Cle Elum Supplementation and Research Facility. The present experiment addresses these needs by comparing the behavior of CESRF's two rearing treatments with wild spring chinook smolts.

The goals of the present study was to assess differences between rearing treatments (OCT, SNT and Wild), in terms of 1) amount of time spent by smolts in cover after introduction into the tank, 2) effectiveness of cover utilization, 3) influence of the presents of other conspecifics on time smolts spent in cover, and 4) to quantify the smolt's susceptibility to a predator threat.

## Methods

Groups of five smolts from each of the three treatment groups, (Wild, OCT \& SNT), were placed sequentially into an aquarium. Cover and a predation threat was present in the aquarium. Typically, upon introduction, smolts will dive for cover and remain hidden for several minutes before emerging to explore or school with other smolts. Observers record the amount of time smolts spend in cover and make qualitative assessments of the smolt's cover utilization. Northern Pikeminnows, Ptychocheilus oregonensis, are then allowed to feed on the smolts until approximately $1 / 2$ are consumed. Surviving smolts are then counted and measured by treatment group. These procedures are described in more detail below:

Apparatus: The experiment was conducted in different locations using mostly different apparatus in the two years we ran the study. The following is a general description of the apparatus with year specific differences discussed at the end. The aquarium dimensions were 3.05 m long, 1.22 m wide and 0.91 m deep, with a volume of about 3400 liters. Cover objects were provided, these consisted of about 100 river rocks 5 to 15 cm in size, several irregular concrete pieces measuring about $15-20 \mathrm{~cm}$ and 2.5 cm thick, and small dead pine tree submerged on the bottom. The rocks, tree and concrete pieces were arranged to provide refugia for the smolts, with spaces between and underneath objects large enough for the smolts, but too small for the pikeminnows. Areas underneath the concrete blocks were arranged to be visible to observers through the tank windows.

Lighting was provided by a rack of 4 incandescent lights suspended over the aquarium which were controlled by both a dimmer and a timer switch. Normally, these lights were only used while making behavior observations, and room lights were also turned off during observations. Room lights were normally left off except when needed by us for setting up experiments. Some ambient light was also available through windows, thus fish were not subjected to changes in day length from their natural environment.

A 1HP irrigation pump was used to provide a current through the aquarium. Water was pulled from the drain on the bottom of one end of the aquarium, and pumped into the headworks at the other end. Hereafter, the drain end will be referred to as the "downstream" end, and the headworks the "upstream end". The headworks consisted of two parallel 50 mm diameter PVC pipes submerged at about 5 cm and 17 cm from the bottom. Water exited these pipes through a series of 7 mm holes pointed towards the far downstream end. Water velocities approximately 30 cm downstream of the headworks, where the smolts tended to school measured about $0.6 \mathrm{~m} \mathrm{~s}^{-1}$. The current degenerated rapidly into vortices and was barely perceivable in the downstream half of the tank.

Elastomer marked OCT and SNT smolts, spawned and reared at the Cle Elum Facility, were used for ease of identification. The elastomer mark was injected into the clear tissue behind the fish's left or right eye, identifying the fish as OCT or SNT. The side (left or right) and color (red, orange and green) is alternated between OCT and SNT and the three acclimation sites, (Easton, Clark Flat, and Jack Creek) each brood year. All hatchery fish were adipose fin clipped. These marks had been applied in October-November as a part of the marking program for all Cle Elum hatchery fish. For this experiment, a single raceway pair was selected each year as the source of OCT and SNT fish in order to reduce the genetic differences between these groups. Wild fish were collected at smolt tagging stations at Roza and Prosser dams, and had no clips or marks. All fish were kept in holding tanks for periods of 1 to 10 days before used in this experiment.

Northern Pikeminnows, Ptychocheilus oregonensis, were used as the predation threat in these studies. This species is considered a major predator of salmon smolts in the Columbia basin, (Poe, et all 1991, and Zimmerman 1999), and is the subject of a bounty program aimed at reducing their numbers, (see www.pikeminnow.org for details). These were collected via boat electroshocking from the Yakima River. While the experiment was in progress, pikeminnows were not fed except for the spring Chinook smolts they consumed during the predation phase of the experiment.

Only relatively large pikeminnow, $(350-560 \mathrm{~mm})$, were selected. These proved to be very hearty and resistant to handling mortality. Of the 16 original minnows captured in February 2000, and used at the Cle Elum facility, most were alive one year later, the only mortalities were two that jumped out of the aquarium.

Initially, we confined the pikeminnows in an area of the downstream end of the aquarium with a sliding partition during our observations of the smolts, and then release them to begin the predation test. We had assumed that the Pikeminnows would feed voraciously on the smolts, and thought we would need to block their access to the smolts while making the cover behavior observations. This expectation was not born out in the first year of the experiment. Generally, the pikeminnows showed little interest in the smolts while the lights were on, or when we were observing them. For all replicates presented in this paper, we did not employ sliding partition, thus allowing the pikeminnows access to the whole aquarium and smolts. This normally worked well, although three of the 180 smolts used were eaten during the observation periods of this experiment. Otherwise all predation that occurred when the room lights were dimmed below the point where we could make observations, or when we were not present.

Normally we allowed each replicate to continue until approximately $1 / 2$ of the smolts had been consumed. This normally took from 2 to 5 days. The exception was when we needed to cycle a replicate early in order to complete the entire set. This occurred in 2000 for replicate \#1 (2 hours), and replicates \#3 and \#4 (24 hours).

Groups of five smolts from each of the three treatment groups (OCT, SNT and Wild) were introduced separately into the observation tank. At the start of each replicate, smolts were netted from their holding tanks, and transferred to a 20 liter bucket, and given a 10 minute recovery period. The bucket was then half submerged and gently poured into behavior arena. Each group was observed for 30 minutes, then the next group was introduced. The order of introduction was counterbalanced across replicates to control for a hypothesized effect that the presents of smolts already in the tank would effect the cover seeking behaviors of newly introduced smolts. The order of introduction of the smolts was completely counterbalanced yielding 6 orders of introduction.

During the observation period we noted the position (in or out of cover) of each smolt, and the time when the smolt emerged from cover up to a maximum score of 30 minutes. The observation period for each replicate ended 30 minutes after the 5 smolts comprising the last group was added. At this point, approximately 90 minutes after the introduction of the first group of smolts, typically most or all of the 15 smolts would be out of cover and swimming in the tank with the majority of these schooling in an open area in the high velocity zone created by the recirculating pump.

At the end of the replicate, all surviving smolts were collected, anesthetized with tricaine methanesulfonate, (MS222), and measured. Initially, we did not measure smolts at the start of the replicate due to concerns that the anesthesia, stress and trauma would adversely affect the behavior study. This procedure, however, only gave us lengths of the surviving smolts, making it difficult to analyze length as a covariate to survival. We started with replicate \#10 in 2000, and continuing in 2001, we anesthetized and measured the smolts, and placed them in separate holding tanks for 24 hour recovery period prior to the start of the each replication. Specific smolts eaten by the pikeminnows was determined from individual fish lengths recorded prior to initiation of the experiment, which were deduced from the lengths of fish missing at the end of the experiment.
$\mathbf{2 0 0 0}$ v.s. 2001 Laboratory Setup. This experiment was initially set up in the incubation room at CESRF, and experiments run in 2000 were conducted there. Due to water permit problems and the over use of CESRF's well fields, this experiment was moved in 2001 to a wet lab at the Central Washington University's campus in Ellensburg, Washington.

In 2000, pumped ground water from the CESRF's well fie ld was used for the aquarium and fish holding tanks, the temperature was a nearly constant $9.8^{\circ} \mathrm{C}$. The water delivery system ensured the water was degassed and oxygenated. Water flow into the tanks was not measured, but flow through was sufficient to replace the total volume several times per day. In 2001, Ellensburg city water was used in a recirculating system, and water temperatures were controlled with chillers (Frigid Units, Inc, Model RT-430-F). The city water was dechlorinated and aerated for several days before use. Additionally, "Instant Ocean" sea salts were added to increase the salinity to about 0.1 ppt . Water temperatures were adjusted to match water temperatures in the Yakima River that the smolts were experiencing at the time of collection (about $14^{\circ} \mathrm{C}$.).

In 2000, a fiberglass tank with Plexiglas windows was used. The ability of observers to view all areas of this tank was compromised due to the spaces between windows. In 2001, this tank was substituted for a glass aquarium with a stainless steal bottom and framing, and of approximately the same dimensions as the original fiberglass tank. This new tank afforded much improved abilities to observe the smolts.

In 2001, a "predation only" tank was added to the experiment, to increase the number of replicates and experimental power to differentiate the survival rates between groups of smolts. This was a modular Living Stream fiberglass tank, which we arranged to form a 2700 liter donut shaped tank. This tank lacked good observation windows, so no behavior observations were attempted. An airlift style water pump was constructed of PVC pipe fitted with an airstone, and this was used to both oxygenate and circulate water around the tank. Due to concerns that gravel would damage this tank, gray and black colored PVC pipe fittings of various sizes and shapes were used to provide cover. These pipe fittings were thought to provide comparable if not greater refugia than did the river gravel. Procedures for handling and measuring the smolts were the same as for the glass tank. The introduction method differed though, as we were not making behavior observations on these groups. Instead, after the initial measurement, all smolts were placed together in section of the tank separated from the pike minnows by removable barriers. The smolts were given 24 hours to recover from the anesthesia and handling before the barriers were removed to allow the pikeminnows access to the smolts, and to allow the smolts access to all areas of the tank.

## RESULTS

Typically, upon introduction, the smolts immediately dove to the bottom of the tank, and those that chose to hide under cover would do so in the first 10-15 seconds, where they would remain for periods up to an hour before emerging. Typically, once emerged, the smolts would swim to an open area of the tank just downstream of the head box that had current provided by the recirculating pump. There they would maintain station 5 to 20 cm above the bottom. They generally remained in that area for the duration of the experiment, though sometimes would explore the rest of the tank. Occasionally the smolts would return to cover for periods of time, but this was the exception rather than the rule.

In general, wild smolts tended to be in or close to cover mo re often during the observation periods than the two hatchery groups following introduction to the aquarium. Also, many wild smolts were observed to lay on the bottom, in narrow places between stones or under the substrate at times without appreciable movement. They also appeared more adept at finding crevices or places to hide under rocks. Contrarily, hatchery fish appeared to take longer, and be less adept at finding hiding spots, and when in hiding spots they had nearly constant fin movements, and were rarely perfectly still. The difference in hiding and remaining motionless appeared to be one of degree, not kind.

Time to Leave Cover. The time for the smolts to leave cover during the 30 minutes following their introduction to the aquarium was analyzed in two ways. First using a 3X6X2 factorial, which combined the three rearing conditions with the six counterbalanced orders of placement, and the two years (2000 and 2001) that the experiment was conducted in. The 6 counterbalanced orders of placement represented all the possible orders three treatment groups, for example Wild, SNT, OCT was one order, and SNT, Wild, OCT was another order). The resulting 36 conditions contained 5 smolts each ${ }^{1}$. A second analysis recoded the order variable to represent the order that each group of fish was introduced (first, second or third). This resulted in a 3X3X2 factorial design, with 18 conditions of 5 smolts each.

The order in which the smolt groups were introduced appeared to make little difference. Using either the 3X6X2 or 3X3X2 factorial design, the main effect of order fails to reach statistical significance ( for the 3X6X2 design, $\mathrm{F}(1,144)=0.481, \mathrm{p}=0.790$. For the 3 X 3 X 2 design, $\mathrm{F}(1,162)=0.218, \mathrm{p}=0.804$. The interaction between treatment and order also fails to reach statistical significance. For the 3 X 6 X 2 design, $\mathrm{F}(10,144)=1.321, \mathrm{p}=0.224$. For the 3 X 3 X 2 design, $\mathrm{F}(4,166)=0.499$, $\mathrm{p}=0.736$.

The weighted means analysis indicated that in the trials conducted in 2001, the smolts left cover significantly sooner than smolts did in 2000. The average time in 2000 was 11.1 minutes, but only 5.4 minutes in 2001 . Using the $3 X 6 X 2$ design, the main effect of Year gives $\mathrm{F}(1,144)=17.084, \mathrm{p}=0.000$. Interactions between year and treatment did not reach statistical significance $\mathrm{F}(2,144)=1.310, \mathrm{p}=0.273$.

[^35]Several changes in the experimental design probably account for this difference, 1) the 2001 smolts were older than their 2000 counterparts, as the 2000 observations were performed mostly in April, whereas the 2001 replicates were conducted mostly in May of that year, thus the 2001 smolts were about a month older during a critical time in their normal outmigration and smoltification stage of development. A second difference was in water temperatures, in 2000 all trials were conducted at about $10^{\circ} \mathrm{C}$, whereas in 2001 water temperatures were set to $14.4^{\circ} \mathrm{C}$, and a third may be due to the configuration of the observation aquariums.

The effect main effect of treatment was seen to be significant, $\mathrm{F}(2,144)=4.148, \mathrm{p}=0.018$. The results are summarized in Figure 1. Using a subsequent pairwise comparison (SYSTAT Tukey HSD multiple comparison procedure), the SNT fish were found to have spent significantly less time in cover than did the wild for both years combined ( $\mathrm{p}=0.023$ ). However when analyzing the data separately by year, the comparisons are more ambiguous. In 2001 the difference between the wild and SNT only approached significance ( $\mathrm{p}=0.072$ ), and no difference was seen between the OCT and Wild. In 2001, the difference between the OCT and SNT approached significance, but no difference was seen between either of the two hatchery group and the wild group, but the difference between OCT and SNT approached significance ( $\mathrm{p}=0.093$ ), (Table 1.) In both years, the SNT spent the least amount of time in cover of the three groups.


Figure 1. Average Time Spent in Cover after Introduction to the Aquarium. (all replicates, both years). Standard error bars of the means are show.

| Tukey HSD Multiple Comparisons. |  |  |  |
| :---: | :---: | :---: | :---: |
| Matrix of pairwise mean differences: <br> (Minutes in cover) | Matrix of pairwise comparison probabilities: |  |  |
| Both years: Using model MSE of 99.720 with 177 df. |  |  |  |
| OCT SNT WILD | OCT | SNT | WILD |
| OCT 0 | OCT 1 |  |  |
| SNT -3.185 0 | SNT 0.19 | 1 |  |
| WILD 1.5964 .782 | WILD 0.656 | 0.023 |  |
| For 2000: Using model MSE of 121.108 with 87 df. |  |  |  |
| OCT SNT WILD | OCT | SNT | WILD |
| OCT 0 | OCT 1 |  |  |
| SNT -2.113 0 | SNT 0.742 | 1 |  |
| WILD $4.169 \quad 6.282 \quad 0$ | WILD 0.312 | 0.072 |  |
| For 2001: Using model MSE of 62.518 with 87 df. |  |  |  |
| OCT SNT WILD | OCT | SNT | WILD |
| OCT 0 | OCT 1 |  |  |
| SNT -4.312 0 | SNT 0.093 | 1 |  |
| WILD $-1.169 \quad 3.142$ 0 | WILD 0.835 | 0.278 |  |

Table 1. Pairwise mean differences between treatment groups (left) and the statistical significance of those differences (right). Table shows mean differences for all replicates combined, and for 2000 and 2001 analyzed separately. Bolded red highlights comparisons found significant at $p<0.05$, and violet indicates comparisons only approaching significance, $0.05<p<0.10$.

Survival to a predation challenge: A total of 18 replications with smolts from each of the three rearing conditions were run to assess relative survival rates to a predation challenge (Northern pikeminnows). Fourteen replicates had five smolts from each rearing condition, and three replicates only three or four due to mortalities during the acclimation period ${ }^{2}$. One replicate in 2000 had six wild and four OCT smolts due to a wild smolt jumping from its holding tank to the OCT tank. Six replicates were run in 2000, and 12 in 2002. Data from a total of 261 smolts are presented in these results. In 2001 we measured all smolts before the start of each replicate in order to be able to analyze length as a covariate to survival. Unfortunately, we only measured surviving smolts in 2000, and so are unable to analyze that data using length as a covariate as we did not have length data on the smolts that were consumed.

Overall, using data from both years, the effect of rearing condition on survival failed to reach significance ( $\mathrm{F}(2,252$ ) $=0.704, \mathrm{p}=0.496$ ). That overall survival rates were higher in 2000 is an artifact of the experimental conditions, water temperature at the Cle Elum facility were only $10^{\circ} \mathrm{C}$, to cold to induce sufficient feeding activity in the pikemouth for us to consistently obtain the target level of $50 \%$ consumption rates in the time window for this experiment. In 2001 we were able to manipulate water temperatures to mirror temperatures in the river, which also had the effect of increasing pikeminnow feeding rates.

Using the 2001 data, where survival vs. length comparisons could be made, the effect of fork length on survival was found to be significant $(\mathrm{F}(1,164)=8.514, \mathrm{p}=0.004)$, with larger ${ }^{3}$ hatchery smolts surviving at higher rates than smaller ones, (Figure 2). An ANOVA on the survival of small smolts, ( $<130 \mathrm{~mm}$ ), found a significant effect on rearing condition, $(\mathrm{F}(2,89)=3.357, \mathrm{p}=0.039)$. This relationship did not hold for larger smolts, $(\mathrm{F}(2,70)=1.202$, $\mathrm{p}=0.307$ ). A subsequent comparison of the mean survival of small smolts found that the Wild had a $31.6 \%$ higher survival rate than did the OCT smolts, (Table 2).

[^36]

Figure 2. Survival of large (red) and small (blue) smolts to a predation challenge. Mean survival and standard error bars are shown. Numbers represent sample size. Trend line and error bars for large wild fish are not show due to small sample size ( 6 fish).

| PAIRWISE MEAN | DIFFERENCES |  |  | PAIRWISE | COMPARISON PROBABILITIES |  |  |  |
| ---: | :---: | :---: | :--- | :--- | :--- | :--- | :--- | :--- |
| OCT | SNT | WILD |  |  | OCT | SNT | WILD |  |
| OCT | 0.000 |  |  |  | OCT | 1.000 |  |  |
| SNT | 0.098 | 0.000 |  | SNT | 0.774 | 1.000 |  |  |
| WILD | 0.316 | 0.218 | 0.000 | WILD | 0.033 | 0.172 | 1.000 |  |

Table 2. Tukey HSD Multiple Comparisons of survival rates for small (<=130mm) smolts).

The wild smolts used in 2001averaged about 11 mm smaller than the hatchery smolts, ( 123 mm compared to 134 for the OCT and SNT combined. This difference was significant $(\mathrm{F}(2,171)=20.130, \mathrm{p}<0.001)$. The difference between OCT and SNT smolts was small ( 133.9 vs. 134.7 mm ) and fails to reach statistical significance.

Survival rates of OCT and SNT smolts were similar within size classes. When these treatment groups are pooled we find a significant effect of size class on survival, $(\mathrm{F}(3,109)=4.421, \mathrm{p}=0.006)$, with the largest size class measured enjoying a $41 \%$ higher survival rate than the smallest size class, (Figure 2). The survival rate for wild fish in Figure 2 appears negatively correlated with size, however this trend fails to reach statistical significance.


Figure 2 Length vs. smolt survival to a Northern Pikeminnow predation challenge, 2001. Left graph shows length frequencies of smolts used in the experiment, right shows the percent surviving by size class. Yellow bars on the right graph show average survival across treatments for each size class. There were no wild smolts in the largest size class (141-158mm).

## DISCUSSION

The analysis of the amount of time spent in cover after introduction by treatment group showed that the wild smolts tended to stay in cover longer than either the SNT fish and the two hatchery-reared groups did not significantly differ from one another on this measure. In addition to this quantitative evidence of superior use of cover by the wild smolts, several qualitative observations also support the view that the wild smolts are more adept at using cover relative to their hatchery-reared counterparts. For example, wild smolts were observed to be in tight proximity to cover, often under cover objects, more often than hatchery smolts who tended, in the main, only to be close to cover rather than "in it." Also, when in cover, wild smolts tended to remain motionless, resting on the substrate, whereas hatchery smolts tended to paddle with their fins even when being more or less stationary. Because of the constant movement, and perhaps larger size, we found hatchery smolts generally easier to locate, even when in the same hiding places as the wild. It is probable that visually oriented predators would also find hatchery fish more easily for the same reasons.

The analysis of the time for the smolts to leave cover after introduction into the test tank also revealed no significant effect of order of introduction and no interaction between rearing condition and order. Thus, smolts of a specific rearing group were not influenced to leave cover by the presence or absence of other smolts in the tank. This was contrary to our expectations, as Chinook Salmon are known for their schooling behaviors. For example in one study, observers found that hatchery salmon drifting downstream from their release site tended to "pull" wild salmon from cover in what has become known as the "pied piper effect":
"Upon seeing the hatchery salmon, wild salmon and steelhead moved from their stations and
aggregated with hatchery fish. When aggregated with hatchery salmon, wild salmon and steelhead
appeared to behave like hatchery fish, with no apparent social structure, and they moved with the
hatchery fish in to areas wild fish would not normally use." (Hillman and Mullan, 1989).
Other examples of behavioral interactions between groups of fish abound in the literature, and indeed we observed a number of interactions between the salmon smolts once they emerged from cover and became more acclimated to their new surroundings. Thus the lack of effect of other fish in the tank on time to leave cover suggests that the fishes initial dive for cover and time spent there represents a response to the smolt's involuntary introduction to a novel environment, and the time spent hiding and acclimating to the new environment is not influenced by the presents or absence of conspecifics.

Several authors have found differences in coloration between hatchery and wild fish, and attributed increased vulnerability of hatchery fish to decreased crypsis in stream environments, (Maynard et all., 1996 and Donnelly and Whoriskey 1991, 1993). Indeed, one of the objectives of the SNT treatment at the CESRF was to raise fish in raceways painted with varied naturalistic colors in an attempt to produce fish who's coloration was more like their wild counterparts. However we could not discern any consistent difference in coloration between the OCT, SNT or Wild fish, although variations were observed within each group.

Northern Pikeminnows are thought to be mainly corpuscular feeders. A study of movements of radio tagged pikeminnow below two dams on the Snake river showed that these fish are most active during the corpuscular, (sunrise and sunset), periods of the day, and least active immediately following these periods, and intermediately active during the rest of the day, (Isaak and Bjornn, 1996). Another study that extrapolated diurnal feeding patterns from stomach contents found differences in diurnal consumption rates between locations in the John Day dam pool on the Columbia River, (Vigg, et. all., 1991a). Within 1 km downstream of the McNary dam, peak feeding rates by Northern Pikeminnow was observed from dawn to around 0900 AM, and again from midnight to around 0300 AM. However, in the rest of the John Day dam pool (from 1 km downstream of the McNary dam to the John Day dam), feeding rates of Northern Pikeminnow although showed a diurnal pattern, did not appear to be as corpuscular as in the dam tailrace. In that part of the river, feeding rates were seen to increase from sunrise through the day, then dropped to a low level toward evening and remaining low through the night. No mention was made in these papers how the affects of artificial lights around the McNary dam may have influenced pikeminnow activity, or how the presents of injured and disoriented smolts in the tailrace may have affected diurnal feeding patterns.

In our observations, the pikeminnow tended to rest on the bottom when room lights or tank lights were on, and showed little interest in the smolts. However, when light levels were lowered to the point where we could only see vague outlines of fish, pikeminnow were observed to raise off the bottom, and stalk the smolts from underneath and behind. It is doubtful, in our opinion, that cryptic coloration would play a significant roll in predator avoidance in these circumstances. Other behaviors, such as the smolts ability to detect pikeminnow in close proximity, or escape response, positioning relative to other smolts, or ability to avoid detection through lack of movement probably explain the survival advantage observed in the wild fish.

The analysis of the survival data showed an overall similar survival between the three groups of fish, and this was seen in both years the experiment was conducted. In the 2001 data, where survival could be analyzed as a function of size, we saw a significant positive correlation of survival with fish fork length for hatchery fish, but not for wild. The largest size group of hatchery fish had a survival rate comparable to the wild, however the survival rate of the smaller hatchery fish was only about half the wild rate.

The largest Pike minnows available were intentionally selected for use in this study, as previous studies have reported that salmonids were generally an important diet component only for large, older northern pikeminnows (Vigg et al. 1991) and that consumption rates of juvenile salmonids by northern pikeminnow increased exponentially as size increased (Beamesderfer et al. 1996). The length frequencies of Northern Pikeminnow used in this study compared to the overall length frequency distribution of Northern Pikeminnow in the Yakima are shown in Figure 3, (left).

The predicted prey diet composition of the Yakima Northern Pikeminnow population and the sizes of Spring Chinook smolts are shown in Figure 3 (right). In the wild, pikeminnow of the sizes used in this study can and do consume prey items up to the largest sized smolt used in this study $(240 \mathrm{~mm})$, however the majority of their diet consists of fish considerably less than the average sized smolt ( 130 mm ). Data on predator vs. prey lengths from fish captured in the wild can suggest what size of fish a predator targets, however it is not definitive as it is not known precisely what sizes of fish were available to the predator. There is also evidence that body depth is more important than length to piscivorous fish in selecting prey, (Hambright, 1991). Data from this experiment shows that the Pikeminnows preferentially preyed on the smaller fish, at least in this experimental setting where only one species of prey was available, and in a limited size range.


Figure 3. Left, comparison of pikeminnow sizes in the Yakima River (blue) to thos e used in this experiment, (yellow). Right, predicted size composition of pikeminnow diets (blue) to the size of wild and hatchery spring chinook smolts captured at Roza dam in 2000 (yellow and pink). The predicted NPM diet composition is based on the size distribution of Pikeminnows in the Yakima (left graph) and the diet composition of various sized Pikemouth shown in Figure 4.

The Cle Elum hatchery OCT and SNT smolt population, are in fact larger than the wild smolt population. Of fish passing Roza in March through May of 2000, hatchery smolts averaged 116.9 mm , and wild smolts were 111.5 mm , a small, but statistically significant difference. By the time they reached Prosser Dam, the fish averaged 137.8 and 128.1 mm .


Figure 4. Prey size vs. Northern Pikeminnow sizes for pikeminnow captured in the Yakima River 1998-2000. Pink are salmonids, blue are non salmonid fish consumed by pikeminnow. Length of prey items are directly measured when relatively whole, or estimated from bones when partially digested.

Survival of Out migrating Smolts. Smolt monitoring programs operated by the Yakama Nation Fisheries Program are used to compare relative survival rates of hatchery and wild spring Chinook salmon smolts using passive integrated transponders (PIT tags). The PIT tags provide a unique identification number for each fish, and the tag can be interrogated as the fish passes through smolt monitoring stations designed for this purpose. There is one such site on the Yakima river at Prosser Dam, and three on the Columbia River, McNary Dam, John Day Dam, and the Bonneville Dam. The McNary Dam provides some of the best information for comparing smolt survival of Yakima river releases because McNary detection efficiencies can be estimated using further downstream detections at Bonneville and John Dams. These detection efficiencies can be used to expand the PIT-tag detection rates at

McNary Dam to obtain an index of survival. This expanded survival estimate is considered an index, not as an absolute figure, as not all of the assumptions that go into the model are likely to hold, (Neeley 2002a and 2002b).

Approximately 40,000 CESRF hatchery fish are PIT tagged each year before release, the tags allocated equally between raceways and treatments. Downstream detections of these fish can be used to compare survival of the OCT vs. SNT treatment groups. The survival index of outmigrating smolts to McNary for SNT fish has been greater than OCT fish in each of the three years 1999 through 2001, (Table 3). The difference is small, only $0.8 \%$ averaged over three years, and fails to reach statistical significance, ( $\mathrm{p}=0.26$ ), (Neeley 2002b).

Table 3. Survival index of outmigrating OCT and SNT fish from the acclimation ponds to McNary Dam. A weighted logistic analysis of variation was used to test for survival differences, using acclimation site, treatment and year as factors, (Neeley, 2002b).

| Out - <br> Migration <br> Year | McNary Survival Index |  |  | Difference (SNT - <br> OCT) |
| :---: | :---: | :---: | :---: | :---: |
| 1999 | $50.61 \%$ | $50.13 \%$ | Type 1 <br> error, P |  |
| 2000 | $38.12 \%$ | $37.75 \%$ | $+0.48 \%$ | 0.41 |
| $\mathbf{2 0 0 1}$ | $\mathbf{2 4 . 5 5 \%}$ | $\mathbf{2 3 . 0 1 \%}$ | $+0.37 \%$ | 0.36 |
| Mean | $37.76 \%$ | $36.96 \%$ | $+\mathbf{1 . 5 4 \%}$ | $\mathbf{0 . 0 3}$ |

Differences in survival indexes between years are probably explained by environmental conditions. 1999 was a year of higher than normal flows during the outmigration period, with accompanying cooler temperatures and more turbid conditions. 2000 was an average water year, and 2001 was a year of severe drought.

Another monitoring project involves the collection and tagging of wild and hatchery fish at Roza Dam, which is 60.3 kilometers downstream of Clark Flat, the lowest of the three acclimation sites. At Roza Dam a portion of the river is diverted into a canal for power generation and irrigation purposes. Fish screens mounted in the canal forebay prevent fish from entering the canal, and these fish are bypassed back to the river. Outmigrating smolts can be captured in a holding pen mounted in the bypass unit.

The Roza smolt tagging operation has been run each year since 1999. Wild smolts are PIT tagged, and hatchery smolts are scanned for the presents of an existing PIT tag, and those fish without pre-existing PIT tags are PIT tagged. This results in three groups PIT tagged fish, wild, hatchery smolts tagged at Roza, and hatchery fish PIT tagged several months previously at CESRF. A logistic analysis of variance variation was used to compare relative survival rates to McNary dam of hatchery and wild fish tagged and released at Roza, (Table 4).

Table 4. Survival indexes and difference in sizes of hatchery and wild spring smolts. Fish are tagged and released at Roza Dam and interrogated 697 kilometers downstream at McNary Dam. Indexes shown only include data from time periods where estimates of both wild and hatchery survival were available. Bolded numbers show where differences were statistically significant, (from Neeley 2002b).

| Year | McNary Survival Index* |  |  | Average size at Tagging (mm) |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Wild | Hatchery | Difference | Wild | Hatchery | Difference |
| 1999 | $61.1 \%$ | $50.9 \%$ | $10.2 \%$, <br> $\mathrm{p}=0.252$ | $\mathbf{1 2 1 . 0}$ | $\mathbf{1 2 6 . 4}$ | $\mathbf{- 5 . 4}$ <br> $\mathbf{p}=\mathbf{0 . 0 0 7}$ |
| 2000 | $\mathbf{4 7 . 6 \%}$ | $\mathbf{3 0 . 6 \%}$ | $\mathbf{1 7 . 0 \%}$, <br> $\mathbf{p}=\mathbf{0 . 0 0 2}$ | 110.4 | 110.7 | -0.3 <br> $\mathrm{p}=0.750$ |
| 2001 | $\mathbf{1 5 . 1 \%}$ | $\mathbf{2 2 . 2 \%}$ | $\mathbf{- 7 . 1 \%}$, <br> $\mathbf{p}=\mathbf{0 . 0 0 8}$ | $\mathbf{1 1 8 . 0}$ | $\mathbf{1 2 8 . 7}$ | $\mathbf{- 1 0 . 7}$ <br> $\mathbf{p}<\mathbf{0 . 0 0 1}$ |
|  |  |  |  |  |  |  |
| *Mean over common blocks (adjusted). |  |  |  |  |  |  |

Wild fish significantly outperformed hatchery smolts in 2000, when their mean sizes were similar. Hatchery fish though outperformed wild in 2001 when their mean size was significantly larger. There was no significant difference in survival or migration timing of the hatchery fish tagged at Roza, and those previously tagged at CESRF, indicating the PIT tagging itself does not induce a significant short term mortality or substantially effect
outmigration timing. This is not to say that the process of capturing and handling fish has no effect, as this could not be tested with these data. The results of the two hatchery groups are shown pooled in the above table.

Adult Returns. As of this writing, there is only one year of adult 4-year-old return data to compare, that of the 1999 smolts which returned in 2001. The overall return rate to Roza Dam of the SNT was $1.42 \%$ and $1.31 \%$ for OCT, (Neeley 2002b). The SNT outperformed the OCT in 4 out of 5 raceway pairs, and the overall return rate was about $8.4 \%$ higher for the SNT. This was marginally significant ( $\mathrm{p}=0.07$ ) based on an approximate 1 -sided t -test. All returning hatchery fish are interrogated at Roza, but only those wild fish that are collected for brood stock are interrogated, the majority released without scanning for PIT tags, making it difficult to compare hatchery vs. wild adult return rates using Roza adult data.

In conclusion, it appears from both our study and monitoring studies of out migrating smolts that wild fish enjoy a survival advantage over hatchery fish when their sizes are similar, and that there is little appreciable difference between the OCT and SNT treatments in smolt survival in our laboratory or from release site to McNary Dam.

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## APPENDIX F

2002 Annual Report Indirect Predation

# IntSTATS 

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# 2002 Annual Report <br> Indirect Predation 

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Submitted April 17, 2003

## Summary

There is evidence from 2002 outmigrants that survival to McNary Dam of hatcheryproduced Spring Chinook Smolt increases with an increase in the number of fish volitionally exiting acclimation ponds. Even though survival for these 2002 outmigrants also increased with increased release-site stream flow, the relation of survival to fish number appears to be independent of stream flow's effect.

There was no evidence of a change in survival with a change in volitional release number for 2000 and 2001 outmigrants.

## Introduction

The indirect predation study was initially envisioned as a study to determine whether survival was partially a function of the number of outmigrating smolt. The concept was that either 1) a large number of outmigrating smolt might confuse predators, thereby the increasing smolt survival, or 2) a large number of outmigrating smolt might attract a predator swarm, thereby decreasing smolt survival. It would be difficult, perhaps impossible, to develop an in-river experiment that could test these hypotheses.

Bruce Watson ${ }^{1}$ felt that it would be possible to use smolt survival to McNary Dam on the mid-Columbia River for PIT-tagged smolt detected in the bypass system at Prosser dam on the lower Yakima River to assess the effect of indirect predation. Under my recommendation, the statistical tool he used was to regress ${ }^{2}$ estimates of survival to McNary of fish passing Prosser on estimates of total smolt passage at Prosser. The notion was that a significant positive regression-coefficient estimate on flow would

[^37]indicate that survival would tend to increase with fish number and that a significant negative regression-coefficient estimate would indicate that survival would tend to decrease with increasing fish number. Unfortunately it proved difficult to isolate the effect of fish passage on survival from the effects of other uncontrollable factors such as flow ${ }^{3}$.

A different method was considered for the year-2000 through year-2002 Spring Chinook hatchery outmigrants and is presented here. Hatchery Spring Chinook volitionally exit acclimation sites. There is little variation in flow into the acclimation ponds. Therefore, survival to McNary for fish leaving the ponds within prescribed time periods were related to the number (release number) of fish volitionally leaving during those periods under the assumption the release number would not be related to stream flow. The decision was made to partition the volitional release period into equal releaseperiod segments, arbitrarily set at ten continuous days each. Estimates of smolt survival to McNary and the number of fish exiting the acclimation ponds for each release period constituted the primary database. The coefficient resulting from the logistic regression of smolt survival to McNary on the number of fish leaving the raceway was used a measure of the effect of fish number on size.

## Methodology and Findings

The methods of estimating survival and the regression methods are discussed in Appendix A.

The starting date used for the equal length periods was March 15, which was the date on which the exit screens from the acclimation ponds were pulled. The periods used in the analysis were March 15-24; March 25-April 3; April 4-April 13, April 14-April 23, April 24-May 3, May 4-May 13, and May 14-May 23. There was evidence of fish leakage from the acclimation ponds prior to March 15 in outmigration 2000; however, since free access to the river was not possible during this period because the screens had not been pulled, pre-March $15^{\text {th }}$ releases were not used. For all years, fish remaining in the ponds after May 23 were not used because some of the fish had to be eventually forced out of the ponds; therefore, their release could not be considered to be volitional. Further, there has been a high rate of precocialism in these hatchery fish populations; and, if precocial fish tend to remain in the ponds, the survival index of fish leaving after May 23 may be artificially low because the precocial fish may have survived without outmigrating to McNary Dam.

The databases used in the logistic regression are presented in Table 1. The underlying data used to develop the database in Table 1 are given in Appendix B. The weighted ${ }^{4}$ logistic analyses of variation for the survival index regressed on site indicators

[^38]and number of fish released and stream flow are presented in Table 2 along with other measures Data and analysis summaries are presented for the following outmigration years: a. 2000, b. 2001, and c. 2002. Analysis was not possible for outmigration-year 1999 (the first year of acclimation site operations) because volitional release monitoring was not possible.

Outmigration- year 2002 was the only year in which the regression of survival index on release number was significant ( $\mathrm{P}<0.001$, Table 2.c.). In that year, the sign of the logistic regression coefficient was positive, suggesting an increase in survival with increasing number of fish. However, as indicated in Table 1, within each acclimation site ${ }^{5}$ within each year, the Pearson's correlation coefficients between release number and stream flow were positive (significantly so for some sites in outmigration- years 2000 and 2001, $\mathrm{P}<0.05$; Tables 1.a. and 1.b.). The implication is that fish number increased as stream flow increased even though water flow into the ponds was kept relatively constant. Even if flows into the acclimation ponds do not vary with stream flow, it is possible that there are stream-flow clues in the water entering the ponds of flow streamflow changes that stimulate fish to move out of the ponds when stream flows increases (e.g., chemical changes or turbidity changes in water entering the pond that may be related to stream flow changes).

Because of the correlations between release number and stream flow, the survival index was regressed on flow as well as release number. Daily flows were obtained from the Bureau of Reclamation's WEB site,

## http://mac 1.pn.usbr.gov/yakima/yakwebarcread.html,

for the bureau's monitor sites on the Yakima River at Easton and at Cle Elum and on Teanaway Creek near its confluence with the Yakima near Cle Elum. Weighted ${ }^{6}$ flow estimates are also presented in Table 1. Based on a conversation with Mark Johnston ${ }^{7}$, the Easton monitor-site flows were used for the Easton acclimation site, the combined Cle Elum and Teanaway monitor-site flows were used for Clark Flats acclimation site, and the Teanaway monitor-site flows were used for the Jack Creek acclimation site. Another flow- monitor site further upstream on the Teanaway (below Lambert) may have been a more appropriate flow- monitoring site for the Jack Creek acclimation ponds; however, that flow-monitoring site did not have enough flow data points from the three years to be useful.

As was the case for the logistic regression of survival index on release number, only in outmigration-year 2002 was the logistic regression on flow significant; and, again as with number released, the logistic coefficient of survival index on flow was positive,

[^39]indicating that as flow increased, survival increased. To get a handle on whether the survival index was affected by number released independent of flow's effect, survival index was regressed on both number released and flow. The effect of release number on the survival index adjusted for the effect flow ("release number | flow" in Table 2) as well as the effect of flow on the survival index adjusted for the effect release number ("flow | release number" in Table 2) was then assessed. Both adjusted coefficients are significantly different, suggesting that, if the logistic model used accurately portrays the relation between survival index and the predictor variables (number volitionally released and stream flow), then 1) an increase in release number may have increased the survival index above what might have been attributed to an associated increase in flow, and 2) an increase in stream flow may have increased the survival index above what might have been attributed to an associated increase in release number. Further, the adjustment for flow had almost no effect on either the magnitude or the significance level of the logistic regression coefficient of survival index on release number (unadjusted-for-flow coefficient $=0.000118$ and adjusted- for-flow coefficient $=0.000112$; unadjusted - forflow $\mathrm{P}=0.0005$ and adjusted-for-flow $\mathrm{P}=0.0008$, Table 2.c.). These findings, combined with the fact that Pearson's correlation coefficients between number released and stream flow were of low magnitude and were not significant for all three sites in outmigration year $2001(\mathrm{P}=0.309, \mathrm{P}=0.952$, and $\mathrm{P}=0.274$ respectively for Clark Flats, Easton, and Jack Creek, Table 1.c.), indicates that the year-2002 increase in the survival index with increase release number is independent of stream flow. However, it should be borne in mind that, for the three years of study, this relation was only manifest in outmigrationyear 2002.

It should be noted that in outmigration- year 2002, individual acclimation pond survival to McNary Dam was associated with the severity of Bacterial Kidney Disease ${ }^{8}$ (BKD) measured on a pond-by-pond basis (2002 Annual Report: OCT-SNT Survival by Doug Neeley). However, since the analyses presented in this report are based on data pooled over ponds within acclimation site, I know of no reason to suspect that the BKD variability biased this reports findings.

Figures 1.a through 1.c. respectively plot for outmigration-years 2000, 2001, and 2002 the survival indices against stream flows and the resulting logistic fit unadjusted for flow.

[^40]Table 1. Release-to-McNary Survival-Index Estimates for Fish Volitionally Exiting Acclimation Ponds during Ten-Day Intervals along with Number of Fish Released and Flow Estimates (estimates after May 23 not used in assessment).

## a. Year-2000 Outmigrants

| Clark Flats |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Ending <br> Release <br> Date | Number <br> Released <br> (Rel) | McNary <br> Passage <br> (McN) | Survival <br> Index (McN/Rel) | Weighted* Flow |
| 3/24/00 | 327 | 98.2 | 0.30039261 | 1613.8 |
| 4/3/00 | 572 | 261.1 | 0.45655581 | 3109.1 |
| 4/13/00 | 1780 | 646.3 | 0.363093223 | 3512.1 |
| 4/23/00 | 2458 | 807.8 | 0.328658083 | 3543.9 |
| 5/3/00 | 1063 | 335.4 | 0.315528427 | 2177.3 |
| 5/13/00 | 1114 | 467.6 | 0.419775291 | 1920.7 |
| 5/23/00 | 1840 | 1007.5 | 0.547532016 | 2632.4 |
| Clark | Rel | Correlation between Num. Rel., Flow |  |  |
| Flats | 9154 | Correlation | 0.673671 | Type 1 P |
|  |  | DF | 6 | 0.0670 |
| Easton |  |  |  |  |
| Ending <br> Release <br> Date | Number <br> Released <br> (Rel) | McNary <br> Passage <br> (McN) | Survival <br> Index <br> (McN/Rel) | Weighted* Flow |
| 3/24/00 | 113 | 22.6 | 0.200297691 | 422.4 |
| 4/3/00 | 2378 | 839.4 | 0.352975655 | 788.9 |
| 4/13/00 | 4691 | 1408.3 | 0.300216357 | 979.6 |
| 4/23/00 | 997 | 250.5 | 0.251260483 | 737.1 |
| 5/3/00 | 584 | 218.7 | 0.374557397 | 478.9 |
| 5/13/00 | 386 | 148.0 | 0.38337832 | 589.1 |
| 5/23/00 | 632 | 261.7 | 0.41404838 | 552.4 |
| Easton | Rel | Correlatio | between Num | el., Flow |
|  | 9781 | Correlation | 0.918252 | Type 1 P |
|  |  | DF | 6 | 0.0013 |
| Jack Creek |  |  |  |  |
| Ending <br> Release <br> Date | Number <br> Released <br> (Rel) | McNary <br> Passage <br> (McN) | Survival <br> Index (McN/Rel) | Weighted* Flow |
| 3/24/00 | 340 | 95.8 | 0.28189823 | 361.3 |
| 4/3/00 | 1078 | 284.7 | 0.264090294 | 1131.6 |
| 4/13/00 | 810 | 249.0 | 0.307394516 | 1558.6 |
| 4/23/00 | 1055 | 359.3 | 0.340538197 | 1219.8 |
| 5/3/00 | 447 | 164.6 | 0.368129296 | 655.4 |
| 5/13/00 | 190 | 95.4 | 0.502074178 | 466.3 |
| 5/23/00 | 1847 | 884.6 | 0.478926283 | 970.0 |
| Jack | Rel | Correlation between Num. Rel., Flow |  |  |
| Creek | 5767 | Correlation | 0.554903 | Type 1 P |
|  |  | DF | 6 | 0.1534 |
| Weighted | Rel | Correlation between Num.Rel., Flow |  |  |
| over | 24702 | Correlation | 0.742787 | Type 1 P |
| Sites |  | DF | 18 | 0.0002 |

*Weight is daily number released applied to daily flow

Table 1. (continued)

## b. Year-2001 Outmigrants

| Clark Flats |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Ending <br> Release <br> Date | Number <br> Released <br> (Rel) | McNary Passage (McN) | Survival <br> Index (McN/Rel) | Weighted* <br> Flow |
| 3/24/01 | 1043 | 312.8 | 0.29993 | 1104.9 |
| 4/3/01 | 1063 | 319.4 | 0.30047 | 1501.9 |
| 4/13/01 | 1164 | 407.4 | 0.34998 | 1132.5 |
| 4/23/01 | 1827 | 570.4 | 0.31223 | 1080.0 |
| 5/3/01 | 4361 | 1,171.4 | 0.26862 | 1879.9 |
| 5/13/01 | 1583 | 211.7 | 0.13376 | 1511.9 |
| 5/23/01 | 1814 | 256.8 | 0.14156 | 1464.8 |
| Clark <br> Flats | Num. Rel. $12855$ | Correlation between Num.Rel., Flow  <br> Correlation 0.74246 Type 1P |  |  |
| Easton |  |  |  |  |
| Ending <br> Release <br> Date | Number <br> Released <br> (Rel) | McNary Passage (McN) | Survival <br> Index (McN/Rel) | Weighted* Flow |
| 3/24/01 | 2007 | 570.1 | 0.28408 | 371.5 |
| 4/3/01 | 709 | 203.6 | 0.28714 | 547.7 |
| 4/13/01 | 1252 | 336.5 | 0.26874 | 427.6 |
| 4/23/01 | 1178 | 298.9 | 0.25377 | 329.5 |
| 5/3/01 | 2887 | 635.9 | 0.22026 | 478.4 |
| 5/13/01 | 2045 | 351.8 | 0.17204 | 234.2 |
| 5/23/01 | 2780 | 344.4 | 0.12389 | 228.0 |
| Overall | 12935 | 2,741.3 | 0.211925951 |  |
| Easton | Num.Rel. | Correlatio | etween Num | el., Flow |
|  | 12858 | Correlation | -0.42754 | Type 1P |
|  |  | DF | 6 | 0.2907 |
| Jack Creek |  |  |  |  |
| Ending <br> Release <br> Date | Number <br> Released <br> (Rel) | McNary Passage (McN) | Survival <br> Index (McN/Rel) | Weighted* Flow |
| 3/24/01 | 326 | 72.9 | 0.22354 | 247.7 |
| 36984 | 728 | 241.7 | 0.33198 | 441.3 |
| 4/13/01 | 716 | 209.7 | 0.29293 | 205.1 |
| 4/23/01 | 5101 | 1,393.8 | 0.27323 | 223.6 |
| 5/3/01 | 3064 | 690.6 | 0.22539 | 731.8 |
| 5/13/01 | 1626 | 221.1 | 0.13598 | 530.4 |
| 5/23/01 | 1285 | 135.2 | 0.10523 | 522.3 |
| Jack | Num.Rel. | Correlation between Num.Rel., Flow |  |  |
| Creek | 12846 | Correlation | 0.069292 | Type 1 P |
|  |  | DF | 6 | 0.8705 |
| Weighted* over | Num.Rel. 38559 | Correlation between Num. Rel., Flow |  |  |
| Sites |  | DF | 18 | - 0.5906 |

*Weight is daily number released applied to daily flow

Table 1. (continued)

## c. Year-2002 Outmigrants

| Clark Flats |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Ending <br> Release Date | Number <br> Released <br> (Rel) | McNary <br> Passage <br> (McN) | Survival Index (McN/Rel) | Weighted* <br> Flow |
| 3/24/02 | 917 | 288.1 | 0.31417 | 968.8 |
| 4/3/02 | 3931 | 1413.6 | 0.35959 | 1284.9 |
| 4/13/02 | 1761 | 499.4 | 0.28360 | 3633.7 |
| 4/23/02 | 4060 | 1834.1 | 0.45176 | 5085.7 |
| 5/3/02 | 899 | 134.9 | 0.15003 | 2405.3 |
| 5/13/02 | 336 | 64.0 | 0.19051 | 1793.0 |
| Clark <br> Flats | Num. Rel.$12114$ | =-_Correlation between Num. Rel., Flow |  |  |
|  |  | Correlation | 0.41301 | Type 1P |
|  |  | DF | 6 | 0.3092 |
| Easton |  |  |  |  |
| Ending | Number | McNary | Survival |  |
| Release | Released | Passage | Index | Weighted* |
| Date | (Rel) | (McN) | (McN/Rel) | Flow |
| 3/24/02 | 6995 | 1957.9 | 0.27991 | 294.3 |
| 4/3/02 | 913 | 132.5 | 0.14509 | 406.3 |
| 4/13/02 | 576 | 118.5 | 0.20567 | 783.5 |
| 4/23/02 | 2752 | 689.5 | 0.25055 | 1700.8 |
| 5/3/02 | 862 | 187.1 | 0.21703 | 424.6 |
| 5/13/02 | 214 | 38.5 | 0.17972 | 370.7 |
| 5/23/02 | 113 | 26.3 | 0.23243 | 455.0 |
| Easton | Num.Rel. <br> 12425 | Correlation between Num. Rel., Flow |  |  |
|  |  | Correlation DF | $\begin{gathered} 0.02540 \\ 6 \\ \hline \end{gathered}$ | $\begin{gathered} \text { Type } 1 \mathrm{P} \\ 0.9524 \end{gathered}$ |
| Jack Creek |  |  |  |  |
| Ending <br> Release <br> Date | Number <br> Released <br> (Rel) | McNary Passage (McN) | Survival Index (McN/Rel) | Weighted ${ }^{*}$ Flow |
| 3/24/02 | 3135 | 1132.7 | 0.36131 | 304.7 |
| 37349 | 1932 | 549.1 | 0.28423 | 600.8 |
| 4/13/02 | 2035 | 673.2 | 0.33080 | 1612.6 |
| 4/23/02 | 4180 | 1519.8 | 0.36359 | 2519.5 |
| 5/3/02 | 757 | 191.0 | 0.25229 | 1178.3 |
| 5/13/02 | 699 | 166.2 | 0.23780 | 708.8 |
| 5/23/02 | 203 | 56.6 | 0.27870 | 1033.2 |
| Jack | Num.Rel. | Correlation | etween Num | Rel., Flow |
| Creek | 12941 | Correlation | 0.439906 | Type 1 P |
|  |  | DF | 6 | 0.2754 |
| Weighted* over Sites | Num.Rel. 37480 | Correlation between Num.Rel., Flow |  |  |
|  |  | Correlation | 0.293798 | Type 1 P |
|  |  | DF | 18 | 0.2087 |

[^41]Table. 2. Analysis of Variation from Logistic Regression of Hatchery Spring Chinook Release-to-McNary Smolt-Survival Index on Number Released and Flow with Ten-Day Volitional Release Periods Serving as Database

## a. Year-2000 Outmigrants

|  | $\begin{array}{c}\text { Degrees of } \\ \text { Deviance } \\ \text { (Dev) }\end{array}$ |  |  |  | $\begin{array}{c}\text { Mean } \\ \text { (DF) }\end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | \(\left.\begin{array}{c}Deviance <br>

(Dev/DF)\end{array} \quad $$
\begin{array}{c}\text { F- } \\
\text { Ratio }\end{array}
$$ \quad $$
\begin{array}{c}\text { Type 1 } \\
\text { P }\end{array}
$$\right]\)
b. Year-2001 Outmigrants

| Source* | Deviance (Dev) | Degrees of Freedom (DF) | Mean Deviance (Dev/DF) | F- <br> Ratio | $\begin{gathered} \text { Type } 1 \\ \text { P } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Site | 56.77 | 2 | 28.39 | 0.49 | 0.6218 |
| Number** Site | 13.55 | 1 | 13.55 | 0.23 | 0.6353 |
| Flow | 29.29 | 1 | 29.29 | 0.50 | 0.4873 |
| Number \| Flow | 25.71 | 1 | 25.71 | 0.44 | 0.5148 |
| Flow \| Nunber | 41.45 | 1 | 41.45 | 0.71 | 0.4100 |
| Error | 987.62 | 17 | 58.10 |  |  |

c. Year-2002 Outmigrants

|  | Degrees of <br> Freedom <br> (DF $)$ |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Mean <br> Source* | Meviance <br> (Dev/DF) | F- <br> Ratio | Type 1 <br> P |  |  |
| Site | 325.92 | 2 | 162.96 | 8.64 | 0.0026 |
| Number** Site | 343.78 | 1 | 343.78 | 18.22 | 0.0005 |
| Flow | 111.75 | 1 | 111.75 | 5.92 | 0.0263 |
| Number \| Flow | 308.95 | 1 | 308.95 | 16.38 | 0.0008 |
| Flow \| Nunber | 76.92 | 1 | 76.92 | 4.08 | 0.0595 |
| Error | 320.69 | 17 | 18.86 |  |  |

* Within source, the symbol "|" means "adusted for"
** Number is volitional release number

Figure 1. Estimated Release-to-McNary Survival Estimates plotted against Volitional Release Number and Associated Survival-Index Logistic Fits adjusted for Site (number above estimate is ending Julian Date for Ten-Day Volitional Release Period; C.F. --Clark Flats, East-Easton and J.C.--Jack Creek).
a. Outmigration Year 2000

suvival index $=\frac{1}{1+\exp \left(a+0.0000252^{*}[\text { number released })\right]}$; wherein $\mathrm{a}=0.38024$ for Clark Flats, $\mathrm{a}=0.66891$ for Easton, $\mathrm{a}=0.50335$ Jack Creek

Figure 1. (continued)
b. Outmigration Year 2001

suvival index $=\frac{1}{1+\exp \left(a-0.0000328^{*}[\text { number released })\right]}$; wherein
$\mathrm{a}=1.16506$ for Clark Flats, $\mathrm{a}=1.37666$ for Easton, $\mathrm{a}=1.30892$ Jack Creek

Figure 1. (continued)
c. Outmigration Year 2001

suvival index $=\frac{1}{1+\exp \left(a-0.000175^{*}[\text { number released })\right]}$; wherein
$\mathrm{a}=0.96234$ for Clark Flats, $\mathrm{a}=1.65793$ for Easton, $\mathrm{a}=1.03487$ Jack Creek

## Appendix A. Estimation and Analysis Methods

## Estimation of Survival Index

The release-to-McNary smolt-to-smolt survival index in this study is estimated as follows:

wherein a stratum is a group of contiguous McNary detection dates among which the daily detection efficiencies ${ }^{9}$ are sufficiently homogeneous to permit the use of a pooled estimate of the detection efficiency for that stratum. The pooled estimate is a pooling of the daily detection efficiency estimates over all dates within the stratum.

The estimation of the stratum's McNary-detection efficiency is discussed in greater detail in Appendix B of my 2002 Annual Report: OCT-SNT Survival. However, it should be noted that the detection efficiency estimates for the 2000 and 2001 outmigrant used in this annual report on indirect predation are different than those used in the annual reports on OCT-SNT survival. This is because the currently used detection-efficiency estimates are based on all releases of PIT-tagged Spring Chinook into the Yakima River; whereas previous years' OCT-SNT overall survival-index estimates utilized detection efficiencies that were based on only OCT-SNT releases. Next year, overall survival index estimates for 2000 and 2001 outmigrants will be updated.

The strata's detection efficiencies used for estimating survival indices are presented in Appendix B along with the strata's McNary detections, expanded McNary detections, and the survival index estimates.

[^42]
## Logistic Regression Analysis

Weighted logistic regression of release-to-McNary survival-index estimates on predictor variables was performed using release number as the weighting variable instead of performing a more traditional least-squares regression. The reason for using logistic regression procedures are discussed in more detail in Appendix B of my 2002 Annual Report: OCT-SNT Survival. Suffice it to say here, the logit transform of the survival index is assumed to be linearly related to the predictor variable. The logit transform of the estimated survival index ( s ), which is a proportion, is:

$$
\operatorname{logit}(\mathrm{s})=\text { natural } \log \left(\frac{\mathrm{s}}{1-\mathrm{s}}\right)
$$

and the assumed underlying model is

$$
\operatorname{logit}(\mathrm{s})=\text { Site Intercept }+ \text { Coefficient } *(\text { Predictor Variable })
$$

One major benefit of this model is that the predicted value of the survival index,

$$
\mathrm{s}=\frac{1}{1+\mathrm{e}^{\operatorname{logit}}}
$$

can never be negative nor be greater than 1, which is appropriate for a predicted survival proportion. (In the above equation, e is the exponential constant.)

Another assumption behind the logistic regression procedures used in this study is that the variance of the survival index is proportional to what would be expected for binomially distributed proportion.

Variance (s) is proportion al to $\frac{S^{*}(1-S)}{n}$
wherein S is the expected ("true") proportion surviving and n is the number of fish released. The number released varied over releases; the variation in n would contribute to a variation in the variance of the survival proportion (as n increases in the above equation, the variance in $s$ decreases). To stabilize the variance in $s$ with respect to a change $n$, the release number was used as a weighting variable.

## Appendix B. Within Ten-Day Volitional Release Periods: McNary Detections, McNary Detection Efficiencies, and Expanded McNary Detections ${ }^{10}$; and Total Expanded McNary Detections, Release Numbers, and Estimated McNary Survival Indices ${ }^{11}$.

## 1. Outmigration-Year 2000, Clark Flats Volitional Releases

|  | Detection Efficiency Strata |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Beginning McNary Date > Ending McNary Date > | 4/14/00 | $\begin{aligned} & 4 / 15 / 00 \\ & 5 / 16 / 00 \end{aligned}$ | $\begin{aligned} & 5 / 17 / 00 \\ & 5 / 19 / 00 \end{aligned}$ | $\begin{aligned} & 5 / 20 / 00 \\ & 5 / 25 / 00 \end{aligned}$ | $\begin{aligned} & 5 / 26 / 00 \\ & 5 / 27 / 00 \end{aligned}$ | $\begin{aligned} & 5 / 28 / 00 \\ & 5 / 31 / 00 \end{aligned}$ | 6/1/00 | Total |  |  |
| Release Date | McNary Detections |  |  |  |  |  |  |  |  |  |
| 24-Mar-00 | 1 | 26 | 1 | 2 | 0 | 0 | 0 | 30 |  |  |
| 03-Apr-00 | 0 | 70 | 4 | 5 | 0 | 0 | 0 | 79 |  |  |
| 13-Apr-00 | 0 | 174 | 15 | 7 | 0 | 1 | 0 | 197 |  |  |
| 23-Apr-00 | 0 | 193 | 14 | 14 | 11 | 5 | 5 | 242 |  |  |
| 03-May-00 | 0 | 14 | 11 | 23 | 21 | 15 | 6 | 90 |  |  |
| 13-May-00 | 0 | 2 | 3 | 34 | 28 | 45 | 15 | 127 |  |  |
| 23-May-00 | 0 | 0 | 0 | 15 | 47 | 174 | 63 | 299 |  |  |
| 02-Jun-00 | 0 | 0 | 0 | 0 | 0 | 50 | 488 | 538 |  |  |
| 12-Jun-00 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |
|  | Dectection Efficiencies |  |  |  |  |  |  | Total Expanded Detections | Release Number | McNary Survival Index |
|  | 0.422458 | 0.312317 | 0.273358 | 0.2233586 | 0.264487 | 0.337151 | 0.255567 |  |  |  |
| Release Date | Expanded McNary Detections |  |  |  |  |  |  |  |  |  |
| 24-Mar-00 | 2.37 | 83.25 | 3.66 | 8.95 | 0.00 | 0.00 | 0.00 | 98.23 | 327 | 0.300393 |
| 03-Apr-00 | 0.00 | 224.13 | 14.63 | 22.39 | 0.00 | 0.00 | 0.00 | 261.15 | 572 | 0.456556 |
| 13-Apr-00 | 0.00 | 557.13 | 54.87 | 31.34 | 0.00 | 2.97 | 0.00 | 646.31 | 1780 | 0.363093 |
| 23-Apr-00 | 0.00 | 617.96 | 51.21 | 62.68 | 41.59 | 14.83 | 19.56 | 807.84 | 2458 | 0.328658 |
| 03-May-00 | 0.00 | 44.83 | 40.24 | 102.97 | 79.40 | 44.49 | 23.48 | 335.41 | 1063 | 0.315528 |
| 13-May-00 | 0.00 | 6.40 | 10.97 | 152.22 | 105.87 | 133.47 | 58.69 | 467.63 | 1114 | 0.419775 |
| 23-May-00 | 0.00 | 0.00 | 0.00 | 67.16 | 177.70 | 516.09 | 246.51 | 1007.46 | 1840 | 0.547532 |
| 02-Jun-00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 148.30 | 1909.48 | 2057.78 | 4926 | 0.417738 |
| 12-Jun-00 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 263 | 0 |

## 2. Outmigration-Year 2000, Easton Volitional Releases



[^43]${ }^{11}$ Survival Index $=($ Expanded McNary Detections $) /($ Release Number $)$

# Appendix B. Within Ten-Day Volitional Release Periods: McNary Detections, McNary Detection Efficiencies, and Expanded McNary Detections; and Total Expanded McNary Detections, Release Numbers, and Estimated McNary Survival Index (Continued). 

3. Outmigration-Year 2000, Jack Creek Volitional Releases

|  | Detection Efficiency Strata |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Beginning McNary Date > | 1/0/00 | 4/15/00 | 5/17/00 | 5/20/00 | 5/26/00 | 5/28/00 | 6/1/00 |  |  |  |
| Ending McNary Date > | 4/14/00 | 5/16/00 | 5/19/00 | 5/25/00 | 5/27/00 | 5/31/00 | 7/19/2000 | Total |  |  |
| Release Date | McNary Detections |  |  |  |  |  |  |  |  |  |
| 24-Mar-00 | 1 | 25 | 0 | 3 | 0 | 0 | 0 | 29 |  |  |
| 03-Apr-00 | 0 | 56 | 6 | 13 | 2 | 2 | 3 | 82 |  |  |
| 13-Apr-00 | 0 | 47 | 12 | 9 | 3 | 1 | 0 | 72 |  |  |
| 23-Apr-00 | 0 | 21 | 7 | 21 | 13 | 31 | 8 | 101 |  |  |
| 03-May-00 | 0 | 0 | 4 | 12 | 8 | 13 | 7 | 44 |  |  |
| 13-May-00 | 0 | 0 | 0 | 1 | 4 | 15 | 8 | 28 |  |  |
| 23-May-00 | 0 | 0 | 0 | 0 | 6 | 65 | 171 | 242 |  |  |
| 02-Jun-00 | 0 | 0 | 0 | 0 | 0 | 0 | 214 | 214 |  |  |
| 12-Jun-00 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |
| Release Date | Dectection Efficiencies |  |  |  |  |  |  | Total Expanded Detections | Release Number | McNary Survival Index |
|  | 0.422458 | 0.312317 | 0.273358 | 0.2233586 | 0.264487 | 0.337151 | 0.255567 |  |  |  |
|  | Expanded McNary Detections |  |  |  |  |  |  |  |  |  |
| 24-Mar-00 | 2.37 | 80.05 | 0.00 | 13.43 | 0.00 | 0.00 | 0.00 | 95.85 | 340 |  |
| 03-Apr-00 | 0.00 | 179.31 | 21.95 | 58.20 | 7.56 | 5.93 | 11.74 | 284.69 | 1078 | $\begin{aligned} & 0.281898 \\ & 0.264090 \end{aligned}$ |
| 13-Apr-00 | 0.00 | 150.49 | 43.90 | 40.29 | 11.34 | 2.97 | 0.00 | 248.99 | 810 | 0.307395 |
| 23-Apr-00 | $\begin{aligned} & 0.00 \\ & 0.00 \end{aligned}$ | 67.24 | 25.61 | 94.02 | 49.15 | 91.95 | 31.30 | 359.27 | 1055 | 0.340538 |
| 03-May-00 |  | 0.00 | 14.63 | 53.73 | 30.25 | 38.56 | 27.39 | 164.55 | 447 | 0.368129 |
| 13-May-00 | 0.00 | 0.00 | 0.00 | 4.48 | 15.12 | 44.49 | 31.30 | 95.39 | 190 | 0.502074 |
| 23-May-00 | 0.00 | 0.00 | 0.00 | 0.00 | 22.69 | 192.79 | 669.10 | 884.58 | 1847 | 0.478926 |
| 02-Jun-00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 837.35 | 837.35 | 2412 | 0.347161 |
| 12-Jun-00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2 | 0.000000 |

4. Outmigration-Year 2001, Clark Flats Volitional Releases

| Beginning McNary Date > Ending McNary Date > | Detection Efficiency Strata |  |  |  |  | Total Detections |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 4/1/2001 | 5/6/2001 | 5/24/2001 | 5/29/2001 | 6/2/2001 |  |  |  |
|  | 5/5/2001 | 5/23/2001 | 5/28/2001 | 6/1/2001 | 7/31/2001 |  |  |  |
| Release Date | McNary Detections |  |  |  |  |  |  |  |
| 24-Mar-01 | 36 | 199 | 10 | 0 | 0 | 245 |  |  |
| 03-Apr-01 | 31 | 201 | 13 | 3 | 2 | 250 |  |  |
| 13-Apr-01 | 23 | 270 | 18 | 8 | 0 | 319 |  |  |
| 23-Apr-01 | 19 | 393 | 19 | 12 | 4 | 447 |  |  |
| 03-May-01 | 2 | 594 | 202 | 88 | 25 | 911 |  |  |
| 13-May-01 | 0 | 23 | 76 | 43 | 20 | 162 |  |  |
| 23-May-01 | 0 | 0 | 12 | 133 | 49 | 194 |  |  |
| 02-Jun-01 | 0 | 0 | 0 | 0 | 2 | 2 |  |  |
| Release Date | Dectection Efficiencies |  |  |  |  | Total Expanded Detections |  | McNary |
|  | 0.770785 | 0.786313 | 0.766575 | 0.748928 | 0.771139 |  | Release | Survival |
|  | Expanded McNary Detections |  |  |  |  |  | Number | Index |
| 24-Mar-01 | 46.71 | 253.08 | 13.05 | 0.00 | 0.00 | 312.83 | 1043 | 0.299933 |
| 03-Apr-01 | 40.22 | 255.62 | 16.96 | 4.01 | 2.59 | 319.40 | 1063 | 0.300470 |
| 13-Apr-01 | 29.84 | 343.37 | 23.48 | 10.68 | 0.00 | 407.38 | 1164 | 0.349980 |
| 23-Apr-01 | 24.65 | 499.80 | 24.79 | 16.02 | 5.19 | 570.45 | 1827 | 0.312231 |
| 03-May-01 | 2.59 | 755.42 | 263.51 | 117.50 | 32.42 | 1171.45 | 4361 | 0.268619 |
| 13-May-01 | 0.00 | 29.25 | 99.14 | 57.42 | 25.94 | 211.74 | 1583 | 0.133761 |
| 23-May-01 | 0.00 | 0.00 | 15.65 | 177.59 | 63.54 | 256.78 | 1814 | 0.141557 |
| 02-Jun-01 | 0.00 | 0.00 | 0.00 | 0.00 | 2.59 | 2.59 | 116 | 0.022358 |

## Appendix B. Within Ten-Day Volitional Release Periods: McNary Detections, McNary Detection Efficiencies, and Expanded McNary Detections; and Total Expanded McNary Detections, Release Numbers, and Estimated McNary Survival Index (Continued).

5. Outmigration-Year 2001, Easton Volitional Releases

|  | Detection Efficiency Strata |  |  |  |  | Total Detections |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Beginning McNary Date > Ending McNary Date > | $\begin{aligned} & \text { 4/1/2001 } \\ & 5 / 5 / 2001 \end{aligned}$ | $\begin{gathered} 5 / 6 / 2001 \\ 5 / 23 / 2001 \end{gathered}$ | $\begin{aligned} & 5 / 24 / 2001 \\ & 5 / 28 / 2001 \end{aligned}$ | $\begin{gathered} \text { 5/29/2001 } \\ 6 / 1 / 2001 \end{gathered}$ | $\begin{gathered} 6 / 2 / 2001 \\ 7 / 31 / 2001 \end{gathered}$ |  |  |  |
| Release Date | McNary Detections |  |  |  |  |  |  |  |
| 24-Mar-01 | 46 | 326 | 32 | 25 | 16 | 445 |  |  |
| 03-Apr-01 | 13 | 119 | 17 | 6 | 4 | 159 |  |  |
| 13-Apr-01 | 6 | 184 | 35 | 27 | 10 | 262 |  |  |
| 23-Apr-01 | 1 | 144 | 35 | 37 | 15 | 232 |  |  |
| 03-May-01 | 1 | 169 | 144 | 127 | 48 | 489 |  |  |
| 13-May-01 | 0 | 8 | 52 | 140 | 67 | 267 |  |  |
| 23-May-01 | 0 | 0 | 3 | 154 | 104 | 261 |  |  |
| 02-Jun-01 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |
| Release Date | Dectection Efficiencies |  |  |  |  | Total |  | McNary |
|  | 0.770785 | 0.786313 | 0.766575 | 0.748928 | 0.771139 | Expanded | Release | Survival |
|  | Expanded McNary Detections |  |  |  |  | Detections | Number | Index |
| 24-Mar-01 | 59.68 | 414.59 | 41.74 | 33.38 | 20.75 | 570.15 | 2007 | 0.284079 |
| 03-Apr-01 | 16.87 | 151.34 | 22.18 | 8.01 | 5.19 | 203.58 | 709 | 0.287137 |
| 13-Apr-01 | 7.78 | 234.00 | 45.66 | 36.05 | 12.97 | 336.46 | 1252 | 0.268742 |
| 23-Apr-01 | 1.30 | 183.13 | 45.66 | 49.40 | 19.45 | 298.94 | 1178 | 0.253772 |
| 03-May-01 | 1.30 | 214.93 | 187.85 | 169.58 | 62.25 | 635.89 | 2887 | 0.220261 |
| 13-May-01 | 0.00 | 10.17 | 67.83 | 186.93 | 86.88 | 351.83 | 2045 | 0.172042 |
| 23-May-01 | 0.00 | 0.00 | 3.91 | 205.63 | 134.87 | 344.41 | 2780 | 0.123887 |
| 02-Jun-01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 76 | 0.000000 |

6. Outmigration-Year 2001, Jack Creek Volitional Releases

|  | Detection Efficiency Strata |  |  |  |  | Total Detections |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Beginning McNary Date > Ending McNary Date > | $\begin{aligned} & \text { 4/1/2001 } \\ & 5 / 5 / 2001 \end{aligned}$ | $\begin{gathered} 5 / 6 / 2001 \\ 5 / 23 / 2001 \end{gathered}$ | $\begin{aligned} & 5 / 24 / 2001 \\ & 5 / 28 / 2001 \end{aligned}$ | $\begin{gathered} \text { 5/29/2001 } \\ 6 / 1 / 2001 \end{gathered}$ | $\begin{gathered} \text { 6/2/2001 } \\ 7 / 31 / 2001 \end{gathered}$ |  |  |  |
| Release Date | McNary Detections |  |  |  |  |  |  |  |
| 24-Mar-01 | 10 | 44 | 2 | 1 | 0 | 57 |  |  |
| 03-Apr-01 | 17 | 150 | 13 | 6 | 3 | 189 |  |  |
| 13-Apr-01 | 11 | 129 | 17 | 4 | 3 | 164 |  |  |
| 23-Apr-01 | 29 | 785 | 149 | 102 | 21 | 1086 |  |  |
| 03-May-01 | 0 | 189 | 155 | 144 | 43 | 531 |  |  |
| 13-May-01 | 0 | 5 | 18 | 84 | 61 | 168 |  |  |
| 23-May-01 | 0 | 0 | 0 | 43 | 60 | 103 |  |  |
| 02-Jun-01 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |
|  | Dectection Efficiencies |  |  |  |  | Total |  | McNary |
|  | 0.770785 | 0.786313 | 0.766575 | 0.7489276 | 0.771139 | Expanded | Release | Survival |
| Release Date | Expanded McNary Detections |  |  |  |  | Detections | Number | Index |
| 24-Mar-01 | 12.97 | 55.96 | 2.61 | 1.34 | 0.00 | 72.88 | 326 | 0.223544 |
| 03-Apr-01 | 22.06 | 190.76 | 16.96 | 8.01 | 3.89 | 241.68 | 728 | 0.331977 |
| 13-Apr-01 | 14.27 | 164.06 | 22.18 | 5.34 | 3.89 | 209.74 | 716 | 0.292927 |
| 23-Apr-01 | 37.62 | 998.33 | 194.37 | 136.19 | 27.23 | 1393.75 | 5101 | 0.273231 |
| 03-May-01 | 0.00 | 240.36 | 202.20 | 192.27 | 55.76 | 690.60 | 3064 | 0.225391 |
| 13-May-01 | 0.00 | 6.36 | 23.48 | 112.16 | 79.10 | 221.10 | 1626 | 0.135980 |
| 23-May-01 | 0.00 | 0.00 | 0.00 | 57.42 | 77.81 | 135.22 | 1285 | 0.105231 |
| 02-Jun-01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0 | \#DIV/0! |

## Appendix B. Within Ten-Day Volitional Release Periods: McNary Detections, McNary Detection Efficiencies, and Expanded McNary Detections; and Total Expanded McNary Detections, Release Numbers, and Estimated McNary Survival Index (Continued).

7. Outmigration-Year 2002, Clark Flats Volitional Releases

| Beginning McNary Date > Ending McNary Date > | Detection Efficiency Strata |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 3/1/2002 | 4/22/2002 | 4/26/2002 | 4/30/2002 | 5/2/2002 | 5/3/2002 | 5/6/2002 | 5/7/2002 |
|  | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 |
| Release Date | McNary Detections |  |  |  |  |  |  |  |
| 24-Mar-02 | 29 | 19 | 11 | 8 | 11 | 28 | 4 | 8 |
| 03-Apr-02 | 67 | 69 | 70 | 55 | 56 | 203 | 47 | 47 |
| 13-Apr-02 | 5 | 6 | 12 | 23 | 25 | 102 | 37 | 31 |
| 23-Apr-02 | 2 | 7 | 31 | 39 | 59 | 388 | 119 | 166 |
| 03-May-02 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 9 |
| 13-May-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 23-May-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 02-Jun-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Dectection Efficiencies |  |  |  |  |  |  |  |
|  | 0.332561 | 0.388115 | 0.417032 | 0.458902 | 0.513120 | 0.579215 | 0.542672 | 0.495823 |
| Release Date | Expanded McNary Detections |  |  |  |  |  |  |  |
| 24-Mar-02 | 87.20 | 48.95 | 26.38 | 17.43 | 21.44 | 48.34 | 7.37 | 16.13 |
| 03-Apr-02 | 201.47 | 177.78 | 167.85 | 119.85 | 109.14 | 350.47 | 86.61 | 94.79 |
| 13-Apr-02 | 15.03 | 15.46 | 28.77 | 50.12 | 48.72 | 176.10 | 68.18 | 62.52 |
| 23-Apr-02 | 6.01 | 18.04 | 74.33 | 84.99 | 114.98 | 669.87 | 219.29 | 334.80 |
| 03-May-02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 3.69 | 18.15 |
| 13-May-02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 23-May-02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 02-Jun-02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |



## Appendix B. Within Ten-Day Volitional Release Periods: McNary Detections, McNary Detection Efficiencies, and Expanded McNary Detections; and Total Expanded McNary Detections, Release Numbers, and Estimated McNary Survival Index (Continued).

8. Outmigration-Year 2002, Easton Volitional Releases

| Beginning McNary Date > Ending McNary Date > | Detection Efficiency Strata |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 3/1/2002 | 4/22/2002 | 4/26/2002 | 4/30/2002 | 5/2/2002 | 5/3/2002 | 5/6/2002 | 5/7/2002 |
|  | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 |
| Release Date | McNary Detections |  |  |  |  |  |  |  |
| 24-Mar-02 | 29 | 19 | 11 | 8 | 11 | 28 | 4 | 8 |
| 03-Apr-02 | 67 | 69 | 70 | 55 | 56 | 203 | 47 | 47 |
| 13-Apr-02 | 5 | 6 | 12 | 23 | 25 | 102 | 37 | 31 |
| 23-Apr-02 | 2 | 7 | 31 | 39 | 59 | 388 | 119 | 166 |
| 03-May-02 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 9 |
| 13-May-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 23-May-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 02-Jun-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 12-Jun-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Dectection Efficiencies |  |  |  |  |  |  |  |
|  | 0.332561 | 0.388115 | 0.417032 | 0.458902 | 0.513120 | 0.579215 | 0.542672 | 0.495823 |
| Release Date | Expanded McNary Detections |  |  |  |  |  |  |  |
| 24-Mar-02 | 240.56 | 193.241693 | 158.261179 | 128.567679 | 159.806706 | 485.139358 | 178.745089 | 175.46572 |
| 03-Apr-02 | 0 | 7.72966771 | 9.59158658 | 8.71645281 | 7.79544907 | 31.0765425 | 9.2136644 | 20.1684736 |
| 13-Apr-02 | 0 | 0 | 9.59158658 | 8.71645281 | 5.8465868 | 29.3500679 | 9.2136644 | 14.1179315 |
| 23-Apr-02 | 0 | 0 | 0 | 15.2537924 | 5.8465868 | 79.4178308 | 73.7093152 | 102.859215 |
| 03-May-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2.01684736 |
| 13-May-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 23-May-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 02-Jun-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 12-Jun-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


| Beginning McNary Date > Ending McNary Date > | Detection Efficiency Strata |  |  |  |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{array}{r} 5 / 9 / 2002 \\ 5 / 14 / 2002 \end{array}$ | $\begin{aligned} & 5 / 15 / 2002 \\ & 5 / 24 / 2002 \end{aligned}$ | $\begin{aligned} & 5 / 25 / 2002 \\ & 5 / 27 / 2002 \end{aligned}$ | $\begin{aligned} & 5 / 28 / 2002 \\ & 5 / 30 / 2002 \end{aligned}$ | $\begin{aligned} & 5 / 31 / 2002 \\ & 7 / 31 / 2002 \end{aligned}$ |  |  |  |
| Release Date |  |  |  |  |  |  |  |  |
| 24-Mar-02 | 2 | 4 | 0 | 0 | 0 | 124 |  |  |
| 03-Apr-02 | 21 | 21 | 0 | 1 | 0 | 657 |  |  |
| 13-Apr-02 | 13 | 2 | 0 | 0 | 0 | 256 |  |  |
| 23-Apr-02 | 73 | 53 | 1 | 2 | 0 | 940 |  |  |
| 03-May-02 | 14 | 30 | 0 | 1 | 0 | 56 |  |  |
| 13-May-02 | 0 | 8 | 10 | 3 | 2 | 23 |  |  |
| 23-May-02 | 0 | 3 | 9 | 6 | 1 | 19 |  |  |
| 02-Jun-02 | 0 | 0 | 0 | 2 | 3 | 5 |  |  |
| 12-Jun-02 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |
| Release Date | Dectection Efficiencies |  |  |  |  | Total Expanded Detections | Release | McNary |
|  | 0.443119 | 0.387133 | 0.441514 | 0.252866 | 0.226415 |  |  | Survival |
|  |  |  |  |  |  |  | Number | Index |
| 24-Mar-02 | 94.7827229 | 134.320596 | 9.05974026 | 0 | 0 | 1957.95 | 6995 | 0.27990678 |
| 03-Apr-02 | 9.02692599 | 20.664707 | 4.52987013 | 3.95466667 | 0 | 132.468007 | 913 | 0.14509092 |
| 13-Apr-02 | 15.7971205 | 25.8308838 | 0 | 0 | 0 | 118.464294 | 576 | 0.20566718 |
| 23-Apr-02 | 108.323112 | 247.976485 | 22.6493506 | 15.8186667 | 17.6666667 | 689.521021 | 2752 | 0.2505527 |
| 03-May-02 | 15.7971205 | 95.5742701 | 18.1194805 | 15.8186667 | 39.75 | 187.076385 | 862 | 0.21702597 |
| 13-May-02 | 0 | 5.16617676 | 6.79480519 | 0 | 26.5 | 38.460982 | 214 | 0.17972421 |
| 23-May-02 | 0 | 0 | 9.05974026 | 3.95466667 | 13.25 | 26.2644069 | 113 | 0.23242838 |
| 02-Jun-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \#DIV/0! |
| 12-Jun-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \#DIV/0! |

## Appendix B. Within Ten-Day Volitional Release Periods: McNary Detections, McNary Detection Efficiencies, and Expanded McNary Detections; and Total Expanded McNary Detections, Release Numbers, and Estimated McNary Survival Index (Continued).

## 9. Outmigration-Year 2002, Jack Creek Volitional Releases

| Beginning McNary Date > Ending McNary Date > | Detection Efficiency Strata |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 3/1/2002 | 4/22/2002 | 4/26/2002 | 4/30/2002 | 5/2/2002 | 5/3/2002 | 5/6/2002 | 5/7/2002 |
|  | 4/21/2002 | 4/25/2002 | 4/29/2002 | 5/1/2002 | 5/2/2002 | 5/5/2002 | 5/6/2002 | 5/8/2002 |
| Release Date | McNary Detections |  |  |  |  |  |  |  |
| 24-Mar-02 | 29 | 19 | 11 | 8 | 11 | 28 | 4 | 8 |
| 03-Apr-02 | 67 | 69 | 70 | 55 | 56 | 203 | 47 | 47 |
| 13-Apr-02 | 5 | 6 | 12 | 23 | 25 | 102 | 37 | 31 |
| 23-Apr-02 | 2 | 7 | 31 | 39 | 59 | 388 | 119 | 166 |
| 03-May-02 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 9 |
| 13-May-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 23-May-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 02-Jun-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Dectection Efficiencies |  |  |  |  |  |  |  |
|  | 0.332561 | 0.388115 | 0.417032 | 0.458902 | 0.513120 | 0.579215 | 0.542672 | 0.495823 |
| Release Date | Expanded McNary Detections |  |  |  |  |  |  |  |
| 24-Mar-02 | 174.40 | 126.2512393 | 119.8948323 | 58.8360565 | 77.9544907 | 271.0565096 | 88.45117828 | 80.6738944 |
| 24-Mar-02 | 42.09755628 | 30.91867084 | 52.7537262 | 43.58226407 | 35.07952082 | 134.6650175 | 44.22558914 | 76.64019968 |
| 03-Apr-02 | 15.03484153 | 12.88277952 | 23.97896646 | 32.68669806 | 40.92610762 | 193.3651533 | 70.02384947 | 102.8592154 |
| 13-Apr-02 | 0 | 2.576555903 | 21.58106981 | 32.68669806 | 48.72155669 | 277.9624079 | 184.2732881 | 288.4091725 |
| 23-Apr-02 | 0 | 0 | 0 | 0 | 0 | 0 | 1.842732881 | 6.05054208 |
| 03-May-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13-May-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 23-May-02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |



## APPENDIX G

Financial Reports


| Yakama Nation-Fisheries Program |  |  |  |  |  | Page 2 of 9 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Project No. |  | 1995-63-25 |  |  |  |  |  |  |
| Project Name: |  | Monitoring \& Evaluation |  |  |  |  |  |  |
| Contract No. : |  | 5881 |  |  |  |  |  |  |
| Period Covered: |  | 4/1/02-4/30/03 |  |  |  |  |  |  |
| Prepared by: |  | Ida Sohappy-lke |  |  |  |  |  |  |
| Contact Person: |  | Ida Sohappy-Ike @ 509-865-6262 x6630 |  |  |  |  |  |  |
| 8458101 M\&E Cont' |  |  |  |  |  |  |  |  |
| Code |  | Approved Budget |  | Prior Claimed |  | Claimed this Invoice |  | Balance |
| 551295 | Vehicle Rental | \$ 7,035.00 |  |  |  |  |  |  |
|  | 100-Modeling |  | \$ 1,238.71 | \$ 763.32 | \$ | 475.39 |  |  |
|  | 107-Chandler |  | \$ 1,358.05 | \$ 1,358.05 | \$ | - |  |  |
|  | 115-Roza |  | \$ | \$ 170.50 | \$ | (170.50) |  |  |
|  | 400-Avian Predation |  | \$ 2,734.56 | \$ 1,556.29 | \$ | 1,178.27 |  |  |
|  | 401-Predation |  | \$ | \$ | \$ | - |  |  |
|  | Sub-Total |  | \$ 5,331.32 | \$ 3,848.16 | \$ | 1,483.16 | \$ | 1,703.68 |
| 561171 | Telephone | \$ 2,424.00 |  |  |  |  |  |  |
|  | 100-Modeling |  | \$ 1,014.17 | \$ 882.29 | \$ | 131.88 |  |  |
|  | 107-Chandler |  | \$ 324.13 | \$ 324.13 | \$ | - |  |  |
|  | 401-Predation |  | \$ 251.58 | \$ 251.58 | \$ | - |  |  |
|  | Sub-Total |  | \$ 1,589.88 | \$ 1,458.00 | \$ | 131.88 | \$ | 834.12 |
| 571111 | Insurance | \$ 1,465.00 |  |  |  |  |  |  |
|  | 100-Modeling |  | \$ 390.78 | \$ | \$ | 390.78 |  |  |
|  | 107-Chandler |  | \$ 73.74 | \$ 73.74 | \$ | - |  |  |
|  | 400-Avian Predation |  | \$ 570.84 | \$ | \$ | 570.84 |  |  |
|  | 401-Predation |  | \$ 424.00 | \$ | \$ | 424.00 |  |  |
|  | Sub-Total |  | \$ 1,459.36 | \$ 73.74 | \$ | 1,385.62 | \$ | 5.64 |
| 581110 | Travel Holding | \$ |  |  |  |  |  |  |
|  | 100-Modeling |  | \$ | \$ 375.18 | \$ | (375.18) |  |  |
|  | 401-Predation |  | \$ | \$ | \$ | - |  |  |
|  | Sub-Total |  | \$ | \$ 375.18 | \$ | (375.18) | \$ | - |
| 581121 | Per Diem | \$ 2,500.00 |  |  |  |  |  |  |
|  | 100-Modeling |  | \$ 3,039.98 | \$ 2,254.38 | \$ | 785.60 |  |  |
|  | 401-Predation |  | \$ | \$ 173.04 | \$ | (173.04) |  |  |
|  | Sub-Total |  | \$ 3,039.98 | \$ 2,427.42 | \$ | 612.56 | \$ | (539.98) |
| 621251 | Indirect Cost | \$ 63,031.00 |  |  |  |  |  |  |
|  | 100-Modeling |  | \$ 14,655.65 | \$ 12,778.08 | \$ | 1,877.57 |  |  |
|  | 105-Juv. Spring Chinook Marking |  | \$ 13,217.69 | \$ 13,217.69 | \$ | - |  |  |
|  | 107-Chandler |  | \$ 6,802.89 | \$ 6,203.15 | \$ | 599.74 |  |  |
|  | 110-Behavior |  | \$ | \$ | \$ | - |  |  |
|  | 115-Roza |  | \$ | \$ 33.25 | \$ | (33.25) |  |  |
|  | 400-Avian Predation |  | \$ 9,922.65 | \$ 8,382.50 | \$ | 1,540.15 |  |  |
|  | 401-Predation |  | \$ 2,138.98 | \$ 991.03 | \$ | 1,147.95 |  |  |
|  | 403-Indirect Predation |  | \$ 3,913.38 | \$ 3,519.24 | \$ | 394.14 |  |  |
|  | 124-Aerial Flights |  | \$ 307.63 | \$ 307.63 | \$ | - |  |  |
|  | 131-Predator Avoidance Training |  | \$ 1,572.21 | \$ 1,536.23 | \$ | 35.98 |  |  |
|  | No Poject |  | \$ (735.39) | \$ | \$ | (735.39) |  |  |
|  | Sub-Total |  | \$ 51,795.69 | \$ 46,968.80 | \$ | 4,826.89 | \$ | 11,235.31 |
| 522251 | Sub-Contracts | \$ 69,200.00 |  |  |  |  |  |  |
|  | 100-Modeling |  | \$ 4,200.00 | \$ 4,200.00 | \$ | - |  |  |
|  | 403-Indirect Predation |  | \$ 10,933.65 | \$ | \$ | 10,933.65 |  |  |
|  | 133-Biometrical Support |  | \$ 42,084.98 | \$ 42,084.98 | \$ | - |  |  |
|  | Sub-Total |  | \$ 57,218.63 | \$ 46,284.98 | \$ | 10,933.65 | \$ | 11,981.37 |
| 651171 | Capital Equipment | \$ 129,600.00 |  |  |  |  |  |  |
|  | 105-Juv. Spring Chinook Marking |  | \$ 129,600.00 | \$ 129,600.00 | \$ | - |  |  |
|  | Sub-Total |  | \$ 129,600.00 | \$ 129,600.00 | \$ | - | \$ | - |
|  | TOTAL: | \$ 588,293.00 | \$ 504,233.24 | \$ 464,271.64 | \$ | 39,961.60 | \$ | 78,178.76 |






## Yakama Nation-Fisheries Program

Project No.
Project Name:
Contract No. :
Period Covered:
Prepared by:

## Contact Person:

1995-63-25
Monitoring \& Evaluation
5881
4/1/02-4/30/03
Ida Sohappy-Ike
Ida Sohappy-Ike @ 509-865-6262 x6630

| 8458106 | Fall Chinook |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Code | Description | Approved Budget |  | YTD Exp |  | Prior Claimed |  | Claimed this Invoice |  | Balance |  |
| 512111 | WAGES | \$ | 47,760.00 |  |  |  |  |  |  |  |  |
|  | 108-Fall Chinook |  |  | \$ | 52,150.44 | \$ | 49,670.18 | \$ | 2,480.26 | \$ | (4,390.44) |
| 519111 | FRINGE | \$ | 9,131.00 |  |  |  |  |  |  |  |  |
|  | 108-Fall Chinook |  |  | \$ | 7,777.42 | \$ | 7,225.42 | \$ | 552.00 | \$ | 1,353.58 |
| 541122 | Sensitive Equipment | \$ | 17,367.00 |  |  |  |  |  |  |  |  |
|  | 108-Fall Chinook |  |  | \$ | 16,691.79 | \$ | 15,055.79 | \$ | 1,636.00 | \$ | 675.21 |
| 551111 | Operating Supplies | \$ | 8,478.00 |  |  |  |  |  |  |  |  |
|  | 108-Fall Chinook |  |  | \$ | 4,997.51 | \$ | 4,343.00 | \$ | 654.51 | \$ | 3,480.49 |
| 551295 | Vehicle Rental | \$ | 8,043.00 |  |  |  |  |  |  |  |  |
|  | 108-Fall Chinook |  |  | \$ | 6,189.87 | \$ | 5,518.61 | \$ | 671.26 | \$ | 1,853.13 |
| 561171 | Telephone | \$ | 960.00 |  |  |  |  |  |  |  |  |
|  | 108-Fall Chinook |  |  | \$ | 673.87 | \$ | 590.53 | \$ | 83.34 | \$ | 286.13 |
| 571111 | Insurance | \$ | 867.00 |  |  |  |  |  |  |  |  |
|  | 108-Fall Chinook |  |  | \$ | 101.19 | \$ | - | \$ | 101.19 | \$ | 765.81 |
| 581110 | Travel Holding | \$ | - |  |  |  |  |  |  |  |  |
|  | 108-Fall Chinook |  |  | \$ | - | \$ | - | \$ | - | \$ | - |
| 581121 | Per Diem | \$ | 2,800.00 |  |  |  |  |  |  |  |  |
|  | 108-Fall Chinook |  |  | \$ | 2,612.38 | \$ | 2,612.38 | \$ | - | \$ | 187.62 |
| 621251 | Indirect Cost | \$ | 18,605.00 |  |  |  |  |  |  |  |  |
|  | 108-Fall Chinook |  |  | \$ | 17,763.19 | \$ | 16,463.87 | \$ | 1,299.32 | \$ | 841.81 |
| TOTAL: |  | \$ | 114,011.00 |  | 108,957.66 | \$ | 101,479.78 | \$ | 7,477.88 | \$ | 5,053.34 |



I CERTIFY THIS IS TRUE AND ACCURATE

Chairman
Yakama Tribal Council


## APPENDIX H

Equipment Inventory List

ORIGINAL SHEET BEFORE 5/6/03 UPDATES


| CAPTTALIZED EQUIPment list for bpa | Contract | PROJECT | S.T. 4/22/03 |
| :---: | :---: | :---: | :---: |
| YKFP Monitoring \& Evaluation | Number | NUMBER |  |


Condition: 4. Good 5. Fair 6. Poor 7. Salvageale 8. MissingStolen - HIS SHEET UPDATED AS OF 5/6/03 BY S.T

| No. $\begin{gathered}\text { ITEM } \\ \text { DESCRIPTION }\end{gathered}$ | VENDOR | MAKE | SERIAL NUMBER | YEAR | FUND NUMBER | PROPERTY NUMBER | NOC. | $\begin{aligned} & \text { ITEM } \\ & \text { COST } \end{aligned}$ | location CODE | COMMENTS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 AGIILE MODULATORIDEMODULATOR | C \& C COMMUNICATIONS | M ${ }^{\text {369CADO,375CADO,S/H }}$ |  | 1999 | 8458109.1 |  | 66507 | 3,763.00 |  |  |
| 2 LIBRARY SOFTWARE | VISION 1 |  |  | 1999 | 8458109.1 |  | 67161 | 895.00 |  |  |
| 3 VIIEOO MONITORING EQUIP | VIISION 1 | PCWPIII DUAL PROC, S/H |  | 1999 | 8458109.1 |  | ${ }^{73588}$ | 15.515.00 |  |  |
| ${ }^{3} 4$ POWER SUPPLY | MECCOINC. | PS120-12 | PN10000226 | 1999 | 8458109.1 |  | 67167 | 95.00 |  |  |
| 5 2FISH CULTURE TANKS | AQUA CENTER |  |  | 1999 | 8458109.2 |  | ${ }^{72942}$ | 8,497.00 | ${ }_{5 / 4}$ |  |
| 6 LEFT RTN DESK W/ORGANIZER | INSIDE OREGON |  |  | 1999 | 8458109.2 |  | 72941 | 1,665.00 | $1 / 4$ |  |
| 7 PENTIUM PROCESSOR | GATEWAY COMPANIES |  |  | 1999 | 8458109.3 |  | 62545 | 2,881.48 | 5/4 |  |
| 8 HP COLOR LASER JET 4500N | PACIIC F FIRST COMPUTERS |  |  | 1999 | 8458109.7 |  | 56115 | 3,100.00 |  |  |
| 2WANDSTYLE TAGDETECTOR | NORTHWEST MARIIEE TEC | 1078810790 |  | 1999 | 8458109.7 |  | ${ }^{74731}$ | 13,000.00 |  |  |
| 10 3/PENTIUM MIICOMPUTERS | GATEWAYCOMPANIES | LP MINI TWR TB3 GP7 | 18596571 | ${ }^{1999}$ | 8458109.7 |  | ${ }^{77732}$ | 5,535.00 |  |  |
| 11. HP JET PRINTERDIGITITALVIDEOCREAT | GATEWAYCOMPANIES |  |  | 1999 | 8458109.7 |  | ${ }^{74732}$ | 1,557.00 |  |  |
| 12 CHINOOK OBSERVATION AQUARIUM | MILLER GLASS Corp | 1/2 CLR GLASS,STILES FRAM |  | 1999 | 8458109.7 |  | 74726 | 5,545.00 |  |  |
| 13 CCI-GS SERERES 3.OKG | TRICOSTALINDUSTRIES |  |  | 1999 | 8458109.7 |  | 72952 | 432.11 |  |  |
| ${ }^{14}{ }^{14}$ PENTIUMMIIPROCESSOR NOTEBOOK | GATEWAY COMPANIES |  |  | 1999 | 8458109.7 |  | ${ }^{75155}$ | 10,473.00 |  |  |
| 15 $11^{12}$ ' FLTED UTL TRLR | INDEPENDENT TRALLER \& EQUIP | 2-T8 UTILTIY TRALER |  | $\stackrel{1999}{1999}$ | ${ }^{84585109.7}$ |  |  |  |  |  |
| 16. PENİUMPROCESSOR | GATEWAYCOMPANIES |  |  |  | 8458109.7 |  | ${ }_{\text {TOTAL }}$ | $\xrightarrow{49,6800.59}$ |  |  |





Location: 1. Headquarters 2. Chander 3. Prosser 4. Roza 5. Nelson Springs 6. Hatchery 7. Klickitat 8. Cle Elum 9. WDFW 10. Missing/Stole
Condition: 4. Good 5. Fair 6. Poor 7. Salvageable 8. Missing/Stolen

| No. | ITEM DESCRIPTION | VENDOR | MAKE/ MODEL | SERIAL NUMBER | YEAR | FUND NUMBER | PROPERTY NUMBER | Doc. NUMBER | ITEM COST | location CONDITION CODE | COMMENTS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | OFFICE XP PRO | IINCOMMAND | OEM |  | 2002 | 8458101.100 |  | 128636 | 779.90 | 5/4 |  |
| 2 | HI BACK CHAIR | INSIDE OREGON ENTERP |  |  | 2002 | 8458101.100 |  | 127827 | 350.00 | 5/4 |  |
| 3 | $32 \times 71$ COPY BOARD | TRIBUNE OFFICE SUPPL | $32 \times 71$ COPY BOARD |  | 2002 | 8458101.100 |  | 122273 | 2,395.00 | 5/4 |  |
| 4 | SNORKEL, MASKS, BOOTS | QUALITY SCUBA |  |  | 2002 | 8458101.100 |  | 116845 | 675.00 | 5/4 |  |
| 5 | $10 \times 42$ BINOCULARS | CABELA'S PROMOTION |  |  | 2002 | 8458101.100 |  | 113385 | 994.99 | 1/4 |  |
| 6 | RAFTING SUPPLIES:SEAT/PUMPETC | NORTHWEST RIVER SUPP | \#1220,1222,1709,1703, |  | 2002 | 8458101.400 |  | 130412 | 816.00 |  |  |
| 7 | $11^{\prime}$ OTTER SELF BAILING RAFT | NORTHWEST RIVER SUPP | \#1126 |  | 2002 | 8458101.400 |  | ${ }^{130412}$ | 1,436.00 | 5/4 |  |
| 8 | OAR WRAPPING PROTECTOR | LAVRO |  |  | 2002 | 8458101.400 |  | 130413 | 510.00 | 5/4 |  |
| 9 | (2) $10 \times 42$ BINOCULARS | CABELA'S PROMOTION |  |  | 2002 | 8458101.400 |  | 130411 | 1,989.98 | 5/4 |  |
| 10 | NEOSTRECTH WADERS 3MM | CABELA'S PROMOTION |  |  | 2002 | 8458101.400 |  | 130411 | 82.75 | 5/4 |  |
| 11 | TAGGING UNIT | NORTHWEST MARINE TEC | AUTO TAGGING UNIT |  | 2002 | 8458101.105 |  | 117544 | 129,600.00 | 5/4 |  |
| 12 | KEPCO POWER TRANSDUCER | HALCO INC |  |  | 2002 | 8458102.105 |  | 121034 | 908.64 |  |  |
| 13 | 19" MONITORW/INTEL PROC | GATEWAY COMPANIES | MONITOR/PROCESSOR | 1900980004575 | 2002 | 8458104.114 |  | 130658 | 1,598.00 | 1/4 |  |
| 14 | JVC 14" MONITOR | THE GOOD GUYS | AV14F703 |  | 2002 | 8458104.114 |  | 121247 | 1,045.96 | 1/4 |  |
| 15 | DIGITAL CAMERA | CDW, GOVT. INC. | MVC-CD400 | S010419649J | 2002 | 8458104.114 |  | 730415 | 786.00 | 1/4 |  |
| 16 | 450L INTEL NOTEBOOK | GATEWAY COMPANIES |  |  | 2002 | 8458106.108 |  | 123965 | 1,792.00 | $1 / 4$ |  |
| 17 | 4- DUAL FREQ READER | BIOMARK | $125 \mathrm{KH2} 134.2$ |  | 2002 | 8458106.108 |  | 123858 | 10,810.00 | 1/4 |  |
| 18 | DIGITAL CAMERA | CDW, GOVT. INC. | MVC-CD400 | S010419691G | 2002 | 8458106.108 |  | 130415 | 786.00 | $1 / 4$ |  |
| 19 | PROTECH TOOL BOXES | PEPIN 4 WHEELIN |  |  | 2002 | 8458106.108 |  | 129740 | 850.00 | 1/4 |  |
| 20 | FISHING SUPPLIES | NORTHWEST RIVER SUPP |  |  | 2002 | 8458106.108 |  | 123857 | 2,453.79 | 1/4 |  |
| 21 | NEOPN WADERS, PARKAS,RAINGR | CABELA'S PROMOTION |  |  | 2002 | 8458106.108 |  | 123095 | 869.55 |  |  |
| 22 | (2) SRX 400 RECVR/DATA LOGGER | LOTEX WIRELESS | W-5 |  | 2002 | 8458107.109 |  | 130414 | 9,990.00 |  |  |
| 23 | (2)4MHzRECVR | LOTEX WIRELESS | W31C. 148-152 |  | 2002 | 8458107.109 |  | 130414 | 19.100.00 |  |  |
|  | (4) ELEMENT ANTENNA'S | LOTEX WIRELESS |  |  | 2002 | 8458107.109 |  | 130414 | 880.94 |  |  |
|  | GPS UNIT | FORESTRY SUPPLIERS |  |  | 2002 | 8458107.109 |  | 113384 | 370.06 |  |  |
| 26 | 3 V MICRO CODED FISH TRNSMITRS | LOTEX WIRELESS |  |  | 2002 | 8458107.109 |  | 117307 | 23,068.89 |  |  |
|  |  |  |  |  |  |  |  | TOTAL | 214,939.45 |  |  |


[^0]:    1 Data is summarized from Neeley (2003).

[^1]:    ${ }^{1}$ Expansions involved dividing the number of daily detections by the McNary detection efficiency for that day of McNary passage. Methods of estimating detection efficiencies and the detection efficiency estimates are given in 2002 Annual Report: OCT -SNT Survival by Doug Neeley.

[^2]:    ${ }^{2}$ The figures essentially start with releases made after the beginning of the calendar year. However, in brood years 1998 and 2000 (respectively outmigration years 2000 and 2002), releases were made earlier than January 1. Survival indices for these earlier releases are not presented in the figures, but the estimates are given in tables given in Appendix A. However, survival-index estimates for any releases made before mid-March may be underestimates since McNary detectors were not on line until late March or early April. Some fish released before mid-March may have passed McNary and may not have been included in the McNary detections.

[^3]:    ${ }^{3}$ Refer to 2002 Annual Report: OCT-SNT Survival by Doug Neeley, Consultant to the Yakama Nation.

[^4]:    *Variable is survial index, weight is number of released fish
    ** Respectively adjusted and unadjusted for common Julian Week effects

[^5]:    *Variable is survial index, weight is number of released fish
    ** Respectively adjusted and unadjusted for common Julian Week effects

[^6]:    *Variable is survial index, weight is number of released fish
    ** Respectively adjusted and unadjusted for common Julian Week effects

[^7]:    ${ }^{1}$ For the 1997 brood, the number of fish tagged-per-raceway served as the base for survival estimates. For 1998 and subsequent brood years, fish volitionally leaving the raceways were read by PIT-tag detectors, and this number of detected fish per raceway served as the base for survival estimates.

[^8]:    ${ }^{2}$ Methods of estimating the release-to-McNary survival index are presented in Appendix A.
    ${ }^{3}$ Appendix A contains a discussion on logistic analysis of variation.
    ${ }^{4}$ For Spring Chinook, brood years 1997 through 1999 respectively correspond to outmigration years 1999 through 2001. Data summaries from pre-2002 brood years are presented in 2001 Annual Report: OCTSNT Survival by Doug Neeley. There were somewhat different methods of estimating survival indices in previous years, and those methods are discussed in Appendix A.
    ${ }^{5}$ Ray Brunson (United States Fish and Wildlife Service, Olympia, Washington) provided disease data. Between 59 and 61 fish were sampled and measured for BKD severity per raceway. Refer to Table 1.e.

[^9]:    ${ }^{6}$ The two pairs of raceways for which the OCT ratio exceeded that of the SNT were both at Jack Creek, which had three pairs total. This probably resulted in the fact that the Site x Treatment interaction was significant at the $10 \%$ level $[P=0.079$ in Table 1.b.2)].
    ${ }^{7}$ There may have also been differences in survival effects of the different parental sets. Such differences would be reflected in the Block source of the analysis of variation. The significant levels for the release-toMcNary and the released-to-tagged-number ratio respectively were $\mathrm{P}=0.143$ [Table 1.b.1)] and $\mathrm{P}=0.003$ [Table 1.b.2)].

[^10]:    ${ }^{8}$ The weighted estimated logistic covariate coefficient as sociated with the error source of Table 1.c.1) was -1.1883 for logit of release-to-McNary survival index regressed on BKD severity mean as well as site, raceway-pair, treatment and site x treatment interaction indicators.
    ${ }^{9}$ The weighted estimated logistic covariate coefficient associated with the error source of Table 1.c.2) was -0.6419 for logit of released-to-tagged-number ratio regressed on BKD severity mean as well as site, raceway-pair, treatment and site x treatment interaction indicators.

[^11]:    ${ }^{10}$ Weighted Pearson correlation of BKD index and logit of survival index is -0.816 (weight is number released). Weighted Pearson correlation of BKD index and logit of released-to-tagged-number ration is -0.693 (weight is number tagged).

[^12]:    * ND--Not Detected, VL--Very Low, L--Low, M--Moderate, H--High, VH--Very High
    ** Mean = Sum over ranks of product of rank and relative frequency within acclimation pond

[^13]:    ${ }^{11}$ The acclimation pond screens were pulled on March 15

[^14]:    ${ }^{12}$ Weights are the number of fish PIT-tagged for each raceway.
    ${ }^{13}$ Based on number of fish PIT-tagged in raceways

[^15]:    ${ }^{14}$ Weights are the number of PIT-tagged detections leaving the raceway.

[^16]:    ${ }^{15}$ Based on number of PIT-tagged fish detected volitionally leaving the raceways.

[^17]:    ${ }^{16}$ There were no Jack Creek releases for the 1997 brood.

[^18]:    ${ }^{17}$ Recommended reading on logistic regression: McCullagh, P. and Nelder, J.A. (1989) Generalized Linear Models ( $2{ }^{\text {nd }}$ edition), Chapman and Hall, London.

[^19]:    ${ }^{18}$ The daily McNary detection efficiency is the proportion of PIT-tagged fish passing McNary that are actually detected at McNary. It is the total number of fish jointly detected at McNary on the McNary date and that are also detected at downstream dams (John Day and Bonneville) divided by the total detected at the downstream dams that are estimated to have passed McNary on that date.
    ${ }^{19}$ In recent years experiments were conducted at John Day and Bonneville that varied the proportion of flow spilled in the daytime relative to the proportion spilled at night. To offset the electric power lost at one dam during a given period, contravening action was often taken at the other dam (Personal Communication, Rock Peters, U.S. Army Corps of Engineers, Portland, Oregon.) Given this situation, it was deemed more appropriate to pool John Day and Bonneville Dam-based estimates of the McNary detection rate. This means that the fish detected at both John Day and Bonneville dams were used twice to estimate the McNary detection efficiency (an effective "sampling with replacement").

[^20]:    ${ }^{20}$ Recall that the number tagged served as the base for the 1997-brood release-to-McNary smolt survival index because the number of fish volitionally released was not available.

[^21]:    ${ }^{21}$ Release periods indicted in table by ending exit date which is the last volitional-release detection day for the period, the next volitional-release day being associated with the next period.

[^22]:    ${ }^{22}$ This was done for the 1999, 2000, 2001, and 2002 emigrants, except that Bonneville Powerhouse 1 was omitted from the 2001 detection efficiency estimation. There were few Powerhouse 1 detections of fish in 2001 because Powerhouse 1 was essentially offline due to the extremely low flows in 2001.

[^23]:    ${ }^{23}$ What is meant by offset is: If McNary Julian date d's relative distribution was being estimated from the twelve downstream adjacent dates, then, for the $1^{\text {st }}$ preceding downstream adjacent day, the joint count from McNary day d-1 was used; for the $2^{\text {nd }}$ preceding downstream adjacent day, the joint count from McNary day d-2 was used; ... for the $6^{\text {th }}$ preceding downstream adjacent day, the joint count from McNary day d-6 was used; for the $1^{\text {st }}$ following downstream adjacent day, the joint count from McNary day d+1 was used; for the $2^{\text {nd }}$ following downstream adjacent day, the joint count from McNary day $\mathrm{d}+2$ was used; ... for the $6^{\text {th }}$ following downstream adjacent day, the joint count from McNary day d+6 was used.

[^24]:    * A subsequent shifting of stratum boundaries actually resulted in a separation between Julian dates 151 and 152 giving lower error mean deviance; however, this left less than 20 joint detections in last stratum, so the separation between 150 and 151 maintained
    ** Omitted and terminated stepwise process because $10 \%$ significance level not attained ( $P=0.121$ )

[^25]:    ${ }^{24}$ Spill proportion is McNary spill flow divided by total McNary discharge. Turbine proportion is McNary turbine flow divided by total McNary discharge. McNary flow database used was provided by Fish Passage Center, Portland, Oregon. Database originally created by U.S. Army Corps of Engineers.

[^26]:    ${ }^{1}$ For Coho, release-years 1999, 2000, 2001, and 2002 release years respectively corresponds to brood-years 1998, 1999, and 2000.

[^27]:    ${ }^{2}$ For Fall Chinook, release-years 2000, 2001, and 2002 release years respectively corresponds to broodyears 1999, 2000, and 2001.
    ${ }^{3}$ The stock was comprised of 60,560 Yakima and 23,308 Willard Coho prior to PIT-tagging

[^28]:    * logit transform = natural log [(survival index)/[1 -(survival index)]; refer to Appendix A.

[^29]:    ${ }^{4}$ The only Yakima release showing no evidence of leakage was the erroneous extra early release from Easton.
    ${ }^{5}$ The stock is actually mixed Yakima and Willard broodstock.
    ${ }^{6}$ Estimated Julian dates when $25 \%, 50 \%$, and $75 \%$ of total McNary passage occurred.
    ${ }^{7}$ This ignores the Easton Yakima brood releases for which the "early" release was erroneously made on Mach 28 instead of May 6.

[^30]:    * Actually Extra Early
    ** Actually mixed Willard and Yakima broodstock

[^31]:    ${ }^{8}$ Spill proportion is McNary spill flow divided by total McNary discharge. Turbine proportion is McNary turbine flow divided by total McNary discharge. McNary flow database used was provided by Fish Passage Center, Portland, Oregon. Database originally created by U.S. Army Corps of Engineers.

[^32]:    ${ }^{9}$ Spill proportion is McNary spill flow divided by total McNary discharge. Turbine proportion is McNary turbine flow divided by total McNary discharge. McNary flow database used was provided by Fish Passage Center, Portland, Oregon. Database originally created by U.S. Army Corps of Engineers.

[^33]:    ${ }^{13}$ Personal Communication, Rock Peters, U.S. Corps of Engineers, Portland, Oregon
    ${ }^{14}$ In 2001 only the counts from Powerhouse 2 at Bonneville were used because there were almost no detections at Powerhouse 1 due to low flows and the resulting limited use of Powerhouse 1.

[^34]:    ${ }^{15}$ Individual data estimates are given in Appendix Table 2.c.

[^35]:    1 with the exception that in one replication done in 2000, a wild fish apparently jumped from it's holding tank into the OCT holding tank, resulting in 6 wild and 4 OCT fish were run in the OCT, Wild, SNT replicate.

[^36]:    ${ }^{2}$ The acclimation period is the 24 hour time period starting when the smolts are selected and measured for a replicate, and ending just before they are introduced into the tank with the pikeminnows.
    ${ }^{3}$ The mean, median, and mode lengths of all smolts used in 2001 was $130.4,128$ and 126 mm respectively.

[^37]:    ${ }^{1}$ Formally the fisheries biologist working on the indirect predation study for the Yakima Nation, currently with Mobrand Biometrics, Vashon Island, Washington.
    ${ }^{2}$ Logistic regression of smolt survival on total passage was the specific regression tools used.

[^38]:    ${ }^{3}$ For example, there is evidence that, as flow increases, passage tends to increase. It was not possible to meaningfully isolate the effect of number of fish on survival from the effects of flow and other measured variables
    ${ }^{4}$ The weights used were number of fish released. The reason for the weighting is explained in Appendix A.

[^39]:    ${ }^{5}$ The hatchery is located at Cle Elum. The acclimation ponds are Clark Flats and Easton on the upper Yakima River and Jack Creek on the Teanaway.
    ${ }^{6}$ Time-period flow estimates are weighted means of daily flows within the period for the site, the weights being the daily number of fish detected exiting the site.
    ${ }^{7}$ Fisheries Biologist, Yakama Nation, Toppenish Washington.

[^40]:    ${ }^{8}$ BKD severity measures provided by Data provided by Ray Brunson (United States Fish and Wildlife Service, Olympia, Washington.]

[^41]:    *Weight is daily number released applied to daily flow

[^42]:    ${ }^{9}$ The daily McNary detection efficiency is the proportion of PIT-tagged fish passing McNary that are actually detected at McNary. It is the total number of fish jointly detected at McNary on the McNary date and that are also detected at downstream dams (John Day and Bonneville) divided by the total detected at the downstream dams that are estimated to have passed McNary on that date.

[^43]:    ${ }^{10}$ Expanded McNary Detections $=($ McNary Detections $) /($ Detection Efficiency $)$

